

Trichoptera (Insecta) in water hyacinth roots: evaluation of the influence of exotic mussel and environmental factors

Trichoptera (Insecta) em raízes de aguapé: avaliação da influência de mexilhão exótico e fatores ambientais

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Abstract: Aim: Information on the influence of mussel macrofouling in invertebrate communities usually have the initial assumption of negative interference. **Methods:** We analyzed this relationship in a community of aquatic invertebrates associated to roots of *Eichhornia crassipes* in 15 shallow marginal lakes in the Pantanal National Park and surroundings. We sampled quadrants of floating vegetation, identified the aggregate fauna and evaluated the density effect of *Limnoperna fortunei*, as well as abiotic factors of the Trichoptera community using ordinances and multivariate regressions. **Results:** We found no significant relationship between the abundance of mussels on the macrophytes and the Trichoptera larvae. However, we observed an interference of oxygen on the structure and density of genera. The composition and abundance of the phytophylous caddisfly community is influenced by the depth and the concentration of oxygen dissolved in the marginal regions of the lakes. **Conclusions:** We suggest that the absence of the effects of the assessed *L. fortunei* in the community is related to the ‘biotic resistance’, in which the phytophylous caddisfly demonstrated adaptation to an environment characterized by hypoxic conditions in the dry season. *L. fortunei* was limited by the depth and reduction of oxygen, presenting lower density in lakes with such characteristics.

Keywords: aquatic invertebrates; bio invaders; golden mussel; wetlands.

Resumo: Objetivo: Informações sobre a influência de macroaglomerados de mexilhão em comunidades de invertebrados geralmente apresentam o pressuposto inicial de interferência negativa. **Métodos:** Analisamos essa relação em uma comunidade de invertebrados aquáticos associados a raízes de *Eichhornia crassipes* em 15 lagoas marginais rasas, no Parque Nacional do Pantanal e áreas adjacentes. Amostramos quadrantes da vegetação flutuante, identificamos a fauna associada e avaliamos o efeito da abundância de *Limnoperna fortunei* e fatores abióticos sobre a comunidade de Trichoptera, utilizando ordenações e regressões multivariadas. **Resultados:** Não encontramos relação significativa da abundância de mexilhão nas macrófitas sob as larvas Trichoptera. Entretanto, observamos a interferência do oxigênio na estrutura e abundância dos gêneros. A composição e abundância da comunidade de tricópteros fitófilos é influenciada pela profundidade e oxigênio dissolvido na região marginal das lagoas. **Conclusões:** Sugerimos que a ausência de efeitos de *L. fortunei* na comunidade avaliada está relacionada com a ‘resistência biótica’ na qual os tricópteros fitófilos demonstraram adaptação ao ambiente caracterizado por condições de hipoxia na estação seca, enquanto que *L. fortunei* foi limitado pela profundidade e redução de oxigênio, apresentando menores densidades em lagoas com essas características.

Palavras-chave: invertebrados aquáticos; bioinvasores; mexilhão dourado; áreas úmidas.

1. Introduction

The Caddisfly is a diverse group of insects with immature aquatic stages. They normally occur in streams with cooler oxygenated water (Wiggins, 2000; Bispo et al., 2006), but can also thrive in floodplains (Pescador et al., 1995) mainly in association with other aquatic plants. The water hyacinth roots (*Eichhornia crassipes* Mart. Solms-Laubach), are a special biotope that harbors a complex aquatic community, comprised of periphyton, zooplankton and invertebrates (“interhizon”) (Ohtaka et al., 2011). This microhabitat includes a diversified fauna of insects (Poi de Neiff & Carignan, 1997; Rocha-Ramírez et al., 2007; Silva & Henry, 2013) and the Trichoptera is a special element related to the trophic chain and ecological process (Wiggins, 2000).

