

HEAVY METALS IN SOILS AND PLANTS IN MANGO ORCHARDS IN PETROLINA, PERNAMBUCO, BRAZIL⁽¹⁾

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SUMMARY

The monitoring of heavy metal concentrations in areas under intensive agriculture is essential for the agricultural sustainability and food safety. This paper evaluates the total contents of heavy metals in soils and mango trees in orchards of different ages (6, 7, 8, 9, 10, 11, 14, 16, 17, 19, and 26 years) in Petrolina, Pernambuco, Brazil. Soil samples were taken from the layers 0-20 cm and 20-40 cm, and mango leaves were collected in the growth stage. Areas of native vegetation (Caatinga) adjacent to the cultivated areas were used for comparison. The total concentrations of heavy metals (Cu, Cr, Fe, Zn, Mn, Ni, and Pb) were determined in soils and leaves. In general, mango cultivation led to Cu and Zn accumulation in the soil surface and to a reduction in the contents of Ni, Pb, Mn, and Fe in surface and subsurface. Since contamination by Cu, Zn, and Cr was detected, these areas must be monitored to prevent negative environmental impacts. For instance, the presence of Cr in mango tree leaves indicates the need to investigate the source of the element in these orchards. The management strategies of the different companies led to deficiency or excess of some metals in the evaluated areas. However, the Fe and Mn levels were adequate for the mineral nutrition of mango in all areas.

Index terms: Trace elements, agricultural inputs, guiding values, contamination.

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RESUMO: METAIS PESADOS EM SOLOS E PLANTAS EM POMARES DE MANGA, EM PETROLINA, PERNAMBUCO

O monitoramento dos teores de metais pesados nas áreas com uso intensivo de insumos agrícolas é de grande importância para a sustentabilidade agrícola e para a segurança alimentar. O objetivo deste trabalho foi avaliar os teores totais de metais pesados em solo e planta em áreas de mangueira com diferentes tempos de cultivo (6, 7, 8, 9, 10, 11, 14, 16, 17, 19 e 26 anos), em Petrolina (PE). Foram coletadas amostras de solo nas profundidades de 0-20 e 20-40 cm e de folhas de mangueira na fase vegetativa. Áreas de vegetação nativa (caatinga), adjacentes às áreas cultivadas, foram utilizadas para comparação. As concentrações totais de metais pesados (Cu, Cr, Fe, Zn, Mn, Ni e Pb) foram determinadas em solos e folhas. O cultivo de mangueira proporcionou acúmulo superficial dos metais Cu e Zn e redução dos teores de Ni, Pb, Mn e Fe, tanto em superfície como em subsuperfície. As áreas cultivadas apresentaram contaminação por Cu, Zn e Cr, devendo ser feitos monitoramentos nessas áreas para evitar impactos ambientais. Os diferentes tipos de manejo das empresas no cultivo de manga proporcionaram deficiência e excesso de alguns micronutrientes nas áreas avaliadas. Ferro e Mn, no entanto, supriram a nutrição mineral da mangueira em todas as áreas. A presença de Cr nas folhas de mangueira sugere a necessidade de investigação da ocorrência desse metal nessas áreas cultivadas.

Termos de indexação: Elementos-traço, insumos agrícolas, valores orientadores, contaminação.

INTRODUCTION

The Brazilian fruit production is nationally and internationally relevant, accounting for around 25 % of the total national agricultural production and of 10 % of the international fruit production, respectively (Lacerda & Lacerda, 2004; Medeiros et al., 2005). One of the fruits with a high commercial value is mango (*Mangifera indica* L.) (Albuquerque et al., 1999). In northeastern Brazil, mango is a key element of the regional economy, produced for exportation (Ribeiro et al., 2009), mainly for the European and North American markets (Lucafé & Boteon, 2001), and recently focused on Asia as well. Approximately 92 % of the country's mango exports are produced in this region (Silva, 2008).

The high mango production of this region, of around 25 t ha⁻¹ (Silva, 2008), is the result of a mechanized agriculture with intensive application of agricultural inputs. However, the increase in the use of fertilizers and pesticides causes a series of environmental impacts, e.g., contamination by heavy metals, salts, and agrotoxics that leach and flow from the cultivated areas into watersheds and groundwater sources (Berti, 2003) and may also accumulate in the soil.

Fertilizers, for example, despite providing crops with nutrients, contain heavy metals that are potentially toxic to human health and to the environment (Peris et al., 2008). These elements also frequently form part of the active components of pesticides. Thus, the continued use of fertilizers and pesticides was identified as the primary pathway for the introduction of metals into agricultural soils (Núñez et al., 2006). Organic fertilizers, depending

on the origin, can also release heavy metals in the soil, originating mainly from additives in animal feed (O'Neill, 1993; Mendes et al., 2010).

In the Brazilian literature, data that reflect the changes in the soil chemical properties as related to the time of mango cultivation, mainly focused on the behavior of heavy metals in these soils, are scarce. Therefore, the behavior of these elements in the soil-plant system should be evaluated, in order to find solutions or control methods to monitor the accumulation of these metals in soil and plants, reducing the risk of environmental pollution (Oliveira et al., 2008).

In this context, due to the importance of environmental monitoring of crops with a high contribution of agrochemical inputs, as in the case of mango farming in Northeastern Brazil, this study had the objective of evaluating heavy metal contents (Cu, Cr, Fe, Mn, Ni, Pb, and Zn) of soil and plants in mango orchards after different cultivation periods.

MATERIAL AND METHODS

The soils and plant samples analyzed were collected from mango orchards in Petrolina, Pernambuco, Brazil (9° 23' 38.97" S, 40° 30' 34.58" W).

