



BIOLOGICAL SCIENCES

Diversity and turnover in a rocky shore intertidal community of an upwelling region (Arraial do Cabo, Brazil)

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Abstract: The present study describes the community diversity and gives a seasonal distribution of an intertidal macroalgal assemblage at Prainha Beach, Arraial do Cabo City in Rio de Janeiro state, Brazil. Sampling was performed during four seasons of a 1-year period. Organisms were sampled and photographed using photoquadrats to quantify relative coverage. Ninety-six taxa of macroalgae were registered, including 23 chlorophytes, 19 ochrophytes, and 54 rhodophytes, while three barnacles, seven mollusks, one isopod, and one polychaete were recorded among the zoobenthos. In the upper intertidal zone, the coverage was frequently dominated by invertebrates. Macroalgae always dominated the middle and lower zones, covering almost 100 % in both zones. The highest values of species diversity (H'), richness (S), and turnover rates were found during the transition from spring to summer, which coincided with the upwelling period. A comparison with the 1980's flora revealed that the major changes in the macroalgal assemblage were among species belonging to the same genera, unless new additions were provided. The combination of photoquadrats and minimally destructive sampling allowed a detailed description of the composition and structural characteristics of the intertidal zone, a methodology that should be applied to study protected marine areas.

Key words: intertidal benthic communities, macroalgae, southwestern atlantic ocean, seasonal variability, upwelling system.

INTRODUCTION

Rocky shores are dynamic coastal environments that are considered to be zones of transition between aquatic and terrestrial ecosystems and are inhabited predominantly by specially adapted marine organisms (Ros et al. 1985). The diversity and productivity of rocky shores are associated with the presence of marine macroalgae, which are the main primary producers and serve as substrate, shelter, and reproduction sites for many vertebrates and marine invertebrates (Széchy et al. 2001, Tano et al. 2016). Besides, the presence of a variety of organisms there may serve as indicators of environmental conditions

(Murray et al. 2006, Pinedo et al. 2007, Borja et al. 2012). For instance, Ulvophyceae may present higher coverage or be favored in anthropogenic disturbed sites (Faveri et al. 2010, Teichberg et al. 2010, Scherner et al. 2013), while the coverage of Phaeophyceae is reduced or locally absent under these conditions (Oliveira & Qi 2003, Menconi et al. 2012, Széchy et al. 2017). Zonation patterns and natural fluctuations are also broadly investigated concepts at different rocky shores worldwide (Masi et al. 2009, Gómez & Huovinen 2011, Stevčić et al. 2017, Vinagre et al. 2017, Little et al. 2018).

The coast of Brazil's Rio de Janeiro state is on the eastern coast of South America and is at

the edge of two marine provinces: the Tropical Southwestern Atlantic and the Warm Temperate Southwestern Atlantic (Spalding et al. 2007). The region of Arraial do Cabo within this state is often treated as part of the border between these provinces, and because of this, Prainha Beach (within this region) can be considered an ecotone area. This area is seasonally under the influence of upwelling by the South Atlantic Central Water (SACW), which constitutes a discontinuous biogeographic barrier for marine fauna and flora in the Southwestern Atlantic and is thus recorded as the northern or southern limit of many species' distributions (Yoneshigue 1985, Guimaraens & Coutinho 1996, Floeter & Soares-Gomes 1999). The fisherman population of Arraial do Cabo demanded the establishment of a marine reserve to protect the integrity of the marine biodiversity in this area, which is a locally important economic resource (Mendonça et al. 2013). Thus, the Reserva Extrativista Marinha do Arraial do Cabo (RESEX) was established in 1997 to assure the sustainable utilization of marine resources by the traditional fishermen populations and the recreational and other activities under management norms (Silva 2004, ICMBIO 2017). These particular conditions support the high biodiversity of species with tropical and temperate affinities in the region, leading to the establishment of several lists and spatiotemporal comparisons of macroalgae in this area, particularly those of Yoneshigue (1985), Guimaraens & Coutinho (1996), and Brasileiro et al. (2009).

Several methodologies have been used to track the diversity of the marine organisms (Gomèz & Huovinen 2011, Trivedi et al. 2016, Perera-Valderrama et al. 2017, Rohde et al. 2017), however, the use of photographic sampling has arisen recently as a particularly quick and non-destructive alternative method to traditional destructive sampling techniques (Solan et al.

2003, Durden et al. 2016, Beisiegel et al. 2018). This method, allied to a new generation of software, facilitates statistical analyses and reporting of the occurrence of species in several habitats (Tanner et al. 2015, Arefin 2016, Gomes-Pereira et al. 2016, Romero-Ramirez et al. 2016).

Combining evidence from the RESEX implemented in Arraial do Cabo, including its oligotrophic status (Guimaraens & Coutinho 2000, Jara et al. 2006, Coelho-Souza et al. 2017) and the concept of natural fluctuations induced by environmental and biological factors (Hoek 1982, Ives & Carpenter 2007, Nascimento et al. 2014), we hypothesized that the species of marine flora at Prainha has changed over the years by its ecological equivalent (i.e. species with similar ecological function) and, consequently, there has been a large turnover of the species in this area. Thus, the aims of the present study were to (I.) analyze the diversity of an intertidal community of a rocky shore at Prainha and the turnover of its macroalgae assemblage throughout all four seasons, (II.) assess the change in the macroalgal composition of this community over an interval of almost 30 years, and (III.) evaluate the efficiency of different combinations of destructive and photographic approaches to characterize the species coverage and diversity of this rocky shore benthic community.

MATERIALS AND METHODS

Sampling was performed seasonally between 2011 and 2012 at Prainha Beach, Arraial do Cabo city, Rio de Janeiro state, Brazil (22°57'22.1"S, 42°01'36.5"W). It was an integrative approach of a non-destructive method with a minor destructive method and always performed at low tide on the left side of the rocky shore located outside the beach. The area sampled was approximately 10 m in length, which was horizontally delimited

with a line transect above the intertidal region and divided into three vertical zones (upper, middle, and lower) based on the typically predominant organisms found in each zone. At each vertical zone, five 900 cm² quadrats were randomly placed for photographic sampling. In addition, we manually collected macroalgae specimens from the transect area and those living around the rocky shore (outside the transect area) within a 40 min period to assist identification of local biodiversity, none marine invertebrates were collected. These collections were done to obtain estimates of the species richness and turnover while also sustaining the photoquadratic identifications. Specimens were collected, separated according to their intertidal zone, and preserved in a solution of 4 % formalin diluted in seawater for later identification.

At the laboratory, macroalgal taxa were identified to the lowest level possible with Olympus CX40 optical microscopy and Olympus SZ51 stereoscopic microscope, following Taylor (1972), Littler & Littler (2000) and Dawes & Mathieson (2008). Nomenclatural updates followed Flora 2020 (<http://floradobrasil.jbrj.gov.br>) and AlgaeBase (Guiry & Guiry 2019). The resulting list of macroalgae was compared to the data provided by Yoneshigue (1985) (Table I) to the same locality. Species that were considered as new occurrences for the area were deposited in the Herbarium Jorge Pedro Pereira Carauta (HUNI). The turnovers between seasons in the present study and between the species list from the current study and that proposed by Yoneshigue (1985), were calculated based on the concept of extinction (absence of species) and immigration (appearance of species) of all species along the shore throughout time as suggested in Brown & Kodrick-Brown (1977) and Magurran (2006).

The coverage of the community by each taxon was analyzed using Coral Point Count with

Excel extensions (CPCe) (Kohler & Gill 2006) via an adapted database containing the names of the organisms in the studied region. When possible, identification of organisms from the photoquadrats was performed to the species level. However, it was not possible to identify small organisms to the species level using photographic analysis, in which case they were identified to the morphological or order level. In the coverage analysis, all taxa identified were treated as Community Components (CCs) based on their taxonomic or morphological level. Thirty points were randomly placed on each photograph for coverage analysis. The number of points needed to adequately represent the coverage of the community was established according to Silva et al. (2015). Afterwards, Shannon-Wiener (H') and Evenness (J) diversity indices were calculated based on the coverage of each and all species (Magurran 2006).