The exotic bivalve *Limnoperna fortunei* (Dunker 1857) (Bivalvia, Mytilidae) also lives in association with macrophytes in the aquatic system of Pantanal. Their larvae are fixed and develop in floating mats of water hyacinths (*E. crassipes*, *Eichhornia azurea* (Swartz) Kunth) (Marçal & Callil, 2008), which act as dispersal agents floating downriver or driven by wind (Pott et al., 2011). The high reproduction rate and clustering behavior gives this bivalve a high capacity for installation and accelerated population growth (Boltovskoy et al., 2006). The presence of factors limiting the invasion and establishment of populations, which are related to limnological and hydrological characteristics of rivers (Oliveira et al., 2011; Boltovskoy, 2015), as well as chemical, physical, and biological control mechanisms (Soares et al., 2009; Boltovskoy, 2015) have been investigated. But after more than 20 years of study, effective solutions to limit the spread of *L. fortunei* larvae have not been found and the species is still likely to invade new Brazilian watersheds.

Since the introduction of *L. fortunei* in South America, the understanding of interactions between particular communities or populations (Sax et al., 2005) has become a challenge to researchers, as structural and functional alterations have been demonstrated (Mansur et al., 2003; Brugnoli et al., 2005). The filtering activity of *L. fortunei* reduces the biomass of phytoplankton, periphyton and zooplankton populations, which causes a cascading effect in the aquatic ecosystem (Brugnoli et al., 2005). Studies predicting the deleterious effect of the golden mussel on the community of aquatic invertebrates, concluded that some groups of typically opportunistic invertebrates (Hirudinea,

Oligochaeta, Nematoda and Gastropoda) were favored due to the increase of structural complexity by macrofouling of *L. fortunei* (Darrigran et al., 1998; Sylvester et al., 2007a; Darrigran & Damborenea, 2009, 2011; Burlakova et al., 2012). The facilitative effects of exotic mussels on benthic species are strongly positively related to its density (Bruno et al., 2005), and dense populations of golden mussels alter the aquatic trophic structure of taxa considered resistant to changes in the environment. However, the effects of this invasive species is still little understood, above all, in habitats, groups of invertebrates/vertebrates and peculiar ecosystems. Biological processes emphasizing interactions between native-exotic species, have been approached for further discussion related to “biotic-resistance” (Simberloff & Von Holle, 1999), once the increased biotic and ecological resistance to ecosystem is a disadvantage to the establishment of invasive species (Espinola & Júlio Junior, 2007). Given this context, the evaluation of the golden mussels effect on the structure of populations and specific communities has great relevance.

This study aimed to evaluate the influenced of two predictors in the structure of Trichoptera phytophylous: i) abundance of *L. fortunei* and ii) limnological variables in floodplain lakes in the Paraguay river in the dry season, in addition to presenting new data about the composition, richness, abundance and density of Trichoptera associated with macrophytes.

2. Material and Methods

This study was carried out in fifteen marginal lakes to the Upper Paraguay River, between the Pantanal National Park and a Private Reserve, RPPN Acurizal, which are protected areas located on southwest part of Mato Grosso state, at the border of Brazil and Bolivia (Figure 1). The Paraguay River begins in the Chapada do Parecis plateau, becoming a tributary of the Paraná River which in turn, empties into the La Plata basin that extends over an area of 3,100,000 km². The climate is warm with an average temperature of 25°C, humidity of about 77%, and average annual rainfall of 1,070 mm, which is more intense from November to March -known regionally by flood periods- and dry season from April to October, considered the draining and dry periods (Oliveira et al., 2006). In the dry season, the floodplain lakes of the Paraguay River may present depths of up to 2 meters (Marchese et al., 2005) and may be connected to the main channel permanently or only during flood