The 11 study areas with different histories of fertilization and liming belong to eight agricultural companies producing mainly for mango exportation. In general, fertilizing by agricultural companies provides 40 L/plant of organic matter (OM) annually from animal sources, such as goat or cattle manure or other, commercial products used as organic

fertilizers. The companies use different N doses and sources, the most common of which is ammonium sulfate, at rates of 300 - 500 g/plant. The most commonly used K source was potassium sulfate, at rates of 100 - 800 g/plant. For phosphorus, one company reportedly applies 1,000 g/plant of superphosphate. Some companies utilize other types of fertilizers, such as magnesium sulfate, calcium nitrate, zinc sulfate, iron sulfate, and boric acid, among others. In the different phenological stages of the crop, the highest fertilizer rates are applied during the production period. The acidity in the soil surface and subsurface is reduced, when necessary, by lime and gypsum application. All companies water their crops daily by micro-sprinkler irrigation. To control pests and diseases, all companies adopted the integrated fruit production (IFP) program, with applications of pesticide spraying.

Soil samples were collected from orchards with mango trees of 11 ages (6, 7, 8, 9, 10, 11, 14, 16, 17, 19, and 26 years since planting) and from their respective reference areas (Caatinga biome with no or minimal human interference). Soil from two layers (0-20 and 20-40 cm) was sampled on the edges of the area under the tree canopy. For this purpose, the areas were divided into three equal plots, covering an area of 1,040 or 1,800 m², with a tree spacing of 8 x 5 m or 10 x 10 m, respectively. Of each plot, 15 points were randomly sampled to form the composite sample.

To evaluate the contents of heavy metals in the mango plants, four leaves/plant were collected, in the four cardinal directions, at the average canopy height. To collect the samples, the cultivated area was divided into three equal plots. On each plot, 15 trees were randomly selected in the crop rows, at the points of soil sampling, with the exception of the 8-year-old orchard, from which no leaves were collected because, unlike the other areas, the crop was in the production stage Normal mature and recently matured branches were sampled, taking the leaves from the middle section of the penultimate or last offshoot of the branch (Embrapa, 2000).

The soil samples were air-dried, ground and sieved through 2 mm mesh. To determine the total contents of the metals Cu, Cr, Fe, Zn, Mn, Cd, Ni, and Pb, sub-samples of these soils were ground in an agate mortar and sieved through 0.3 mm (ABNT no. 50); a stainless steel screen was used to avoid contamination. The soil and plant samples were digested by method 3051A (USEPA, 1998). The metals were determined in an atomic absorption spectrophotometer (AAAnalyst 800 Perkin Elmer), using the flame technique.

The analysis quality was controlled by comparison with samples with known metal contents, certified by the NIST (National Institute of Standards and Technology). SRM 2709 San Joaquin soil (Baseline trace element concentrations) and SRM 1570a (Trace

elements in spinach leaves) were digested and every 30 samples were analyzed.

The statistical analysis was performed in a 2 x 2 factorial arrangement (two environments and two layers) with three replicates, for a total of 132 experimental units. For the chemical analysis of plants, 10 cultivation periods were considered with three replicates, for a total of 30 experimental units. The statistical analysis was based on the application of the F test to analysis of variance, correlation analysis, and Tukey's Test ($p < 0.05$), using Statistical Analysis System software (SAS, 1999).

RESULTS AND DISCUSSION

Recoveries (in %) of the elements expressed in comparison to NIST values were as follows: Zn (97 %), Pb (101 %), Fe (119 %), Ni (83 %), Mn (103 %), and Cr (90 %). From the leaves, the following percentages were recovered: Cu (80 %), Mn (108 %), Zn (85 %), Pb (145 %), Ni (135 %), and Cd (77 %). The Cu contents in the 0-20 cm layer varied from 3.14 to 40.10 mg kg⁻¹ in the cultivated areas (CA) and from undetectable (UD) to 16.07 mg kg⁻¹ in the reference areas (RA). Significant Cu accumulations ($p < 0.05$) were found in this layer, with greater contents in the areas after 7, 10, 11, 17, and 26 years of cultivation (Figure 1). In the 9-year-old orchard, a statistically significant difference was also observed but with a higher content in the RA, most likely due to a low number of fertilizer applications on that property. The Cu levels were highest in the areas under cultivation for 10 and 26 years, with a greater difference from their respective reference areas. This finding indicates that Cu accumulation is most likely due to the use of fungicides and fertilizers. In general, a trend was observed of greater Cu contents with increasing orchard age. Analyzing soils under grapes in the same region, Costa (2009) found greater Cu contents in the vineyards than in Caatinga areas and attributed this increase to the use of agro-chemicals.

For the 20-40 cm layer, the Cu content varied from 2.26 to 20.09 mg kg⁻¹ in the orchards and from UD to 9.92 mg kg⁻¹ in the Caatinga. Significant differences were found in the RA after 7, 10, and 26 years. Greater differences were observed in the RA after 6, 11, and 19 years. At this depth, the Cu content was less affected by agricultural practices.

In relation to the effect of the depth, the results show decreasing Cu levels with increasing depth in both environments, with greater differences in the areas after 6, 7, 11, 16, 17, and 26 years of cultivation. These decreases were due to the strong interactions of Cu with soil OM due to the greater selectivity of the colloid surfaces for the metal (Nascimento & Fontes, 2004). Other authors also found this decrease with increasing depth, which was attributed to the

low Cu mobility, the bonds with the OM through the formation of stable organometallic compounds, and the low solubility of Cu (Williams et al., 1980; Oliveira & Matiazzo, 2001; Komárek et al., 2008; Costa, 2009).

The maximum values of Cu found in this study are below the prevention value (60 mg kg^{-1}) indicated by Brazilian legislation (CONAMA, 2009), which allows the conclusion that the content of this metal in the soils studied would not affect the environmental quality. However, when compared to the values obtained by Biondi et al. (2010) for the state of Pernambuco, specifically in a low rainfall region, the Cu levels were higher in the 6, 11, 14, and 16 year-old orchards than the natural content in Ultisols, both at the surface (3.18 mg kg^{-1}) and in the subsurface (2.90 mg kg^{-1}), indicating Cu contamination of the soils. The same was observed in the areas after 7, 8, 9, 17, and 19 years of cultivation when compared to the Cu content of the

Entisol Quartzipsamments in the surface (0.50 mg kg^{-1}) and the subsurface (1.28 mg kg^{-1}). In the 10 and 26 year-old orchards, the Cu contents were also higher than the Cu values found in Oxisols, (surface 2.03 mg kg^{-1}) and in the subsurface (3.40 mg kg^{-1}). Mendes et al. (2010) found a significant increase in the Cu level in soil under melon, with values as high as 102.5 and $125.33 \text{ mg kg}^{-1}$, above the prevention value of 60 mg kg^{-1} (CONAMA, 2009).