RESULTS

Coverage analysis

The upper zone of the rocky shore was dominated by invertebrates, which made up more than half of the CCs in all seasons, and the invertebrate cover was higher during winter, summer, and autumn, whereas algae had a higher coverage than invertebrates in the spring (Figure 1). For instance, *Tetraclita stalactifera* Lamarck 1818 (TET) was the most abundant species during the winter, spring, and summer and *Mytilaster solisianus* d'Orbigny 1842 (MYT) in autumn, winter and summer. In autumn, however, the crustose non-coralline rhodophyte *Hildenbrandia rubra* (Sommerfelt) Meneghini (HIL) was the most abundant species. Community components with less percentage coverages were identified during the year and included taxa such as TET, *Amphibalanus amphitrite* Darwin 1854 (AMA), *Perna perna* Linnaeus 1758

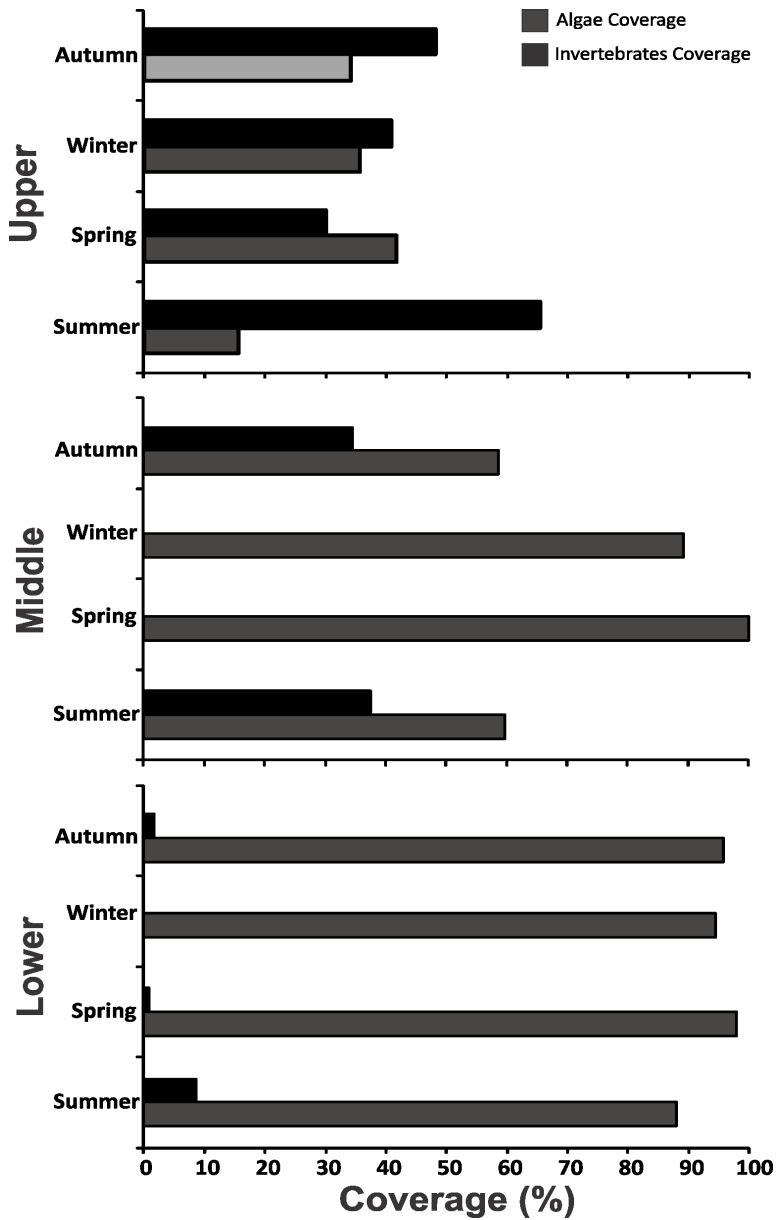


Figure 1. Sum of coverage Mean of Macroalgae and Invertebrate in each intertidal zone.

(PER), *Chthamalus bisinuatus* Pilsbry 1916 (CHT), crustose Corallinales and laminar forms of *Ulva* spp. (including *Ulva fasciata* Delile and *Ulva rigida* C. Agardh). Spring was the season with the highest diversity (16 CCs), while autumn had the lowest (10 CCs) (Figure 2).

In the middle and lower zones, macroalgae were the dominant organisms on the substrate throughout the year (Figure 1). The most important CC with the highest coverage was the

group of articulated Corallinales (ACO), however, other algae were also found in both zones, e.g., laminar forms of *Ulva* spp., crustose Corallinales (CCO), *Padina* spp. (PAD), *Colpomenia sinuosa* (Mertens ex Roth) Derbès & Solier (COL), and *Hypnea* spp. (HYP). Some invertebrates that were characteristic of the upper zone, such as TET, AMA and PER were also observed in the middle zone during the summer (Figure 2).

Table I: Comparison of intertidal species list from Prainha by Yoneshigue (1985) and the current study. Species that were listed only in the present study are highlighted with bold, species registered within Prainha by Yoneshigue (1985) and by the present study are also marked with “*” and species listed only by Yoneshigue (1985) are not in bold. The species recognized by CPCe received the classification codes as follow: “ACO” – Articulated Corallinales; “CEG” – *Centroceras gasparrinii*; “FRH” – Filamentous Rhodophyta; “GEL” – Gelidiales; “GYM” – *Gymnogongrus griffithsiae*; “HIL” – *Hildenbrandia rubra*; “HYP” – *Hypnea* spp.; “LAU” – *Laurencia* spp.; “TRH” – Turfs of Rhodophyta; “CAN” – *Canistrocarpus cervicornis*; “COL” – *Colpomenia sinuosa*; “FOC” – Filamentous Ochrophyta; “CHA” – *Chaetomorpha antennina*; “ENT” – Tubular Ulva; “ULV” – Foliose Ulva; “FCH” – Filamentous Chlorophyta; “TCH” – Turfs of Chlorophyta; “COE” – Erect Codium; “COC” – Crustose Codium; “AMA” – *Amphibalanus amphitrite*; “CHT” – *Chthamalus bisinuatus*; “FIS” – *Fissurella clenchi*; “ISO” – *Isognomon bicolor*; “LOT” – *Lottia subrugosa*; “MYT” – *Mytilaster solisianus*; “OST” – Ostreidae; “PER” – *Perna perna*; “PHR” – *Phragmatopoma caudata*; “STA” – *Stramonita haemastoma*; “-” Not Identified in the Photography; “ND” No data, i.e., not found during this study.

	Community Components
Rhodophyta	
Stylonemataceae	
<i>Bangiopsis dumontioides</i> (P.L.Crouan & H.M.Crouan) V.Krishnmurthy	-
<i>Stylonema alsidii</i> (Zanardini) K.M.Drew *	-
Erythrotrichiaceae	
<i>Erythrotrichia carnea</i> (Dillwyn) J.Agardh *	-
<i>Sahlingia subintegra</i> (Rosenvinge) Kornmann*	-
Bangiaceae	
<i>Bangia fuscopurpurea</i> (Mertens ex Roth) C.Agardh	ND
<i>Porphyra pujalsiae</i> Coll & E.C.Oliveira *	-
<i>Pyropia acanthophora</i> (E.C.Oliveira & Coll) M.C.Oliveira, D.Milstein & E.C.Oliveira	ND
Colaconemataceae	
<i>Colaconema codicola</i> (Børgesen) Stegenga, J.J.Bolton & R.J.Anderson	-
Acrochaetiaceae	
<i>Acrochaetium crassipes</i> (Børgesen) Børgesen	ND
<i>Acrochaetium flexuosum</i> Vickers	ND
<i>Acrochaetium globosum</i> Børgesen	ND
<i>Acrochaetium microscopicum</i> (Nägeli ex Kützing) Nägeli in Nägeli & Cramer	-
Gelidiellaceae	
<i>Millerella pannosa</i> (Feldmann) G.H.Boo & L.Le Gall	ND
Gelidiaceae	
<i>Gelidium pusillum</i> (Stackhouse) Le Jolis *	GEL
Pterocliadiaceae	
<i>Pterocliadiella capillacea</i> (S.G.Gmelin) Santelices & Hommersand *	TRH
<i>Pterocliadiella caerulescens</i> (Kützing) Santelices & Hommersand	TRH
Bonnemaisoniaceae	
<i>Asparagopsis taxiformis</i> (Delile) Trevisan	FRH

Table I (continuation)

Plocamiaceae		
	<i>Plocamium brasiliense</i> (Greville) M. Howe & W.R.Taylor	ND
Lomentariaceae		
	<i>Ceratodictyon variabile</i> (J.Agardh) R.E.Norris	ND
	<i>Lomentaria rawitscheri</i> A.B.Joly	ND
Gracilariaceae		
	<i>Gracilaria cervicornis</i> (Turner) J. Agardh	ND
	<i>Gracilaria mammillaris</i> (Montagne) M.A.Howe	-
Cystocloniaceae		
	<i>Hypnea cervicornis</i> J. Agardh*	HYP
	<i>Hypnea pseudomusciformis</i> Nauer, Cassano & M.C.Oliveira *¹	HYP
	<i>Hypnea spinella</i> (C. Agardh) Kützing *	HYP
Phylloporaceae		
	<i>Gymnogongrus griffithsiae</i> (Turner) C. Martius*	GYM
Gigartinaceae		
	<i>Chondracanthus acicularis</i> (Roth) Fredericq	CHO
	<i>Chondracanthus saundersii</i> C.W.Schneider & C.E.Lane	CHO
	<i>Chondracanthus teedei</i> (Mertens ex Roth) Kützing *	CHO
Corallinaceae		
	<i>Amphiroa beauvoisii</i> J.V.Lamouroux *	ACO
	<i>Amphiroa brasiliana</i> Decaisne	ACO
	<i>Amphiroa fragilissima</i> (Linnaeus) J.V. Lamouroux	ND
	<i>Arthrocardia flabellata</i> (Kützing) Manza	ND
	<i>Arthrocardia variabilis</i> (Harvey) Weber-van Bosse	ACO
	<i>Corallina officinalis</i> Linnaeus	ND
	<i>Jania adhaerens</i> J.V. Lamouroux *	ACO
	<i>Jania capillacea</i> Harvey *	ACO
	<i>Jania rubens</i> (Linnaeus) J.V. Lamouroux	ACO
	<i>Pneophyllum fragile</i> Kützing	ND
Hildenbrandiaceae		
	<i>Hildenbrandia rubra</i> (Sommerfelt) Meneghini *	HIL
Champiaceae		
	<i>Champia parvula</i> (C.Agardh) Harvey	TRH
	<i>Champia vieillardii</i> Kützing	TRH
	<i>Gastroclonium parvum</i> (Hollenberg) C.F.Chang & B.M.Xia	TRH
Rhodymeniaceae		
	<i>Rhodymenia Pseudopalmata</i> (Lamouroux) Silva	ND