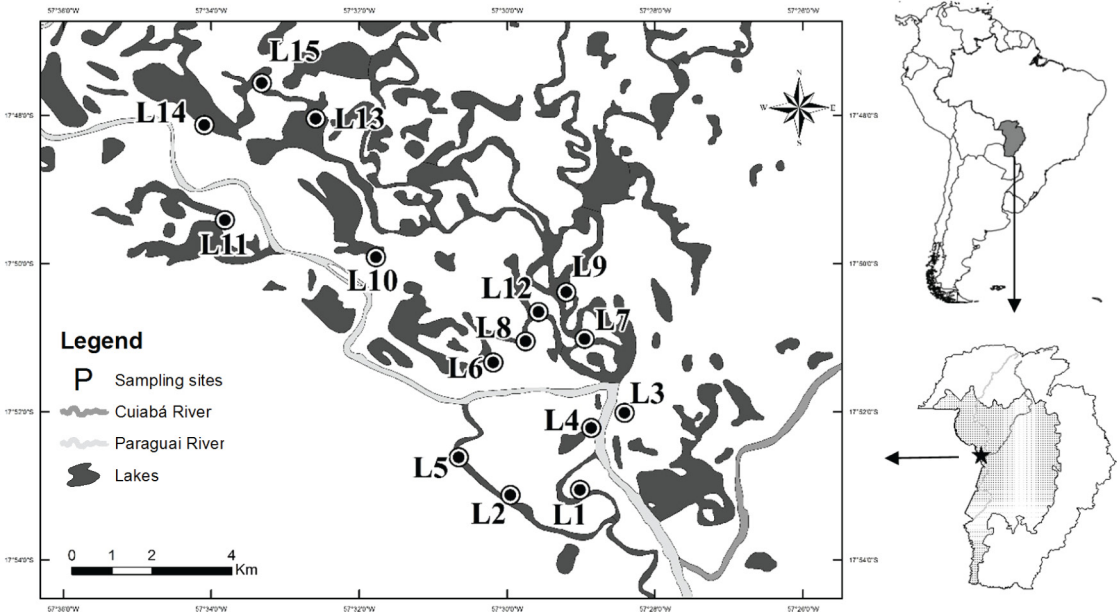


Figure 1. Map of location of the 15 marginal lakes along Paraguay River, on southeast Upper Basin Paraguay (BAP), in southwest part of MT State, Brazil, sampled in September of 2005.

peaks (= oxbows shallow lakes). These environments usually present low concentrations of dissolved oxygen, high organic matter content, and stands of floating macrophytes in different stages of dominance, including *E. crassipes* (Wantzen et al., 2005).

The sampling was conducted in a single campaign in the dry season (in September 2005) in order to avoid interference from disturbances caused by rain in invertebrate animals. The physicochemical variables were determined in each lake (in triplicate) using Secchi disk (depth, transparency) and sensors of each of the parameters (temperature, pH, dissolved oxygen and electrical conductivity). Three samples of the monospecific marginal stands of floating vegetation were sampled from a 25×25 cm square using pruning shears, the total area collected was of 0.1875 m^2 per lake. In the laboratory, the plant's root mass was washed in 2mm, 1mm and 0.25mm mesh, and the resulting part retained in the last sieves was preserved in 70% alcohol. The roots of the plants were sun-dried, placed in an oven at $60 \text{ }^\circ\text{C}$ (about 72 hours), and then weighed to estimate the biomass in grams of dry weight.

The larvae of Trichoptera were screened and identified under a stereoscopic microscope, as described by Domínguez & Fernández (2009). Considering the lake as a sampling unit, the average abundance of Trichoptera was obtained from the replicates. Based on the biomass of roots on the

sampled area, the density of Trichoptera larvae and *L. fortunei* were expressed in number of individuals per 100g of root dry weight ($\text{ind.}100\text{g}^{-1} \text{ DW}$) and individuals per square meter ($\text{ind.}\text{m}^{-2}$).

Before performing statistical tests, the relationship between physical and chemical variables were evaluated through the Spearman Correlation and the variables which presented values above 0.5 ($r_s > 0.5$; $*p < 0.05$; $**p < 0.01$). The pH was correlated to the conductivity (0.657^{**}); the depth was correlated to the dissolved oxygen (0.6385^{**}) and water temperature (0.611^*); the later was also correlated to the conductivity (-0.574^*). Due to covariance with other variables, the temperature was not considered as an environmental predictor. A Principal Component Analysis (PCA) was used to summarize the physical and chemical characteristics of water in the 15 lakes—from the average values—and use the graphical tool *screen plot* to retain only those components that explained most of the variation of the data, as suggested by Gotelli & Ellison (2010) (Figure 2).