In the orchards, significant and positive correlations ($p < 0.05$) were observed between Cu and OM and between Cu and Zn in the 0-20 cm layer (Table 1). For Cu adsorption, organic matter represents one of the main attributes, due to the high degree of selectivity of OM for the element, with the formation of inner-sphere complexes that result from specific adsorption (Guilherme & Anderson, 1998). It is worth highlighting that due to the element's ground state electronic configuration $[\text{Ar}]3d^{10}4s^1$, Cu is highly reactive with the carboxyl and phenolic groups of OM, resulting in a high binding energy (Croué et al., 2003; Casali et al., 2008). The lack of correlation between Cu and P at the surface may be related to the vertical dislocation of P in organic and inorganic forms caused by the low P adsorption capacity of these sandy soils (Rheinheimer et al., 2003; Galvão et al., 2008).

In the reference areas, the correlation of Cu with Ni, Pb, Mn, Zn, Cr, and Fe was highly significant and positive ($p < 0.01$), suggesting a common origin of these elements from the source material.

The Ni content in the 0-20 cm layer varied from UD to 0.65 mg kg^{-1} in the CA and from UD to 24.67 mg kg^{-1} in the RA (Figure 2). Significant differences were found between the CA and RA in this layer after 6, 11, 14, 16, 17, and 19 years, with greater contents in the RA and much lower in the CA. These findings indicate that the agricultural management did not affect Ni accumulation in the orchards. These higher values of Ni in the RA, mainly than in orchards under cultivation for 14 and 16 years with Ni contents of 24.67 mg kg^{-1} , indicate a strong influence of the source material, while the reduction in the contents of this metal in the orchards in relation to the reference areas may be caused by removal of the crop or by leaching into deeper soil layers.

The higher Ni levels found in the CA are much lower than the prevention value (30 mg kg^{-1}) of CONAMA (2009). In comparison to the quality reference values of Pernambuco (Biondi et al., 2010), the orchards after 11 and 14 years of cultivation have higher Ni contents than the areas in the Sertão region, belonging to the Ultisol class, in the subsurface (1.10 mg kg^{-1}). The high correlation between Ni and OM, especially in the reference areas, is most likely due to the decrease in soil OM with cultivation (data not shown). This fact was also observed for the other metals studied (Table 1).

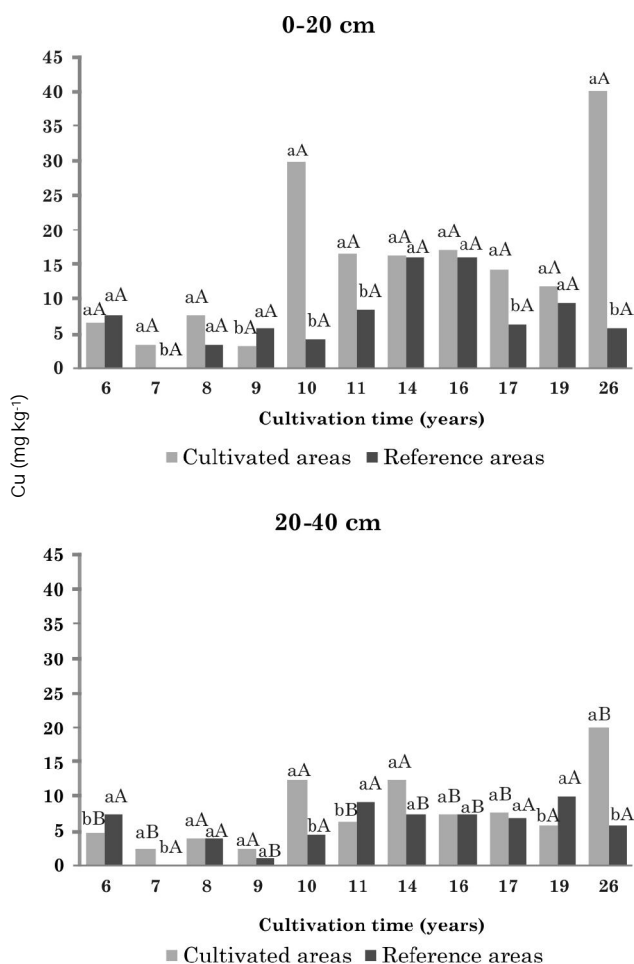


Figure 1. Total Cu contents in cultivated (CA) and reference areas (RA) in two layers (0-20 and 20-40 cm) after different cultivation periods. Means followed by lowercase letters between environments (CA and RA) and capital letters between layers (0-20 and 20-40) did not differ statistically by Tukey's test ($p < 0.05$).

The contents of Pb in the 0-20 cm layer varied from UD to 1.32 mg kg⁻¹ in the CA and from UD to 9.29 mg kg⁻¹ in the RA. Significant differences were observed in the surface layer of the orchards after 6, 11, 14, 16, and 19 years of cultivation (Figure 3), with higher Pb contents found in the RA. These results show that Pb accumulation in the orchard areas was not affected by agricultural practices. In the 20-40 cm layer, the contents of Pb varied from UD

to 3.47 mg kg⁻¹ in the CA and from UD to 7.88 mg kg⁻¹ in the RA. In contrast, Mendes et al. (2010) found higher Pb concentrations in melon plantations than in the soils with native vegetation. The low values found in the CA may be a result of harvest removal.