Table I (continuation)

Callithamniaceae	
<i>Aglaothamnion uruguayense</i> (W.R.Taylor) N.E.Aponte, D.L.Ballantine & J.N.Norris	ND
Ceramiaceae	
<i>Centroceras clavulatum</i> (C. Agardh) Montagne	ND
<i>Centroceras gasparrinii</i> (Meneghini) Kützing	CEG
<i>Centrocerocolax ubatubensis</i> A.B. Joly *	-
<i>Ceramium brasiliense</i> A.B. Joly *	TRH
<i>Ceramium brevizonatum</i> var. <i>caraibicum</i> H.E.Petersen & Børgesen in Børgesen	TRH
<i>Ceramium clarionense</i> Setchell & N.L.Gardner	TRH
<i>Ceramium codii</i> (H. Richards) Mazoyer	ND
<i>Ceramium dawsonii</i> A.B. Joly *²	TRH
<i>Ceramium lutzburgii</i> O.C. Schmidt	ND
<i>Ceramium tenerrimum</i> (G. Martens) Okamura *	TRH
<i>Ceramium vagans</i> P.C.Silva in P.C. Silva, Meñez & Moe	TRH
<i>Gayliella mazoyerae</i> T.O.Cho, Fredericq & Hommersand ³	ND
<i>Gayliella</i> sp. ⁴	ND
Wrangeliaceae	
<i>Anotrichium tenue</i> (C.Agardh) Nägeli	ND
<i>Wrangelia argus</i> (Montagne) Montagne	-
Spyridiaceae	
<i>Spyridia hypnoides</i> (Bory) Papenfuss *	TRH
Delesseriaceae	
<i>Cryptopleura ramosa</i> (Hudson) Kylin ex Newton	ND
<i>Taenioma perpusillum</i> (J.Agardh) J.Agardh	-
Dasyaceae	
<i>Dasya brasiliensis</i> E.C.Oliveira Filho & Y.Y.Braga	ND
<i>Dasya corymbifera</i> J. Agardh	ND
<i>Dasya ocellata</i> (Grateloup) Harvey in Hooker	FRH
<i>Dasya rigidula</i> (Kützing) Ardissona	TRH
Rhodomelaceae	
<i>Bostrychia binderi</i> Harvey	ND
<i>Bryocladia thyrSIGera</i> (J.Agardh) F.Schmitz in Falkenberg*	TRH
<i>Chondria atropurpurea</i> Harvey	TRH
<i>Herposiphonia bipinnata</i> M. Howe	-
<i>Herposiphonia delicatula</i> Hollenberg	-
<i>Herposiphonia nuda</i> Hollenberg	-
<i>Herposiphonia parca</i> Setchell	-

Table I (continuation)

<i>Herposiphonia secunda</i> (C Agardh) Ambronn	ND
<i>Herposiphonia secunda</i> var. <i>tenella</i> (C. Agardh) M.J. Wynne	ND
<i>Laurencia aldingensis</i> Saito & Womersley	LAU
<i>Laurencia dendroidea</i> J.Agardh	LAU
<i>Melanothamnus ferulaceus</i> (Suhr ex J.Agardh) Díaz-Tapia & Maggs	ND
<i>Melanothamnus gorgoniae</i> (Harvey) Díaz-Tapia & Maggs	-
<i>Melanothamnus sphaerocarpus</i> (Børgesen) Díaz-Tapia & Maggs	-
<i>Murrayella pericladus</i> (C. Agardh) F. Schmitz	ND
<i>Polysiphonia scopulorum</i> Harvey	-
<i>Polysiphonia villum</i> J.Agardh	ND
<i>Xiphosiphonia pennata</i> (C.Agardh) Savoie & G.W.Saunders *	-
Ochrophyta	
Acinetosporaceae	
<i>Feldmannia indica</i> (Sonder) Womersley & A.Bailey	-
<i>Feldmannia irregularis</i> (Kützing) G.Hamel *	-
<i>Feldmannia mitchelliae</i> (Harvey) H.-S.Kim *	-
Bachelotiaceae	
<i>Bachelotia antillarum</i> (Grunow) Gerloff	-
Asteronemataceae	
<i>Asteronema breviararticulatum</i> (J.Agardh) Ouriques & Bouzon *	FOC
Ectocarpaceae	
<i>Ectocarpus fasciculatus</i> Harvey	FOC
<i>Ectocarpus rallsiae</i> Vickers	ND
Chordariaceae	
<i>Levringia brasiliensis</i> (Montagne) A.B.Joly *	FOC
Neoralfsiaceae	
<i>Neoralfsia expansa</i> (J.Agardh) P.-E.Lim & H.Kawai ex Cormaci & G.Furnari	ND
Scytosiphonaceae	
<i>Chnoospora minima</i> (Hering) Papenfuss	FOC
<i>Colpomenia sinuosa</i> (Mertens ex Roth) Derbès & Solier *	COL
<i>Rosenvingea orientalis</i> (J.Agardh) Børgesen	FOC
<i>Rosenvingea sanctae-crucis</i> Børgesen *	FOC
Sphacelariaceae	
<i>Sphacelaria brachygonia</i> Montagne	ND
<i>Sphacelaria novae-hollandiae</i> Sonder	-
<i>Sphacelaria tribuloides</i> Meneghini	-
Dictyotaceae	
<i>Canistrocarpus cervicornis</i> (Kützing) De Paula & De Clerck	CAN

Table I (continuation)

<i>Dictyopteris delicatula</i> Lamouroux	-
<i>Dictyota menstrualis</i> (Hoyt) Schnetter, Hörning & Weber-Peukert ⁵	ND
<i>Padina antillarum</i> (Kützinger) Piccone	PAD
<i>Padina gymnospora</i> (Kützinger) Sonder *	PAD
Sargassaceae	
<i>Sargassum furcatum</i> Kützinger *	-
<i>Sargassum vulgare</i> C.Agardh	-
Chlorophyta	
Ulvaceae	
<i>Ulva compressa</i> Linnaeus *	ENT
<i>Ulva fasciata</i> Delile	ND
<i>Ulva flexuosa</i> Wulfen *	ENT
<i>Ulva lactuca</i> Linnaeus	ULV
<i>Ulva linza</i> Linnaeus	ENT
<i>Ulva prolifera</i> O.F.Müller	ENT
<i>Ulva rigida</i> C. Agardh *	ULV
Ulvellaceae	
<i>Ulvella viridis</i> (Reinke) R.Nielsen, C.J.O'Kelly & B.Wyosor	ND
Kornmanniaceae	
<i>Pseudendoclonium marinum</i> (Reinke) Aleem & E. Schulz	ND
Cladophoraceae	
<i>Chaetomorpha aerea</i> (Dillwyn) Kützinger	-
<i>Chaetomorpha antennina</i> (Bory) Kützinger *	CHA
<i>Chaetomorpha brachygona</i> Harvey *	-
<i>Cladophora rupestris</i> (Linnaeus) Kützinger	ND
<i>Cladophora sericea</i> (Hudson) Kützinger	FCH
<i>Cladophora vagabunda</i> (Linnaeus) Hoek *	FCH
<i>Rhizoclonium riparium</i> (Roth) Harvey	FCH
<i>Willeella brachyclados</i> (Montagne) M.J.Wynne *	FCH
Valoniaceae	
<i>Valonia macrophysa</i> Kützinger	ND
Boodleaceae	
<i>Cladophoropsis macromeres</i> W.R.Taylor	TCH
<i>Phyllocladon anastomosans</i> (Harvey) Kraft & M.J.Wynne	TCH
Bryopsidaceae	
<i>Bryopsis corymbosa</i> J. Agardh	ND
<i>Bryopsis pennata</i> Lamouroux	FCH

Table I (continuation)