The relationship between the average abundance of genera, average abundance of Trichoptera and *L. fortunei* were tested amongst themselves and in relation to the physicochemical variables through a Spearman's correlation ($r_s > 0.5$; $p < 0.05$; $p < 0.01$), after being tested for normality (Shapiro-Wilks; $\alpha = 0.05$) and homocedasticity (Levene; $\alpha = 0.05$), and transformed ($\log(x+1)$). We ordered the average composition and abundance of Trichoptera genera

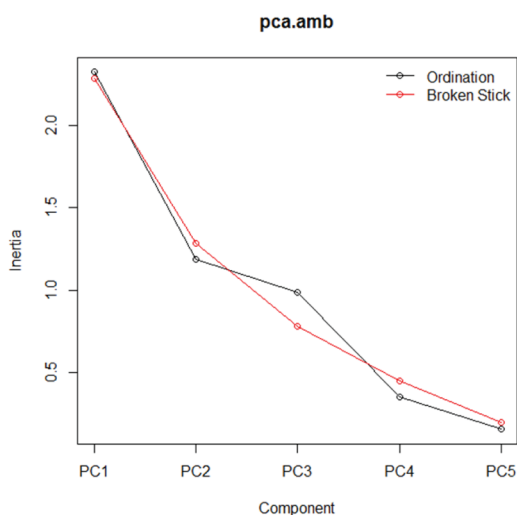


Figure 2. Result of the Principal Components Analysis (PCA) showing the *scree plot* for the physiochemical variables in the 15 marginal lakes along to the Paraguay River, MT.

through the PCoA (Principal Coordinate Analysis) with the Bray-Curtis association index (Legendre & Legendre, 1998). For this purpose, the abundance of each taxon was transformed in $\log(x+1)$, to reduce a heterocedasticity. The relationship between the Trichoptera community (PCoA axes), physiochemical variables (components of the PCA), and abundance of *L. fortunei* was assessed by multivariate inferential analysis by Generalized Linear Model.

The relationship between the Trichoptera community and the physicochemical variables was analyzed using Generalized Linear Models between the axes of PCoA, PCA and the abundance of *L. fortunei*. ($p < 0.05$). We used the models (i) average abundance of Trichoptera (PCoA) in relation to environmental variables (PCA) and abundance of *L. fortunei*, (ii) average abundance of Trichoptera genera in relation to environmental variables (PCA) and abundance of *L. fortunei*, (iii) average abundance of Trichoptera (PCoA) in relation to environmental variables (PCA) and (iv) average abundance of Trichoptera genera (PCoA) in relations to environmental variables (PCA). The data analysis was conducted by using the program R (R Development Core Team, 2013) using functions of the vegan package (Oksanen et al., 2013) and descriptive graphics in the SPSS statistics, version 22.

3. Results

The 15 marginal lakes sampled were shallow (30 to 98.33 cm) with low dissolved oxygen saturation (1.18 to 2.20 mg.l^{-1}) and variable transparency (5.67 to 33.33 cm). The pH values tended to neutrality (6.3 to 7.28); the conductivity showed short variation between 46.27 to 90.4 $\mu\text{S.cm}^{-1}$ and the water temperature ranged from 19.4 to 26.1°C (occasional cold fronts) (Table 1).

The PCA conducted with the environmental variables extracted two axes which explained 70.16% of the total data variation (Figure 3). The first axes of PCA have captured 46.42% of variance of the data and were negatively related with conductivity and positively with dissolved oxygen. The second component explained 23.74% of the data variance, was negatively represented by pH and positively by the transparency of water (Figure 3A and 3B).