The contents of Mn in the 0-20 cm layer varied from 45.47 to 133.20 mg kg⁻¹ in the CA and from 15.29 to 348.40 mg kg⁻¹ in the RA. Almost all areas differed significantly, some having greater contents in the CA

Table 1. Linear correlations between heavy metal contents in soil for cultivated (CA) and reference areas (RA) in two layers

| Areas | | Cu | Ni | Pb | Mn | Zn | Cr | Fe |
|----------|----|--------------------|---------------------|---------------------|---------------------|---------------------|---------------------|---------------------|
| 0-20 cm | | | | | | | | |
| MO | CA | 0.36* | 0.37* | -0.24 ^{ns} | 0.31 ^{ns} | 0.33 ^{ns} | -0.20 ^{ns} | 0.17 ^{ns} |
| | RA | 0.88** | 0.92** | 0.78** | 0.55** | 0.35* | 0.81** | 0.92** |
| P | CA | 0.22 ^{ns} | 0.11 ^{ns} | -0.11 ^{ns} | 0.53** | 0.18 ^{ns} | 0.08 ^{ns} | -0.06 ^{ns} |
| | RA | 0.05 ^{ns} | -0.13 ^{ns} | 0.17 ^{ns} | 0.29 ^{ns} | 0.39* | -0.16 ^{ns} | -0.04 ^{ns} |
| Cu | CA | | 0.04 ^{ns} | 0.02 ^{ns} | 0.01 ^{ns} | 0.36* | -0.17 ^{ns} | 0.33 ^{ns} |
| | RA | | 0.92** | 0.89** | 0.78** | 0.44** | 0.76** | 0.95** |
| Ni | CA | | | 0.12 ^{ns} | -0.01 ^{ns} | -0.18 ^{ns} | -0.09 ^{ns} | 0.43* |
| | RA | | | 0.87** | 0.76** | 0.26 ^{ns} | 0.73** | 0.96** |
| Pb | CA | | | | 0.01 ^{ns} | -0.04 ^{ns} | -0.01 ^{ns} | 0.30 ^{ns} |
| | RA | | | | 0.86* | 0.26 ^{ns} | 0.62** | 0.87* |
| Mn | CA | | | | | 0.49** | -0.22 ^{ns} | -0.09 ^{ns} |
| | RA | | | | | 0.21 ^{ns} | 0.42* | 0.78** |
| Zn | CA | | | | | | -0.19 ^{ns} | 0.06 ^{ns} |
| | RA | | | | | | 0.21 ^{ns} | 0.34* |
| Cr | CA | | | | | | | 0.22 ^{ns} |
| | RA | | | | | | | 0.77** |
| 20-40 cm | | | | | | | | |
| MO | CA | 0.36* | 0.00 ^{ns} | -0.00 ^{ns} | 0.50** | 0.35* | 0.10 ^{ns} | -0.02 ^{ns} |
| | RA | 0.41* | 0.92** | 0.69** | 0.56** | 0.50** | 0.96** | 0.94** |
| P | CA | 0.48** | 0.18 ^{ns} | 0.19 ^{ns} | 0.60** | 0.36* | 0.29 ^{ns} | 0.13 ^{ns} |
| | RA | 0.31 ^{ns} | -0.03 ^{ns} | 0.33 ^{ns} | 0.49** | 0.22 ^{ns} | -0.11 ^{ns} | 0.03 ^{ns} |
| Cu | CA | | 0.19 ^{ns} | 0.24 ^{ns} | 0.11 ^{ns} | 0.12 ^{ns} | -0.09 ^{ns} | 0.24 ^{ns} |
| | RA | | 0.54** | 0.74** | 0.76** | 0.73** | 0.44* | 0.60** |
| Ni | CA | | | 0.37* | -0.03 ^{ns} | -0.19 ^{ns} | -0.01 ^{ns} | 0.73** |
| | RA | | | 0.82* | 0.72** | 0.45** | 0.88** | 0.97** |
| Pb | CA | | | | 0.03 ^{ns} | -0.14 ^{ns} | -0.04 ^{ns} | 0.18 ^{ns} |
| | RA | | | | 0.89** | 0.49** | 0.68** | 0.78** |
| Mn | CA | | | | | 0.50** | 0.11 ^{ns} | -0.00 ^{ns} |
| | RA | | | | | 0.48** | 0.52** | 0.73** |
| Zn | CA | | | | | | -0.07 ^{ns} | -0.13 ^{ns} |
| | RA | | | | | | 0.50** | 0.54** |
| Cr | CA | | | | | | | 0.25 ^{ns} |
| | RA | | | | | | | 0.90** |

*, ** and ns: significant at 5 and 1 % by Tukey's test and not significant, respectively

and others in the RA (Figure 4). The soils contained higher Mn levels in the RA after 6, 11, 14, 16, 17, and 19 years, differing from the areas after 8, 9, 10, and 26 years, where the levels were higher in CA than in RA. In the 20-40 cm layer, the Mn levels varied from 27.53 to 68.27 mg kg⁻¹, and the contents in the RA exceeded those in the CA, in all but the area under cultivation for 8 years, where the CA value was higher than that of the RA. These differences show that in the areas where the Mn levels were lower in the RA than in the CA, a Mn-poor source material had been applied. The other RAs showed that the main Mn sources in soils are the fertilizers, corroborating the data of Micó et al. (2006) and Peris et al. (2008). These low Mn levels in the CA were expected, given that Mn is a micronutrient found in large quantities in the mango peel, and after Fe, Mn is the most abundant

micronutrient in the fruit pulp (Pinto, 2002). According to Magalhães & Borges (2000), approximately 2.71 g of Mn are exported per ton of fruit. These results show the impoverishment of the studied areas, with a clear need to replace this micronutrient in the soil.

In general, the dynamics of Mn was affected by the content of organic matter, with a decrease in Mn levels observed with increasing soil depth, with the exception of the areas cultivated for 11 and 26 years, where Mn was accumulated in the deeper layers.