<i>Bryopsis plumosa</i> (Hudson) C. Agardh	FCH
<i>Bryopsis ramulosa</i> Montagne	FCH
Derbesiaceae	
<i>Derbesia marina</i> (Lyngbye) Solier	ND
<i>Derbesia tenuissima</i> (Moris & De Notaris) P.Crouan & H.Crouan	ND
Codiaceae	
<i>Codium decorticatum</i> (Woodward) M. Howe *	COE
<i>Codium intertextum</i> Collins & Hervey *	COE
<i>Codium isthmocladum</i> Vickers	ND
<i>Codium spongiosum</i> Harvey *	COC
<i>Codium taylorii</i> P.C.Silva *	COE
Caulerpaceae	
<i>Caulerpa racemosa</i> (Forsskål) J.Agardh	-
¹ reported as <i>Hypnea musciformis</i> (Wulfen) J.V. Lamouroux by Yoneshigue (1985)	
² reported as <i>Ceramium taylorii</i> E.Y. Dawson by Yoneshigue (1985)	
³ reported as <i>Ceramium gracillimum</i> var. <i>byssoideum</i> Mazoyer by Yoneshigue (1985)	
⁴ following Wynne (2017), <i>Ceramium flaccidum</i> in Brazil must be considered a non-identified <i>Gayliella</i> species.	
⁵ reported as <i>Dictyota dichotoma</i> (Hudson) J.V. Lamouroux by Yoneshigue (1985)	
Invertebrates	
Arthropoda	
Tetraclitidae	
<i>Tetraclita stalactifera</i> (Lamarck, 1818)	TET
Chthamalidae	
<i>Chthamalus bisinuatus</i> Pilsbry, 1916	CHT
Balanidae	
<i>Amphibalanus amphitrite</i> (Darwin, 1854)	AMA
Ligiidae	
<i>Ligia exotica</i> Roux, 1828	LYG
Mollusca	
Muricidae	
<i>Stramonita haemastoma</i> (Linnaeus 1787)	STA
Fissurellidae	
<i>Fissurella clenchi</i> Pérez Farfante, 1943	FIS
<i>Lottia subrugosa</i> (d'Orbigny, 1846)	LOT
Mytilidae	
<i>Mytilaster solisianus</i> (d'Orbigny, 1842) *	MYT
<i>Perna perna</i> (Linnaeus, 1758) *	PER

Table I (continuation)

Ostreidae	
<i>Cassostrea rhizophorae</i> (Guilding, 1828)	OST
Pteriidae	
<i>Isognomon bicolor</i> (C.B. Adams, 1845)	ISO
Annelida	
Sabellariidae	
<i>Phragmatopoma caudata</i> Krøyer in Mörch, 1863	PHR

In comparison to other seasons and zones, the coverage of the middle zone during the summer presented the highest total number of CCs, making it the most diverse assemblage in terms of coverage found throughout the year (Figure 2). During winter and spring, the middle zone was dominated by articulated Corallinales (ACO, 36-54 % coverage), while invertebrates were absent. Afterwards, in summer and autumn, the dominance of articulated Corallinales was replaced by *Chaetomorpha antennina* (Bory) Kützing (CHA, 18.5 %) and HIL (25.7 %), and the invertebrates returned and made up for more than half of the CCs registered.

In the lower zone, macroalgae accounted for more than 87 % of the coverage (Figure 1). The ACO were the prevailing CC for the whole year. The principal component of this algal group was *Jania crassa* J. V. Lamouroux, but it also included *Amphiroa beauvoisii* J. V. Lamouroux to a lesser extent. The number of registered CCs in the lower zone went from eight during the winter up to 14 in autumn, with invertebrates accounting for two CCs or less during the entire period sampled (Figure 2).

Macroalgal assemblage richness, diversity, and turnover

The Shannon diversity index values obtained demonstrated that the richness and homogeneity of the species present decreased between summer and autumn in the upper

zone (Figure 3). Between spring and summer, the richness in the middle zone increased considerably, although there was no concurrent change in the equitability (Figure 3). This change in the number of species was concomitant with the highest turnover rate for the whole year (Figure 4d). In the transition between spring and summer, six species were not observed, and eight new occurrences appeared in the macroalgal assemblage. This variation was greater than the annual turnover in the region, and thus this was the period out of the year with the largest replacement of organisms (Figure 4d).

In the lower zone, the richness and equitability exhibited little variation, which was reflected in the fact that the diversity of this zone was similar in all seasons of the year (Figure 3). Regarding the turnover rate between winter and spring, eight new species were recorded, while seven were not observed on the rocky shore after this period (Figure 4d).

In terms of the turnover of phyla, the turnover rate of Chlorophyta ranged from 0.27 to 0.42, with the largest change occurring between spring and summer (Figure 4a). Rhodophyta exhibited high turnover rates (> 0.50) for all seasons, indicating constant and major shifts in the composition of red macroalgal assemblages (Figure 4b). Comparing the present study and the former (Yoneshigue 1985) community compositions were compared, green algae were also found to have a lower turnover rate than

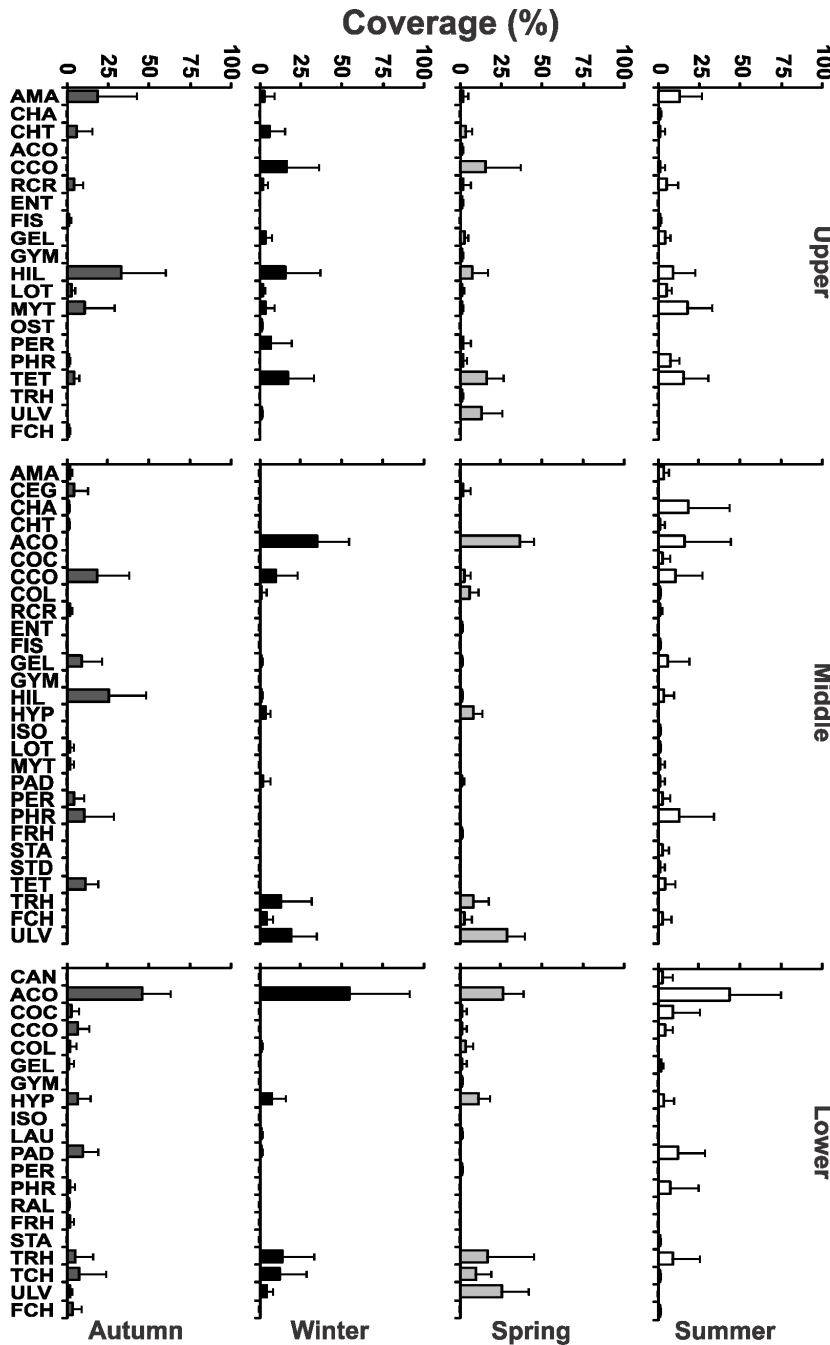


Figure 2. Percent cover of different community components (CC) in each zone. Values presented are means \pm standard deviations.