A total of 4,761 Trichoptera larvae associated with *E. crassipes* roots were sampled, distributed within 13 taxa belonging to the families Hydroptilidae, Leptoceridae, Hydropsychidae and Polycentropodidae (Table 2). The genus *Neotrichia* sp2 dominated with >50% of relative abundance, followed by *Polycentropus* sp. and *Oecetis* sp. (>10%), comprised the three, most of 80% of the total individuals (Table 2). The genera *Flintiella*, *Smicridea* and the morphospecies Hydroptilinae e Leptoceridae were represented by only 1 individual (0.02%) of the total (Table 2).

The density of Trichoptera varied from 31.45 to 572 ind.100g^{-1} DW (L8 and L11) and 357.33 to 4,448 ind.m^{-2} (L3 and L6; respectively). *L. fortunei* density varied from 3.5 to 548.85 ind.100g^{-1} DW and 21 to 3,616 ind.m^{-2} in (L11 and L12; respectively) (Table 3). Ordering the lakes to river gradient (upstream /downstream) we observed that the majority of deeper lakes (± 60 to 100 cm) are upstream and shallower lakes (± 30 to 60 cm) are downstream. Some genera of Trichoptera showed pronounced density in the lakes (1,099 to 4,448 ind.m^{-2}) with lower mean depth in the littoral region (downstream) (Table 3; Figure 4A). These shallower lakes in marginal areas with water hyacinth, totaled about of 70% of Trichoptera larvae sampled (Figure 4A). Already *L. fortunei* were more abundant mostly in deeper lakes upstream (1,483 to 3,616 ind.m^{-2}) (Table 3).

The PCoA axes explained >70% of the variation of community of Trichoptera in composition and abundance data. For the average abundance of

Table 1. Localization and average values of the physicochemical variables of the water of the 15 marginal lakes along the Paraguay River sampled in September of 2005.

Lakes	Geographical coordinates	pH	DO (mg.L ⁻¹)	E.C. (μS.cm ⁻¹)	W.T. (°C)	Transparency (cm)
L1 Desprezo	S17° 53' 53.4" W57° 29' 00.8"	7.28 ± 0.07	1.56 ± 0.28	81.13 ± 7.01	20.57 ± 1.31	10 ± 0.0
L2 Piuval	S17° 53' 37.6" W57° 29' 57.4"	6.69 ± 0.2	1.35 ± 0.03	90.4 ± 14.18	21 ± 0.62	38.33 ± 12.58
L3 Caracará	S17° 52' 50.9" W57° 28' 24.8"	6.51 ± 0.10	1.26 ± 0.06	65.17 ± 0.86	20.13 ± 0.93	13.33 ± 5.77
L4 Joãozinho	S17° 52' 13.2" W57° 28' 51.8"	6.42 ± 0.14	1.18 ± 0.11	67.73 ± 0.40	20.9 ± 0.96	20.00
L5 Comprida	S17° 52' 36.9" W57° 30' 39.3"	6.79 ± 0.30	1.57 ± 0.07	83.43	19.4 ± 1.30	10 ± 0.0
L6 Cláudia	S17° 51' 20.1" W57° 30' 11.2"	6.36 ± 0.14	1.92 ± 0.22	68.63 ± 4.09	23.67 ± 0.25	18.33 ± 2.89
L7 Três Bocas	S17° 51' 21.1" W57° 28' 57.1"	6.43 ± 0.05	2.20 ± 0.27	46.27 ± 5.25	22.67 ± 0.23	5.67 ± 1.15
L8 Sandrinha	S17° 51' 02.9" W57° 29' 45.0"	6.47 ± 0.08	1.83 ± 0.19	69.98 ± 1.35	24.50 ± 0.44	19.3 ± 4.4
L9 Caracarazinho	S17° 50' 23.2" W57° 29' 12.1"	6.3 ± 0.02	1.73 ± 0.09	55.6 ± 0.72	25.23 ± 0.21	11.67 ± 2.89
L10 Inês	S17° 49' 54.9" W57° 31' 46.4"	6.45 ± 0.07	1.46 ± 0.12	71.50 ± 1.14	25.87 ± 0.91	20 ± 0.0
L11 Acurizal	S17° 49' 24.8" W57° 33' 48.9"	6.51 ± 0.06	1.32 ± 0.16	74.13 ± 1.40	21.53 ± 0.81	20.00
L12 Turco	S17° 48' 05.9" W57° 15' 35.9"	6.39 ± 0.07	1.57 ± 0.67	72.23 ± 0.85	22.27 ± 2.37	15.00
L13 Figueira	S17° 48' 32.5" W57° 32' 35.3"	6.36 ± 0.19	1.55 ± 0.05	55.23 ± 25.59	26.10 ± 0.72	15 ± 0.0
L14 Bigueirinho	S17° 48' 07.7" W57° 34' 05.5"	6.43 ± 0.08	1.51 ± 0.10	71.87 ± 4.53	25.83 ± 0.38	11.67 ± 2.89
L15 Canafisto	S17° 47' 33.8" W57° 33' 18.9"	6.34 ± 0.05	1.57 ± 0.06	68.97 ± 0.67	25.77 ± 0.32	20 ± 5.0