The contents of Fe in the 0-20 cm layer varied from 1844 to 6491 mg kg⁻¹ in the CA and from 642 to 34,633 mg kg⁻¹ in the RA. Significant differences were observed for the majority of the areas (Figure 5), with a higher Fe content in the RA, similar to the behavior

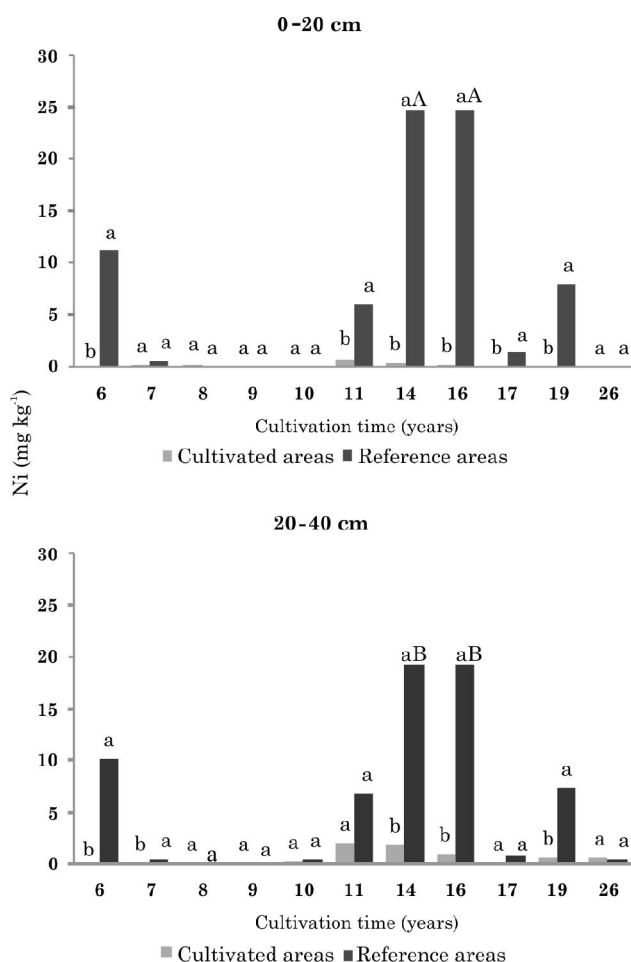


Figure 2. Total Ni contents in cultivated (CA) and reference areas (RA) in two layers (0-20 and 20-40 cm) after different cultivation periods. Means followed by lowercase letters between environments (CA and RA) and capital letters between layers (0-20 and 20-40) did not differ statistically by Tukey's test ($p < 0.05$).

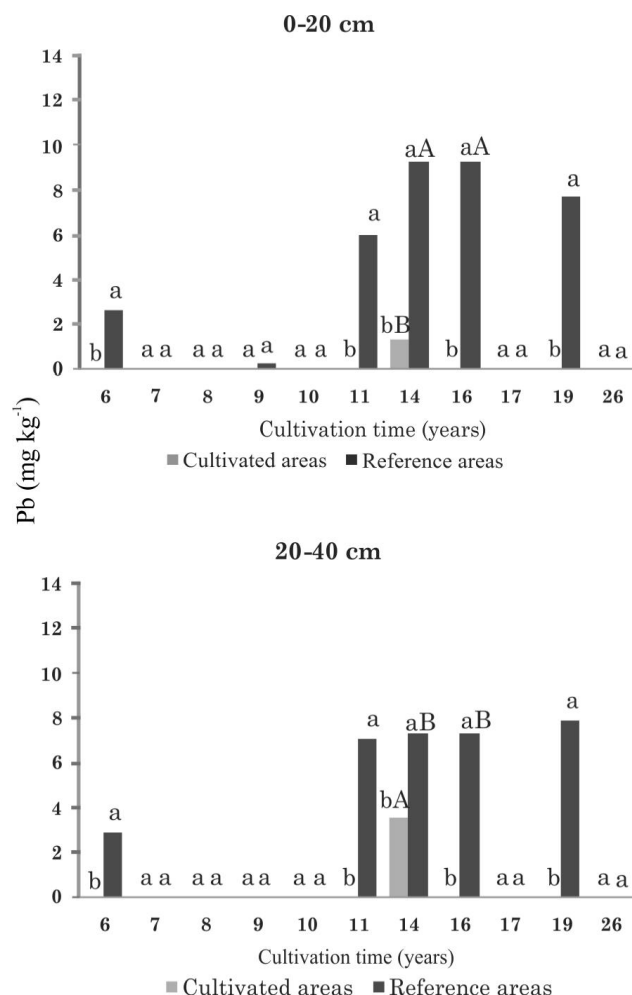


Figure 3. Total Pb contents in cultivated (CA) and reference areas (RA) in two layers (0-20 and 20-40 cm) after different cultivation periods. Means followed by lowercase letters between environments (CA and RA) and capital letters between layers (0-20 and 20-40) did not differ statistically by Tukey's test ($p < 0.05$).

found for Mn. Only the surface layer of the 7-year-old orchard contained a higher Fe level than that of the RA. In the 20-40 cm layer, the Fe level varied from 1,479.3 to 9,361.0 mg kg⁻¹ in the CA and from 758 to 35,000 mg kg⁻¹ in the RA. The Fe levels in the 7 and 9-year-old orchards were higher than in the RA. The reduction of the Fe level in the CA is a result of removal at mango harvest, as Fe is a micronutrient found at high concentrations in mango fruit (Pinto, 2002; Medeiros et al., 2005).

In the RA, positive and significant correlations were observed between Fe and OM in both layers (Table 1). According to Oliveira & Nascimento (2006), the contents of available Fe are linked to the OM fraction, which explains the high correlations between Fe and OM found in this study. This correlation was also

reported by Costa (2009) and corroborates the important role of OM for Fe availability.