AMA = *Amphibalanus amphitrite*; CAN = *Canistrocarpus cervicornis*; CEG = *Centroceras gasparrinii*; CHA = *Chaetomorpha antennina*; CHT = *Chthamalus bisinuatus*; ACO = Articulated Corallinales; COC = *Codium spongiosum*; CCO = Crustose Corallinales; COL = *Colpomenia sinuosa*; RCR = Recruit Cirripedia; ENT = Tubular *Ulva* spp.; FCH = Filamentous Chlorophyta; FIS = *Fissurella clenchi*; FRH = Filamentous Rhodophyta; GEL = Gelidiales; GYM = *Gymnogongrus griffithsiae*; HIL = *Hildenbrandia rubra*; HYP = *Hypnea* spp.; ISO = *Isognomon bicolor*; LAU = *Laurencia* spp.; LOT = *Lottia subrugosa*; MYT = *Mytilaster solisianus*; OST = *Cassostrea rhizophorae*; PAD = *Padina* spp.; PER = *Perna perna*; PHR = *Phragmatopoma caudata*; RAL = *Neoralgsia expansa*; STA = *Stramonita haemastoma*; STD = *Stramonita haemastoma spawning*; TCH = Turfs of Chlorophyta; TET = *Tetraclita stalactifera*; TRH = Turfs of Rhodophyta; ULV = Laminar *Ulva* spp.

Rhodophyta (Figure 4e). However, Ochrophyta was a phylum that underwent lower annual variations than others (Figure 4c), nevertheless, it experienced a higher turnover during the transitions from spring to summer and winter to spring. Comparing to Yoneshigue (1985) dataset, Ochrophyta also showed the lowest turnover

rate, with low exclusion and immigration rates compared to those of other phyla (Figure 4c).

Macroalgal comparison

Yoneshigue (1985) registered 85 species as occurring at Prainha Beach, of which 20 were species of Chlorophyta, 12 were Ochrophyta, and

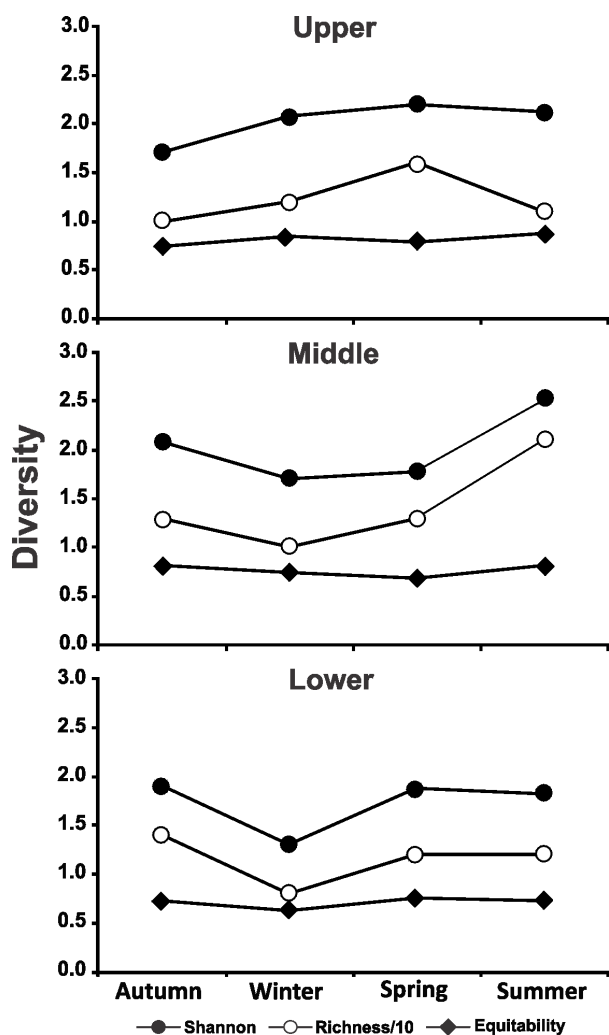


Figure 3. Diversity indices calculated for the upper, middle, and lower zones of the rocky shore at Prainha.

53 were Rhodophyta. In addition, the review of Brasileiro et al. (2009) registered 207 taxa for Arraial do Cabo, of which Prainha had 98 species and was the second most diverse location in the region after Ponta da Cabeça (108 species). The present study found 96 taxa at Prainha, of which 23 belonged to the phylum Chlorophyta, 19 to Ochrophyta, and 54 to Rhodophyta, with the last phylum remaining the most well-represented at Prainha.

Rhodomelaceae and Ceramiaceae (Rhodophyta), were the families with the higher diversity at Prainha, with 12 species (22 % of the

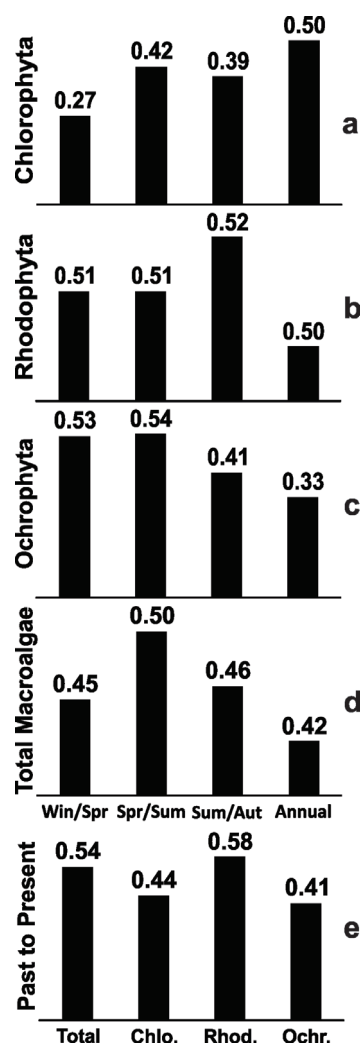


Figure 4. Turnover of taxa found on the rocky shore at Prainha. Turnover is shown for Chlorophyta, Rhodophyta, Ochrophyta, all taxa of macroalgae, and is compared between Yoneshigue (1985) and the present study.

total) and eight species (15 %), respectively. The most well-represented families of Chlorophyta at the study site were Cladophoraceae and Ulvaceae, with seven and six species present, respectively. Comprising together 57 % of the total abundance of the phyla observed. Among Ochrophyta, two families comprised most of the composition: Dictyotaceae, which increased its representation from 15 to 21 % of the total (four species) over time; and Scytosiphonaceae, which increased its representation (two

species) in the community in comparison to that in the early 1980s.

Altogether, compared to past surveys (Yoneshigue 1985, Brasileiro et al. 2009) 55 new occurrences were added for Prainha Beach and 27 for Arraial do Cabo. Considering these new occurrences, Prainha reached a total of 125 taxa, which would make it the most diverse area in Arraial do Cabo. These new records also increased the number of taxa present in Arraial do Cabo, enhancing it from 207 to 234 species.

DISCUSSION

Autumn and winter in this region are periods that are marked by the frequent occurrence of cold fronts and storms. As a result, the water column is mixed in these seasons, increasing the amount of organic and inorganic particles it contains (Coelho-Souza et al. 2017) and causing the senescence of a large part of the coastal flora (Riggs et al. 1998). In addition to these physical disturbances, the prevalence of south-southwest quadrant winds favors the proximity and residence of the Brazilian Current along the coast and prevents upwelling of the SACW. The proximity of the Brazilian Current could lead to inorganic nutrient stress and/or limitation, especially for organisms in the upper zone of the coast, resulting in the decreased abundance of several components of the community and increasing the turnover rates, as was observed during the present study. Under nutrient limitation, resistant macroalgae with shorter canopies are selected (Steneck & Dethier 1994). The same seasonal effects on the benthic community were observed at Ubatuba (state of São Paulo, Brazil), another upwelling region that is influenced by the SACW (Alves et al. 2014). On the other hand, during spring and summer, upwelling events increase in frequency

and intensity at Arraial do Cabo. The water column stratifies more frequently and becomes enriched with inorganic nutrients, such as nitrate and phosphate (Coelho-Souza et al. 2012, 2017). The increased input of such inorganic nutrients could explain the rising number of macroalgal species in the middle and lower zones of the coast that we observed, as well their turnover rates during the beginnings of these periods.

The interactions between macroalgae and invertebrates over the three zones of the intertidal community of Prainha observed in the present study were similar to those described by Yoneshigue (1985), as well as to those in other upwelling regions of the world (Hosegood et al. 2017, Jiang & Wang 2018, Walter et al. 2018). However, some exceptions were observed, such as the differences in the species dominating the upper zone throughout all seasons, wherein *Chthamalus bisinuatus* Pilsbry, 1916 dominated community coverage in the early 1980s (Yoneshigue 1985), while *Tetraclita stalactifera* and *Hildenbrandia rubra* dominated in the present study. The results for the middle and lower zones agreed between this and past studies in relation to species dominance, with one exception during autumn; specifically, Yoneshigue (1985) recorded a large settlement of crustose Corallinales algae in the lower zone in the autumn, whereas our study observed that the complex thalli of articulated Corallinales dominated this season.

Previous studies emphasized the influence of upwelling on the megafaunistic composition of marine communities at other sites around the world (e.g., Lavaleye et al. 2002, Kelaher & Castilla 2005, Carlier et al. 2015). At Cabo Frio, Brazil, in particular, most species have subtropical affinities and food resources are available in larger amounts for the entire food chain (Léo & Pires-Vanin 2006, Braga et al. 2018). Furthermore, in addition to the afore mentioned

physicochemical and/or biological disturbances, invertebrates compete with macroalgae for available space, which is a potentially limiting resource in coastal habitats. This competition thus exerts an influence on the community and likely alters its species richness and diversity (Dayton 1971, Tlig-Zouari et al. 2010).