L1 to L15 = Lakes, D.O. = Dissolved oxygen, E.C. = Electrical conductivity, W.T. = Water temperature.

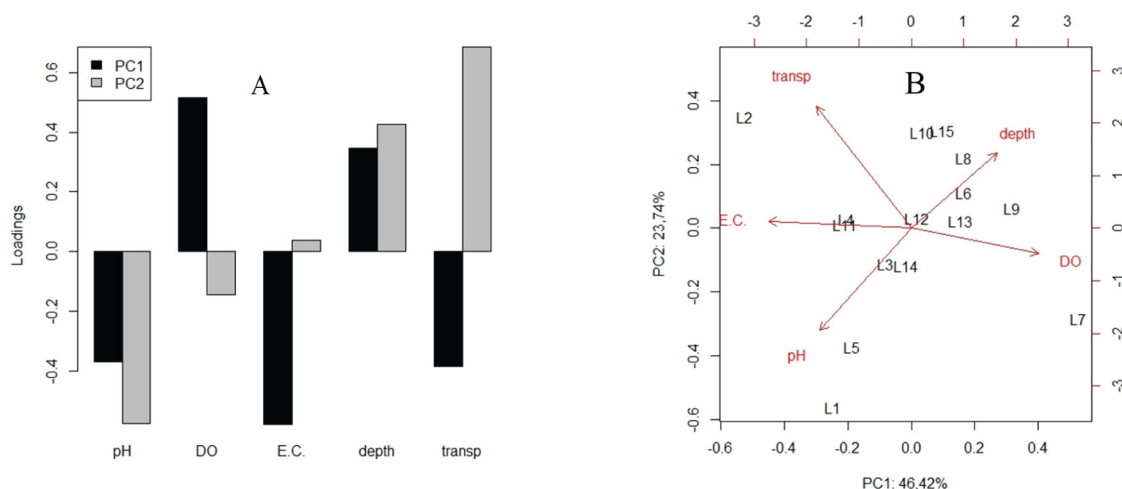


Figure 3. Result of the Principal Components Analysis (PCA) for the physicochemical variables showing the variable, its loadings (A) and the percentage of variance explained through the two first components (B) in the 15 marginal lakes along to the Paraguay River, MT sampled in September of 2005. (B) in the 15 marginal lakes along to the Paraguay River, MT sampled in September of 2005.

Trichoptera in the lakes, both axes of the PCoA represented 99.99% of the data variation.