The Zn levels in the 0-20 cm layer varied from 32.07 to 102.87 mg kg⁻¹ in the CA and from 0.20 to 40.53 mg kg⁻¹ in the RA. In the areas after 7, 10, 14, 17, 19, and 26 years, significant differences were found, with greater Zn contents in the CA (Figure 6), indicating that these areas were affected by alterations caused by Zn accumulation, most likely due to the use of fertilizers and pesticides. In the 20-40 cm layer, the Zn levels varied from 22.40 to 74.20 mg kg⁻¹ in the CA and from 1.73 to 26.20 mg kg⁻¹ in the RA. In most CA, the Zn contents were significantly higher than in the RA. Some companies reported adding zinc sulfate to the fertilizer, as in the 10, 17, 19, and 26-year-old orchards, which explains the greater differences in Zn accumulation in these areas.

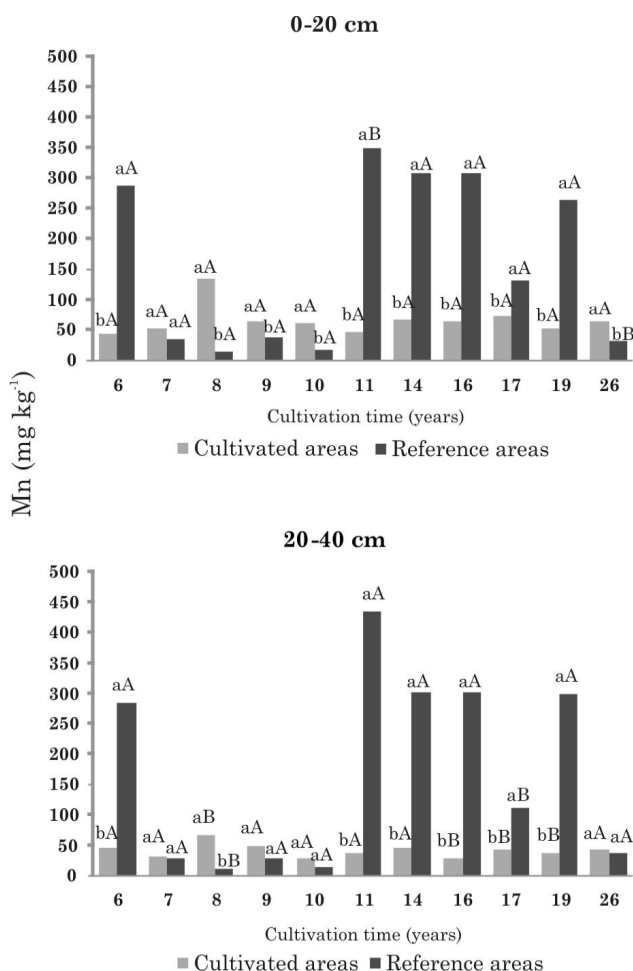


Figure 4. Total contents of Mn in cultivated (CA) and reference areas (RA) in two layers (0-20 and 20-40 cm) after different cultivation periods. Means followed by lowercase letters between environments (CA and RA) and capital letters between layers (0-20 and 20-40) did not differ statistically by Tukey's test ($p < 0.05$).

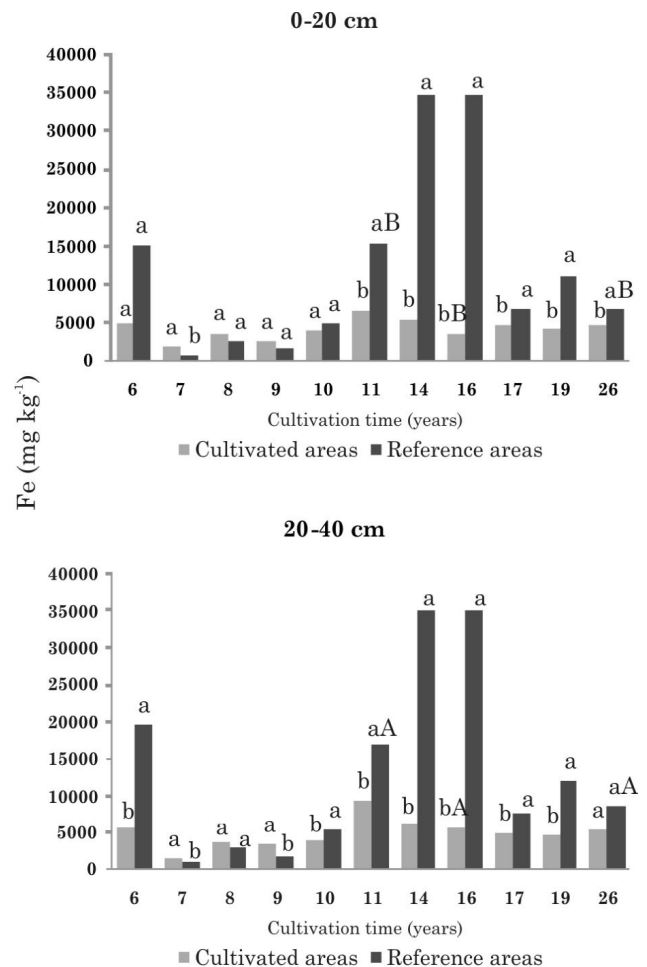


Figure 5. Total contents of Fe in cultivated (CA) and reference areas (RA) in two layers (0-20 and 20-40 cm) after different cultivation periods. Means followed by lowercase letters between environments (CA and RA) and capital letters between layers (0-20 and 20-40) did not differ statistically by Tukey's test ($p < 0.05$).

The Zn levels in the 20-40 cm layer presented significant differences between the 10, 17, 19, and 26-year-old orchards, with decreasing contents in the deeper soil layers. This dynamic was observed in the RA in relation to the 9-year-old orchard, and the inverse was observed for the 19-year-old orchard, with increasing Zn concentration with increasing depth. This reduction in the Zn content from the surface to the subsurface was influenced by changes in the soil pH caused by liming, as Zn is a mobile element in acidic soil. However, Zn is strongly retained by OM and oxides when the pH increases (Nascimento et al., 2002).