Changes in the macroalgal composition and relative occurrence of different algal families were observed at Prainha before and after an interval of almost 30 years. According to Yoneshigue (1985) and Brasileiro et al. (2009), among Chlorophyta families, Cladophoraceae and Codiaceae were the most diverse families, representing half of the green macroalgae found on the rocky shore. In the present study, Cladophoraceae and Ulvaceae were the most well-represented families. *Bryopsis corymbosa* J. Agardh and *Valonia macrophysa* Kützing, whose abundance characterized Prainha to Yoneshigue (1985), were not recorded in the present study, which might indicate that significant alterations in the community have occurred. Furthermore, *B. corymbosa* was replaced by different species of the same genus at this site (*B. pennata* J. V. Lamouroux, *B. plumosa* (Hudson) C. Agardh, and *B. ramulosa* Montagne), increasing the representation of Bryopsidaceae; this change was considered a substitution of one species by its ecological equivalent(s). Nevertheless, other previously reported species were not recorded during the present study, such as *Derbesia marina* (Lyngbye) Solier, *Derbesia tenuinissima* (Moris & De Notaris) P. Crouan & H. Crouan and *Ulvella lens* P. Crouan & H. Crouan. When comparing the previous studies' lists of macroalgae, 12 species of Chlorophyta were added for Prainha, and six new occurrences were recorded for the larger region of Arraial do Cabo: *Bryopsis ramulosa*, *Cladophora sericea* (Hudson) Kützing, *Cladophoropsis macromeres* W. R. Taylor, *Phyllocladon anastomosans*

(Harvey) Kraft & M. J. Wynne, *Ulva prolifera* O. F. Müller, and *Willeella brachyclados* (Montagne) M. J. Wynne.

The most representative Ochrophyta flora reported by Yoneshigue (1985) belonged to the families Acinetosporaceae, Dictyotaceae, and Scytosiphonaceae, with each family comprising 15 % of the Ochrophyta flora. In the present study, Acinetosporaceae, Dictyotaceae, Sargassaceae, and Scytosiphonaceae were the families with the highest coverage. The latter three families form complex thalli important to microhabitat formation. Some species that form less complex thalli were recorded by Yoneshigue (1985) and Brasileiro et al. (2009), but were not registered in the present study. Eleven species of Ochrophyta were added to the list of Yoneshigue (1985) for Prainha: *Bachelotia antillarum* (Grunow) Gerloff, *Canistrocarpus cervicornis* (Kützing) De Paula & De Clerck, *Chnoospora minima* (Hering) Papenfuss, *Sphacelaria novae-hollandiae* Sonder, *Sphacelaria tribuloides* Meneghini, *Ectocarpus fasciculatus* Harvey, *Dictyopteris delicatula* J. V. Lamouroux, *Feldmannia indica* (Sonder) Womersley & A. Bailey, *Padina antillarum* (Kützing) Piccone, *Rosenvingea orientalis* (J. Agardh) Børgesen, and *Sargassum vulgare* C. Agardh. However, only the last five of these algae represented new occurrences for the region of Arraial do Cabo.

The diversity of Rhodophyta families, unlike the other macroalgal phyla, did not change in the present study, comparing to the lists of Yoneshigue (1985) and Brasileiro et al. (2009). On the other hand, one species registered by Yoneshigue (1985), *Ceramium codii* (H. Richards) Mazoyer, was replaced by its ecological equivalent *C. vagans* P. C. Silva in P. C. Silva, Meñez & Moe. The families Callithamniaceae, Lomentariaceae, Gelidiellaceae, Plocamiaceae, and Rhodymeniaceae were recorded by Yoneshigue (1985) at Prainha but were not

found in the present study. In relation to the review by Brasileiro et al. (2009), the present study added 18 rhodophyte species for Arraial do Cabo region: *Arthrocardia variabilis* (Harvey) Weber-van Bosse, *Amphiroa brasiliiana* Decaisne, *Centroceras gasparrinii* (Meneghini) Kützing, *Ceramium clarionense* Setchell & N. L. Gardner, *Dasya ocellata* (Grateloup) Harvey in Hooker, *Taenioma perpusillum* (J. Agardh) J. Agardh, *Chondria atropurpurea* Harvey, *Herposiphonia delicatula* Hollenberg, *Herposiphonia nuda* Hollenberg, *Herposiphonia parca* Setchell, *Laurencia aldingensis* Saito & Womersley, *Laurencia dendroidea* J. Agardh, *Melanothamnus gorgoniae* (Harvey) Díaz-Tapia & Maggs, *Melanothamnus sphaerocarpus* (Børgesen) Díaz-Tapia & Maggs, *Pterocladia caerulescens* (Kützing) Santelices & Hommersand, *Hypnea cervicornis* (J. Agardh), *Chondracanthus saundersii* C. W. Schneider & C. E. Lane, and *Gracilaria mammillaris* (Montagne) M. A. Howe.

The changes in the macroalgal composition observed may have been related to several factors acting over the intervening years. The RESEX was created in 1997, establishing a protection area for the marine resources in the area and at the same time implementing a plan for the sustainable utilization of these resources by the local community (Silva 2004). Additionally, the Companhia Nacional de Álcalis, which was established in 1960 at Prainha, used to constantly discharge warmer water and sodium carbonate (barrilha) into the seawater. This company ceased its operations in 2006 (FEEMA 1988, Pereira 2010). Thus, the absence of chemical material inputs and the creation of a protection policy for this area could have influenced the increases we observed in the number of species at Prainha and, consequently, the area's biodiversity.

The present study recorded a reduced turnover rate of brown algal species in relation

to Yoneshigue's (1985) data. This group of algae is frequently referred to as one that is particularly ecologically susceptible to anthropogenic influences (Ballesteros et al. 2007, Pinedo et al. 2007, Martins et al. 2012). Therefore, this suggests that the results of this study present a favorable diagnosis to the water quality at Prainha. The presence of macroalgae with structurally complex thalli (such as *Canistrocarpus cervicornis*, *Laurencia* spp., and *Sargassum* spp.) is also an indicator of good water quality at the site because in polluted environments, opportunistic taxa (*i.e.* macroalgae with simple thalli) are favored and intensely compete with invertebrates for space on the rocky shore, which decreases the local diversity (Soares-Gomes et al. 2016).

A non-destructive sampling approach is widely used in ecological studies, including in Brazil, due to its ability to analyze rocky shore communities efficiently and cost-effectively without negatively impacting sensitive species (Masi et al. 2016, Beisiegel et al. 2017, Mantelatto et al. 2018). Despite having some limitations (*e.g.*, it can be difficult to visualize small organisms, and requires one to have previous knowledge of the occurrence of relevant species for the area to perform analyses), the photoquadrat technique is still recognized as an important tool for non-destructive environmental monitoring (Pech et al. 2004, Gomes-Pereira et al. 2016, Perkins et al. 2016).

The utilization of the methodology discussed above may similarly provide assistance to monitoring programs aiming to prevent new (invasive) organisms from settling in an area and control the behavior of exotic species that are already established. For instance, such programs detected and continuously monitoring the Caribbean mollusk *Isognomon bicolor* C. B. Adams, 1845, which has been detected at Prainha since the early 1990s (Domaneschi &

Martins 2002). This invasive species has been expanding its geographical distribution in the southwestern Atlantic since the 1980s, and is present from Rio Grande do Norte, Brazil to Uruguay (Domaneschi & Martins 2002, Breves et al. 2014). The monitoring of exotic species in benthic communities is important due to the threat they pose to native biota, and even entire ecosystems (Amaral & Jablonski 2005).

In conclusion, our analyses of the species coverage, richness, and diversity of a rocky shore community demonstrated that seasonal upwelling-downwelling events may lead to species turnover and changes in diversity between seasons (Weir & Schluter 2007). The studied rocky shore community at Prainha is therefore susceptible to alterations during upwelling periods, but the replacement of any lost species was basically driven by their ecological equivalents (*i.e.* species from the same families or genera), which we consider to be relatively natural and expected intra-annual changes. The comparison of the currently registered flora to those registered almost 30 years ago demonstrated (I.) new occurrences; (II.) high turnover, especially among red algae; and low turnover among brown algae. These temporal differences were smaller than our hypothesis had expected, which indicates good water quality at the site, perhaps related to the changes in coastal waste disposal between the previous and present studies. The present study used approaches that combined reduced sampling impacts (only carried out to obtain some specimens for identification) and photoquadrats, which allowed us to make a detailed description of the intertidal composition and structural characteristics of different zones on the rocky shore with minimal impact, and is recommended as a good methodology for the study of protected marine areas. The information obtained and techniques

developed could assist in the management of the RESEX of Arraial do Cabo, for example to improve the supervision and protection of the benthic community that occupies the rocky shore of Prainha.