The relationship between Trichoptera community and the physicochemical variables were analyzed using Generalized Linear Models between the axes of PCoA, PCA and the abundance of *L. fortunei* ($p < 0.05$). We did not find relationship in the model (i), which comprehends average abundance of Trichoptera (PCoA) in relation to the axes of PCA and the abundance of *L. fortunei*. The multivariate multiple regression demonstrated that the ordination of the Trichoptera community in model (ii), in the average abundance of the genera of Trichoptera (PCoA) in relation to the PCA axes and abundance of *L. fortunei* was significant, with the first PCA axes represented mainly by dissolved oxygen and conductivity (Pillai's Trace = 0.717;

$F_{2,10} = 11,444$; $p = 0.003$). The multivariate multiple regression indicated that the community of Trichoptera was not significantly related with *L. fortunei* abundance (Pillai's Trace = 0.050; $F_{2,10} = 0.128$; $p = 0.970$), neither with axes 2 of the PCA (Pillai's Trace = 0.185; $F_{2,10} = 1.028$; $p = 0.396$).

In model (iii), in which we used the average abundance of Trichoptera (PCoA) in relation to the axes of the PCA, we found significant relationship with the PC1 (conductivity and dissolved oxygen) (Pillai's Trace = 0.505; $F_{2,11} = 5,614$; $p = 0.020$) (Figure 4B). In the average abundance of the genera of Trichoptera (PCoA) in relation to the axes of the PCA, and excluding *L. fortunei* in the model (iv), the relationship with the first PCA axes was sharper (Pillai's Trace = 0.7154; $F_{2,11} = 13,8289$; $p = 0.0009$) (Figure 4A).

Table 2. Taxonomic composition, abundance relative (%) and Spearman correlation of genera average abundance, average abundance total of Trichoptera and of *L. fortunei* with physicochemical variables in the 15 marginal lakes along the Paraguay River, MT.

Taxa	%	pH	DO (mg.L ⁻¹)	E.C. (μS.cm ⁻¹)	Depth (cm)
<i>Flintiella</i> sp. Angrisano, 1995	0.02				
<i>Hydroptila</i> sp. Dalman, 1819	0.19				
<i>Neotrichia</i> sp1. Morton, 1905	0.42				
<i>Neotrichia</i> sp2. Morton, 1905	57.55		-0.707**		-0.582*
<i>Oxyethira</i> sp. Eaton, 1873	0.53	-0.569ns		0.676ns	-0.542ns
Hydroptilinae	0.02				
<i>Cyrnellus</i> sp. Banks, 1913	1.93				-0.503ns
<i>Polycentropus</i> sp. Ulmer, 1905	17.66		-0.521*		
<i>Macronemas</i> sp. Pictet, 1836	5.48	0.512**	-0.531ns	0.611ns	
<i>Smicridea</i> sp. McLachlan, 1871	0.02				
<i>Oecetis</i> sp. McLachlan	12.96		-0.639*		-0.643*
<i>Nectopsyche</i> sp. Muller	3.19		-0.629ns		
Leptoceridae	0.02				
Average abundance Trichoptera	100		-0.746**		-0.618*

* $p < 0.05$. ** $p < 0.01$. ns = non significant.

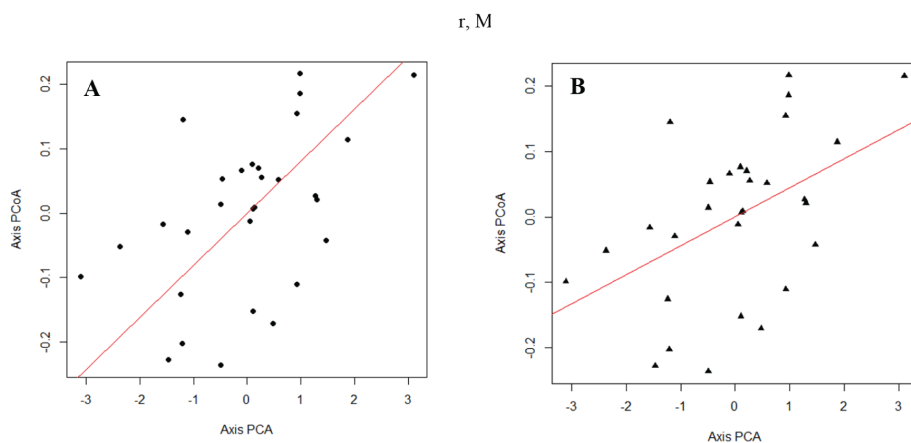


Figure 4. Relationship of the axes PCoA (A) average abundance of Trichoptera genera; (B) average abundance of Trichoptera) associated to *E. crassipes* with PCA axis (physicochemical variables) in the 15 marginal lakes along the Paraguay River, MT, sampled in September 2005.