The Cr levels in the 0-20 cm layer varied from 6.31 to 77.99 mg kg⁻¹ in the CA and from 0.31 to 68.86 mg kg⁻¹ in the RA (Figure 7). In the areas after 6, 7, and 8 years, significantly higher Cr levels were

observed in the CA than the RA, indicating human influence with fertilizers or pesticides, with greatest effects on the 6-year-old orchard. Significant differences were also observed in the areas after 14, 16, 17, and 19 years, but with greater Cr levels in the RA than the CA, showing the influence of the source material on Cr accumulation. In the 20-40 cm layer, the Cr levels varied from 7.23 to 81.03 mg kg⁻¹ in the CA and from UD to 60.39 mg kg⁻¹ in the RA. At this depth, statistically significant differences were observed, with higher Cr levels in the CA for the same areas at the surface, including the 9-year-old orchard. The 6 year-old orchard was the only area with a Cr content above the prevention value (75 mg kg⁻¹) indicated by CONAMA (2009), in both layers. These Cr levels were most likely due to the use of mineral and organic fertilizers. According to Malavolta (2006),

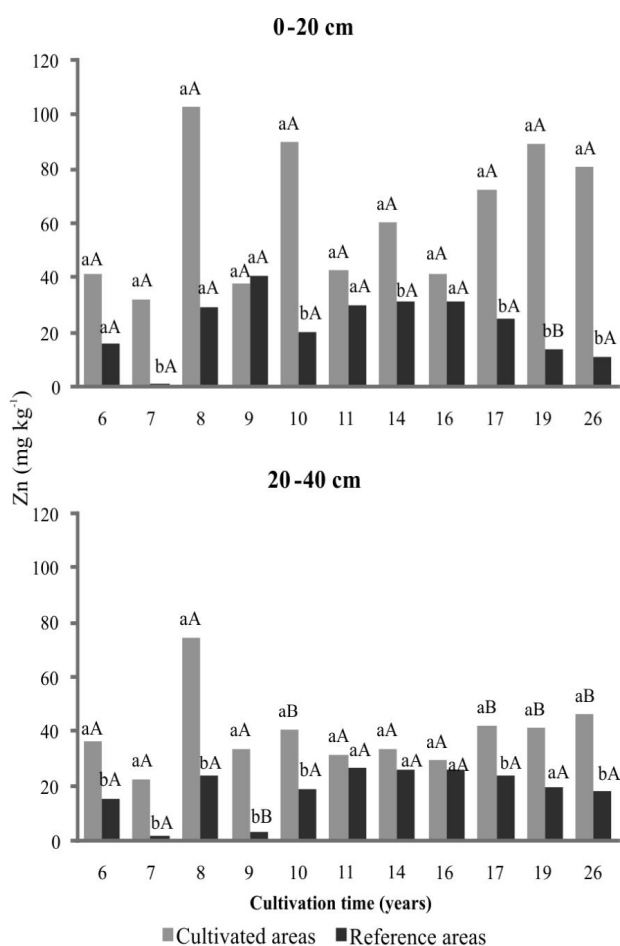


Figure 6. Total Zn contents in cultivated (CA) and reference areas (RA) in two layers (0-20 and 20-40 cm) after different cultivation periods. Means followed by lowercase letters between environments (CA and RA) and capital letters between layers (0-20 and 20-40) did not differ statistically by Tukey's test ($p < 0.05$).

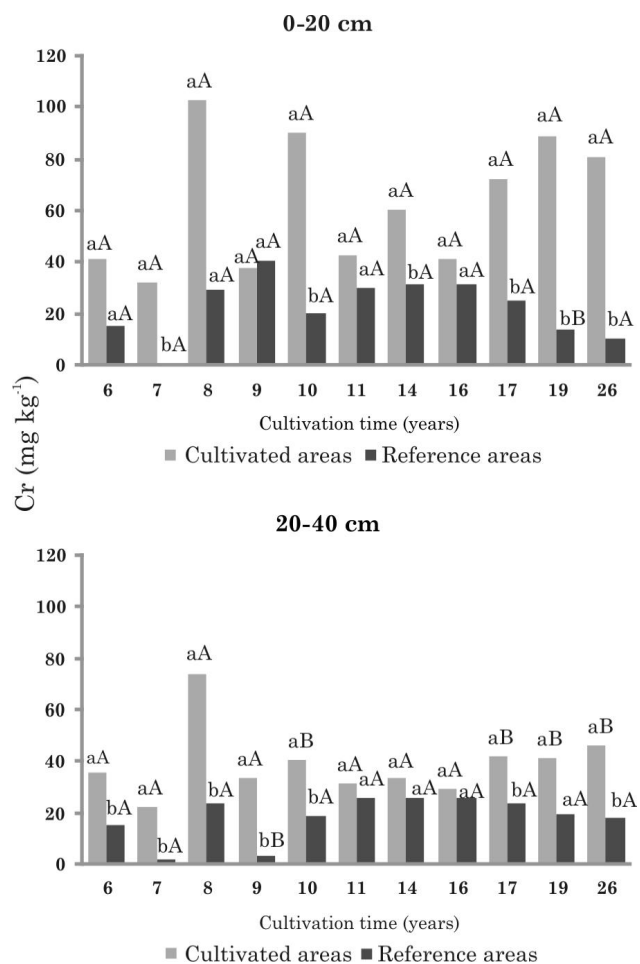


Figure 7. Total Cr levels in cultivated (CA) and reference areas (RA) in two layers (0-20 and 20-40 cm) after different cultivation periods. Means followed by lowercase letters between environments (CA and RA) and capital letters between layers (0-20 and 20-40) did not differ statistically by Tukey's test ($p < 0.05$).

the Cr level is higher in phosphate fertilizers. These results show that this area needs monitoring, in order to minimize the environmental impact. However, the Cr levels in all orchards were above the quality reference values of Pernambuco (Biondi, 2010) compared to the respective soil classes of the Sertão region, with the exception of the subsurface layer of the 10-year-old orchard. Thus, Cr contamination was confirmed in the 6, 11, 14, and 16-year-old orchards compared with the class of Ultisols (12.60 and 14.70 mg kg⁻¹ at the surface and subsurface, respectively), in the 7, 8, 9, 17, and 19-year-old orchards compared to the Entisol Quartzipsamments (2.38 and 3.95 mg kg⁻¹ in the surface and subsurface, respectively) and in the 10 and 26-year-old orchards compared with the class of Oxisols (10.05 and 11.93 mg kg⁻¹ in the surface and subsurface, respectively).