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REFERENCES

- ALVES BGR, GÜTH AZ, BÍCEGO MC, GAETA SA & SUMIDA PYG. 2014. Benthic community structure and organic matter variation in response to oceanographic events on the Brazilian SE inner shelf. *Cont Shelf Res* 85: 106-116.
- AMARAL ACZ & JABLONSKI S. 2005. Conservation of marine and coastal biodiversity in Brazil. *Conserv Biol* 19(3): 625-631.
- AREFIN AME. 2016. Proposal of a marine protected area surveillance system against illegal vessels using image sensing and image processing. *Acta Ecol Sin* 38(2): 111-116.
- BALLESTEROS E, TORRAS X, PINEDO S, GARCÍA M, MANGIALAJO L & TORRES M. 2007. A new methodology based on littoral community cartography dominated by macroalgae for the implementation of the European Water Framework Directive. *Mar Pollut Bull* 55(1-6): 172-180.
- BEISIEGEL K, DARR A, GOGINA M & ZETTLER ML. 2017. Benefits and shortcomings of non-destructive benthic imagery for monitoring hard-bottom habitats. *Mar Pollut Bull* 121(1-2): 5-15.
- BEISIEGEL K, DARR A, ZETTLER ML, FRIEDLAND R, GRÄWE U & GOGINA M. 2018. Understanding the spatial distribution of subtidal reef assemblages in the southern Baltic Sea using towed camera platform imagery. *Estuar Coast Shelf Sci* 207: 82-92.

- BORJAA, ELLIOTT M, HENRIKSEN P & MARBÅ N. 2012. Transitional and coastal waters ecological status assessment: advances and challenges resulting from implementing the European Water Framework Directive. *Hydrobiologia* 704(1): 213-229.
- BRAGA HO, AZEITEIRO UM, OLIVEIRA HMF & PARDAL MA. 2018. Conserving Brazilian Sardine: Fisher's attitudes and knowledge in the Marine Extractive Reserve of Arraial do Cabo, Rio de Janeiro State, Brazil. *Fish Res* 204: 402-411.
- BRASILEIRO PS, YONESHIGUE-VALENTIN Y, BAHIA RG, REIS RP & AMADO FILHO G. 2009. Algas Marinhas Bentônicas da Região de Cabo Frio e Arredores: Síntese do Conhecimento. *Rodriguésia* 60(1): 39-66.
- BREVES A, SCARABINO F & LEONI V. 2014. First records of the non-native bivalve *Isognomon bicolor* (C. B. Adams, 1845) rafting to the Uruguayan coast. *Check List* 10(3): 684-686.
- BROWN JH & KODRICK-BROWN A. 1977. Turnover rates in insular biogeography: Effect of immigration on extinction. *Ecology* 58(2): 445-449.
- CARRIER A ET AL. 2015. Trophic connectivity between offshore upwelling and the inshore food web of Banc d'Arguin (Mauritania): New insights from isotopic analysis. *Estuar Coast Shelf Sci* 165: 149-158.
- COELHO-SOUZA SA, LOPÉZ MS, GUIMARÃES JRD, COUTINHO R & CANDELLA RN. 2012. Biophysical interactions in the Cabo Frio upwelling system, Southeastern Brazil. *Braz J Oceanogr* 60(3): 353-365.
- COELHO-SOUZA SA, PEREIRA GC, LOPEZ MS, GUIMARÃES JRD & COUTINHO R. 2017. Seasonal sources of carbon to the Brazilian upwelling system. *Estuar Coast Shelf Sci* 194: 162-171.
- DAWES CJ & MATHIESON A. 2008. The seaweeds of Florida, 1st ed., University Press of Florida, 591 p.
- DAYTON PK. 1971. Competition, disturbance, and community organization: The provision and subsequent utilization of space in a rocky intertidal community. *Ecol Monogr* 41(4): 351-389.
- DOMANESCHI O & MARTINS CM. 2002. *Isognomon bicolor* (C.B. Adams) (Bivalvia, Isognomonidae): primeiro registro para o Brasil, redescrição da espécie e considerações sobre a ocorrência e distribuição de *Isognomon* na costa brasileira. *Rev Bras Zool* 19(2): 611-627.
- DURDEN JM ET AL. 2016. Perspectives in visual imaging for marine biology and ecology: from acquisition to understanding. In: Hughes RN, Hughes DJ, Smith IP & Dale AC (Eds), *Oceanography and Marine Biology: An Annual Review*, Boca Raton, Florida 54: 1-72.
- FAVERI C, FARIAS JN, SCHERNER F, OLIVEIRA EC & HORTA PA. 2010. Temporal changes in the seaweed flora in Southern Brazil and its potential causes. *Panam J Aquat Sci* 5(2): 350-357.
- FEEMA. 1988. Perfil ambiental do Município de Arraial do Cabo, 1^a ed., Rio de Janeiro, 43 p.
- FLOETER SR & SOARES-GOMES A. 1999. Biogeographic and species richness patterns of Gastropoda on the Southwestern Atlantic. *Rev Brasil Biol* 59(4): 567-575.
- FLORA DO BRASIL. 2020. Jardim Botânico do Rio de Janeiro. Disponível em: <http://floradobrasil.jbrj.gov.br/>. Acessado em 27 de dezembro de 2019.
- GOMES-PEREIRA JN ET AL. 2016. Current and future trends in marine image annotation software. *Prog Oceanogr* 149: 106-120.
- GOMÉZ I & HUOVINEN P. 2011. Morpho-functional patterns and zonation of South Chilean seaweeds: the importance of photosynthetic and bio-optical traits. *Mar Ecol Prog Ser* 422: 77-91.
- GUIMARAENS MA & COUTINHO R. 1996. Spatial and temporal variation of benthic marine algae at the Cabo Frio upwelling region, Rio de Janeiro, Brazil. *Aquat Bot* 52: 283-299.
- GUIMARAENS MA & COUTINHO R. 2000. Temporal and spatial variation of *Ulva* spp. and water properties in the Cabo Frio upwelling region of Brazil. *Aquat Bot* 66(2): 101-114.
- GUIRY MD & GUIRY GM. 2019. *AlgaeBase*. World-wide electronic publication, National University of Ireland, Galway. Available at <http://www.algaebase.org>; Accessed on 16 October 2018.
- HOEK C VAN DEN. 1982. The distribution of benthic marine algae in relation to the temperature regulation of their life histories. *Biol J Linnean Soc* 18(2): 81-144.
- HOSEGOOD PJ, NIGHTINGALE PD, REES AP, WIDDICOMBE CE, WOODWARD EMS, CLARK DR & TORRES RJ. 2017. Nutrient pumping by submesoscale circulations in the mauritanian upwelling system. *Prog Oceanogr* 159: 223-236.
- ICMBIO - INSTITUTO CHICO MENDES DE CONSERVACAO DA BIODIVERSIDADE. 2017. Ministério do Meio Ambiente. Disponível em: <http://www.icmbio.gov.br/portal/unidadesdeconservacao/biomas-brasileiros/marinho/unidades-de-conservacao-marinho/2282-resex-marinha-do-arraial-do-cabo>. Acessado em 23 de maio de 2018.
- IVES AR & CARPENTER SR. 2007. Stability and diversity of ecosystems. *Science* 317(5834): 58-62.

- JARA VC, MIYAMOTO JHS, GAMA BAP, MOLIS M, WAHL M & PEREIRA R. 2006. Limited evidence of interactive disturbance and nutrient effects on the diversity of microbenthic assemblages. *Mar Ecol Prog Ser* 308: 37-48.
- JIANG R & WANG Y. 2018. Modeling the ecosystem response to summer coastal upwelling in the northern South China Sea. *Oceanologia* 60(1): 32-51.
- KELAHER BP & CASTILLA JC. 2005. Habitat characteristics influence macrofaunal communities in coralline turf more than mesoscale coastal upwelling on the coast of Northern Chile. *Estuar Coast Shelf Sci* 63(1-2): 155-165.
- KOHLER KE & GILL SM. 2006. Coral Point Count with Excel extensions (CPCe): A Visual Basic program for the determination of coral and substrate coverage using random point count methodology. *Comput Geosci* 32: 1259-1269.
- LAVALEYE MSS, DUINEVELD GCA, BERGHUIS EM, KOK A & WITBAARD R. 2002. A comparison between the megafauna communities on the N.W. Iberian and Celtic continental margins-effects of coastal upwelling? *Prog Oceanogr* 52(2-4): 459-476.
- LÉO FC & PIRES-VANIN AMS. 2006. Benthic megafauna communities under the influence of the South Atlantic Central Water intrusion onto the Brazilian SE shelf: A comparison between an upwelling and a non-upwelling ecosystem. *J Marine Syst* 60(3-4): 268-284.
- LITTLE C, TROWBRIDGE CD, PILLING GM, COTTRELL DM, PLOWMAN CQ, STIRLING P, MORRITT D & WILLIAMS GA. 2018. Long-term fluctuations in epibiotic bryozoan and hydroid abundances in an Irish sea lough. *Estuar. Coast Shelf Sci* 210: 142-152.
- LITTLER DS & LITTLER MM. 2000. Caribbean reef plants: An identification guide to the reef plants of the Caribbean, Bahamas, Florida and Gulf of Mexico, Offshore Graphics, 542 p.
- MAGURRAN AE. 2006. *Medindo a diversidade biológica*. 1ª ed., Editora UFPR, Curitiba, 262 p.
- MANTELATTO MC, SILVA AG, LOUZADA TS, MCFADDEN CS & CREED JC. 2018. Invasion of aquarium origin soft corals on a tropical rocky reef in the southwest Atlantic, Brazil. *Mar Pollut Bull* 130: 84-94.
- MARTINS CDL ET AL. 2012. The impact of coastal urbanization on the structure of phytobenthic communities in southern Brazil. *Mar Pollut Bull* 64(4): 772-778.
- MASI BP, COUTINHO R & ZALMON IR. 2016. Analytical successional tools of fouling communities submitted to different light effects. *Community Ecol* 17(2): 178-187.
- MASI BP, MACEDO IM & ZALMON IR. 2009. Benthic community zonation in a breakwater on the north coast of the state of Rio de Janeiro, Brazil. *Braz Arch Biol Technol* 52(3): 637-646.
- MENCONI MFG ET AL. 2012. Baía de Guanabara: Síntese do conhecimento ambiental. BR Petrobras. Biodiversidade, Rio de Janeiro 2: 479.
- MENDONÇA TCM, MORAES EA & COSTA MAM. 2013. Turismo e pesca nas Reservas Extrativistas Marinhas de Arraial do Cabo (RJ) e da Prainha do Canto Verde (CE): possibilidades e limites de complementaridade. Rio de Janeiro. *CVT* 13(3): 372-390.
- MURRAY SN, AMBROSE RF & DETHIER MN. 2006. *Monitoring rocky shores*. 1st ed., University of California Press, California, 240 p.
- NASCIMENTO A, COELHO - GOMES C, BARBARINO E & LOURENÇO SO. 2014. Temporal variations of the chemical composition of three seaweeds in two tropical coastal environments. *Open J Mar Sci* 4(2): 118-139.
- OLIVEIRA EC & QI Y. 2003. Decadal changes in a polluted bay as seen from its seaweed flora: the case of Santos Bay in Brazil. *Ambio* 32(6): 403-405.
- PECH D, CONDAL AR, BOURGET E & ARDISSON PL. 2004. Abundance estimation of rocky shore invertebrates at small spatial scale by high-resolution digital photography and digital image analysis. *J Exp Mar Biol Ecol* 299(2): 185-199.
- PEREIRA WLCM. 2010. Vagas da modernidade: a Companhia Nacional de Alcalis em Arraial do Cabo (1943-964). *Estud hist (Rio J.)* 23(46): 321-343.
- PERERA-VALDERRAMA S ET AL. 2017. Temporal dynamic of reef benthic communities in two marine protected areas in the Caribbean. *J Sea Res* 128: 15-24.
- PERKINS NR, FOSTER SD, HILL NA & BARRETT NS. 2016. Image subsampling and point scoring approaches for largescale marine benthic monitoring programs. *Estuar Coast Shelf Sci* 176: 36-46.
- PINEDO S, GARCÍA M, SATTÀ MP, TORRES M & BALLESTEROS E. 2007. Rocky-shore communities as indicators of water quality: A case study in the Northwestern Mediterranean. *Mar Pollut Bull* 55(1-6): 126-135.
- RIGGS SR, AMBROSE WG, COOK JW, SNYDER SW & SNYDER SW. 1998. Sediment production on sediment-starved continental margins; the interrelationship between hardbottoms, sedimentological and benthic community processes, and storm dynamics. *J Sediment Res* 68(1): 155-168.

- ROHDE S, SCHUPP PJ, MARKERT A & WEHRMANN A. 2017. Only half of the truth: Managing invasive alien species by rapid assessment. *Ocean Coast Manage* 146: 26-35.
- ROMERO-RAMIREZ A, GRÉMARE A, BERNARD G, PASCAL L, MAIRE O & DUCHÊNE JC. 2016. Development and validation of a video analysis software for marine benthic applications. *J Marine Syst* 162: 4-17.
- ROS JD, ROMERO J, BALLESTEROS E & GILI JM. 1985. Diving in Blue Water. The Benthos. The Western Mediterranean [(R Margalef Ed)], 1st ed., Buckley: Pergamon Press Oxford, p. 233-295.
- SCHERNER F ET AL. 2013. Coastal urbanization leads to remarkable seaweed species loss and community shifts along the SW Atlantic. *Mar Pollut Bull* 76(1-2): 106-115.
- SILVA PP. 2004. From common property to co-management: lessons from Brazil's first maritime extractive reserve. *Mar Policy* 28(5): 419-428.
- SILVA RP, DE PAULA JC & ARAUJO JM. 2015. Estudo metodológico para análise de cobertura do costão rochoso da Prainha, Arraial do Cabo, Rio de Janeiro. *Revista Cecil* 7: 28-30.
- SOARES-GOMES A, DA GAMA BAP, NETO JAB & THOMPSON F. 2016. An environmental overview of Guanabara Bay, Rio de Janeiro. *Reg Stud Mar Sci* 8: 319-330.
- SOLAN M ET AL. 2003. Towards a greater understanding of pattern, scale and process in marine benthic systems: a picture is worth a thousand worms. *J Exp Mar Biol Ecol* 285: 313-338.
- SPALDING MD ET AL. 2007. Marine ecoregions of the world: A bioregionalization of coastal and shelf areas. *BioScience* 57(7): 573-583.
- STENECKRS & DETHIERMN. 1994. A functional group approach to the structure of algal-dominated communities. *Oikos* 69(3): 476-498.
- STEVČIĆ Č, PÉREZ-MIGUEL M, DRAKE P, TOVAR-SÁNCHEZ A & CUESTA JA. 2017. Macroinvertebrate communities on rocky shores: Impact due to human visitors. *Estuar Coast Shelf Sci* 211: 127-136.
- SZÉCHY MTM, VELOSO VG & PAULA EJ. 2001. Brachyura (Decapoda, Crustacea) of phytobenthic communities of the sublittoral region of rocky shores of Rio de Janeiro and São Paulo, Brazil. *Trop Ecol* 42(2): 231-241.
- SZÉCHY MTM, KOUTSOUKOS VS & BARBOZA CAM. 2017. Long-term decline of brown algal assemblages from southern Brazil under the influence of a nuclear power plant. *Ecol Indic* 80: 258-267.
- TANNER JE, MELLIN C, PARROTT L & BRADSHAW CJA. 2015. Fine-scale benthic biodiversity patterns inferred from image processing. *Ecol Complex* 22: 76-85.
- TANO S, EGGERTSEN M, WIKSTRÖM SA, BERKSTRÖM C, BURIYO AS & HALLING C. 2016. Tropical seaweed beds are important habitats for mobile invertebrate epifauna. *Estuar Coast Shelf Sci* 183(Part A): 1-12.
- TAYLOR WR. 1972. Marine algae of the eastern tropical and subtropical coasts of the Americas. 3rd ed., Michigan: The University of Michigan Press, 870 p.
- TEICHBERG M ET AL. 2010. Eutrophication and macroalgal blooms in temperate and tropical coastal waters: nutrient enrichment experiments with *Ulva* spp. *Global Change Biol* 16(9): 2624-2637.
- TLIG-ZOUARI S, RABAQUI L, FGUIRI H, DIAWARA M & HASSINE OKB. 2010. Spatial diversity of rocky midlittoral macroinvertebrates associated with the endangered species *Patella ferruginea* (Mollusca: Gastropoda) of Tunisian coastline. *Estuar Coast Shelf Sci* 87(2): 275-283.
- TRIVEDI S, ALOUFI AA, ANSARI AA & GHOSH SK. 2016. Role of DNA barcoding in marine biodiversity assessment and conservation: An update. *Saudi J Biol Sci* 23: 161-171.
- VINAGRE PA, VERÍSSIMO H, PAIS-COSTA AJ, HAWKINS SJ, BORJA A, MARQUES JC & NETO JM. 2017. Do structural and functional attributes show concordant responses to disturbance? Evidence from rocky shore macroinvertebrate communities. *Ecol Indic* 75: 57-72.
- WALTER RK, ARMENTA KJ, SHEARER B, ROBBINS I & STEINBECK J. 2018. Coastal upwelling seasonality and variability of temperature and chlorophyll in a small coastal embayment. *Cont Shelf Res* 154: 9-18.
- WEIR JT & SCHLUTER D. 2007. The latitudinal gradient in recent speciation and extinction rates of birds and mammals. *Science* 315(5818): 1574-1576.
- WYNNE MJ. 2017. A checklist of benthic marine algae of the tropical and subtropical Western Atlantic: fourth revision. *Nova Hedwigia Beihefte* 145: 1-202.
- YONESHIGUE Y. 1985. Taxonomie et ecologie des algues marines dans la région de Cabo Frio (Rio de Janeiro) Brésil. Thesis, Université d'Aix-Marseille II, 1(1): 461.

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