The Cu leaf content of the mango trees varied from 4.29 to 63.73 mg kg⁻¹ in the evaluated orchards (Table 2). According to the interpretation ranges of the nutritional status for micronutrients (Quaggio, 1996), the 14-year-old orchard was found to be Cu-deficient. In the areas after 16, 17, and 26 years of cultivation the Cu levels were between the deficient and adequate range, and the 9-year-old orchard had Cu contents above the adequate range. The other areas were within the ideal range. A large part of the Cu found in the leaves was due to the use of Cu-based chemical products for phytosanitary control.

The Ni leaf contents varied from UD to 1.24 mg kg⁻¹ (Table 2). In most areas, the Ni levels were

undetectable or close to zero. Only in the 17-year-old orchard the value was higher. Despite considered essential, the Ni demand of mango is very low.

The Mn leaf concentrations varied from 53.87 to 569.60 mg kg⁻¹ (Table 2). Only in the areas after 7 and 19 years of cultivation, the Mn contents were within the range considered adequate (Quaggio, 1996), while in all other orchards the values were clearly higher than the adequate level. Similar results of Mn excess in the mango leaves were also reported by Assis et al. (2004), Rozane et al. (2007), Pinto et al. (2009), and Galli et al. (2009). The high levels may be explained by the high Mn concentration in the soil and by spraying with Mn-based chemical products to control pests and diseases.

The Zn levels in the leaves varied from 7.76 to 37.71 mg kg⁻¹ (Table 2). The areas after 11, 14, and 16 years of cultivation had Zn contents below the level considered deficient (Quaggio, 1996). Only in the areas after 19 and 26 years of cultivation, the Mn contents were within the adequate content range. In the other orchards, the Zn contents varied between deficient and adequate. Pereira et al. (2005) considered Zn to be one of the most important micronutrients for mango. Zn deficiency can cause flower malformation or "Silking" and vegetative malformation or "Witches' Broom", caused by the emission of multiple, small, irregularly shaped, deformed panicles (Silva et al., 2004).

The leaf concentrations of Fe varied from 48.13 to 110.17 mg kg⁻¹ (Table 2). Only in the 9 year-old

Table 2. Heavy metal contents in leaves of mango trees, grown in northeastern Brazil for different cultivation periods

| Cultivation periods | Cu | Ni | Mn | Zn | Fe | Cr |
|---------------------|-----------------------------|------|------------|-----------|------------|------|
| years | mg kg ⁻¹ | | | | | |
| 6 | 22.47 | UD | 503.47 | 17.36 | 55.44 | 2.32 |
| 7 | 32.07 | 0.59 | 53.87 | 11.15 | 62.83 | 1.85 |
| 9 | 63.73 | 0.47 | 250.67 | 12.99 | 48.13 | 1.92 |
| 10 | 10.47 | 0.05 | 429.07 | 19.08 | 80.52 | 4.85 |
| 11 | 16.31 | UD | 212.53 | 9.45 | 110.17 | 3.84 |
| 14 | 4.29 | UD | 569.60 | 7.76 | 76.52 | 3.03 |
| 16 | 9.35 | UD | 358.80 | 7.97 | 65.84 | 1.56 |
| 17 | 9.75 | 1.24 | 408.67 | 11.95 | 76.27 | 0.87 |
| 19 | 19.79 | UD | 67.57 | 37.71 | 56.68 | 0.00 |
| 26 | 8.44 | UD | 135.87 | 25.81 | 60.47 | 0.00 |
| | Range levels ⁽¹⁾ | | | | | |
| Deficient | <5 | * | <10 | <10 | <15 | * |
| Appropriate | 10.0-50.0 | * | 50.0-100.0 | 20.0-40.0 | 50.0-200.0 | * |
| Excessive | - | * | - | >100 | - | * |

UD: Undetectable; * no content range for these elements; - excessive levels for these elements. ⁽¹⁾Quaggio (1996).

orchard, the Fe levels were below the adequate range (Quaggio, 1996). All other areas had Fe levels within the adequate range.

The Cr levels in the leaves varied from UD to 4.85 mg kg⁻¹ (Table 2). Only in the 19 and 26 year-old orchards, no Cr was detected in the mango leaves. There are no reports on the effect of Cr toxicity on mango trees because the toxic effect of Cr in plants varies according to its oxidation state (Cr³⁺ or Cr⁶⁺) and the tolerance of the species (Castilhos et al., 2001). According to Losi et al. (1994), toxic effects of Cr are common in most plants when the foliar concentration of the element exceeds 18 mg kg⁻¹.

CONCLUSIONS

1. The agricultural management practices adopted by the mango-producing companies promoted changes in the heavy metal concentrations of the soils when compared with the soils of the reference areas.

2. In general, the cultivation of mangos caused a surface accumulation of Cu and Zn in the soil and a reduction in the contents of Ni, Pb, Mn, and Fe in both the surface and subsurface.

3. Contamination with Cu, Zn, and Cr was detected in the mango orchards, indicating the need of monitoring, in order to minimize the environmental impact. For instance, the presence of Cr in the leaves of the mango trees shows the need to investigate the source of the element.

4. The different management strategies of the different companies led to a deficiency or excess of some metals that are also micronutrients (Cu, Zn, Fe, Mn) in the evaluated areas. However, the Fe and Mn levels were adequate for the mineral nutrition of mango in all areas.

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