Table 3. Richness of Trichoptera, abundance (average and standard deviation) and density of Trichoptera and *L. fortunei* in relation to depth and biomass of roots of *E. crassipes* in the 15 marginal lakes of the Paraguay River, MT, sampled in September 2005.

Lakes	Trichoptera			<i>L. fortunei</i>			Depth		Biomass roots (g ⁻¹ DW)	
	Richness	Abundance	ind.100g ⁻¹ DW	ind. m ⁻²	Abundance	ind.100g ⁻¹ DW	ind. m ⁻²	(cm)	Average	Total
L1	9	182 ± 101.95	291.85	2912	24 ± 36.37	38.49	384	58.33 ± 20.21	62.36 ± 26.58	187.08
L2	7	149.33 ± 18.01	337.91	2389.33	51.67 ± 85.16	116.91	826	46.67 ± 5.77	44.19 ± 11.84	132.58
L3	8	278 ± 422.34	504.84	4448	17.67 ± 10.07	32.08	282.67	44.00 ± 1.73	55.07 ± 27.35	165.2
L4	11	200.67 ± 131.34	257.23	3210.67	112 ± 41.61	143.57	1.792	30.00	78.01 ± 46.05	234.03
L5	7	71 ± 57.45	133.06	1136	33.33 ± 30.89	62.47	533.33	41.67 ± 20.21	53.36 ± 19.06	160.08
L6	4	22.33 ± 25.32	33.34	357.33	27.67 ± 34.59	41.3	442.67	68.33 ± 7.64	66.99 ± 24.78	200.96
L7	5	27.33 ± 13.32	37.63	437.33	182.33 ± 149.72	251.04	2,917.33	61.67 ± 10.41	72.63 ± 31.59	217.89
L8	6	24 ± 20.88	31.45	384	158.67 ± 212.17	207.95	2,538.67	96.67 ± 15.28	76.3 ± 8.36	228.9
L9	5	68.67 ± 29.69	116.64	1098.67	13.67 ± 11.02	23.21	218.67	78.33 ± 27.54	58.87 ± 13.47	176.61
L10	7	45.67 ± 26.50	37.35	730.66	167.33 ± 20.65	136.86	2,677.33	98.33 ± 12.58	122.26 ± 97.355	366.79
L11	7	218 ± 49.27	572.03	3488	1.33 ± 2.31	3.5	21.33	40.00	38.11 ± 11.73	114.33
L12	10	136.33 ± 113.32	331.09	2181.33	226 ± 204	548.85	3.616	60.00	41.18 ± 12.86	123.53
L13	6	40 ± 31	64.54	640	92.67 ± 51.07	149.51	1,482.67	56.67 ± 10.41	61.98 ± 11.45	185.94
L14	8	44.67 ± 42.25	68.61	714.67	45.33 ± 28.04	69.63	725.33	46.67 ± 22.55	65.10 ± 12.04	195.31
L15	6	79 ± 4.58	168.62	1264	188 ± 114.79	401.28	3.008	90.00 ± 00.00	46.85 ± 17.05	140.55

The multivariate multiple regressions corroborated with the Spearman correlation, because a negative correlation of dissolved oxygen and depth with the average abundance of Trichoptera (Figure 5A1 and A2) and the average abundance of the *Neotrichia* p2. (Figure 5B1 and B2) and *Oecetis*

(Figure 5C1 and C2) occurred. The conductivity, in spite of correlating negatively with the PC1, did not demonstrate a direct influence on the genera of Trichoptera, but affected indirectly the average abundance of *Macronema* sp., by means of correlation with the pH (Figure 5D3).

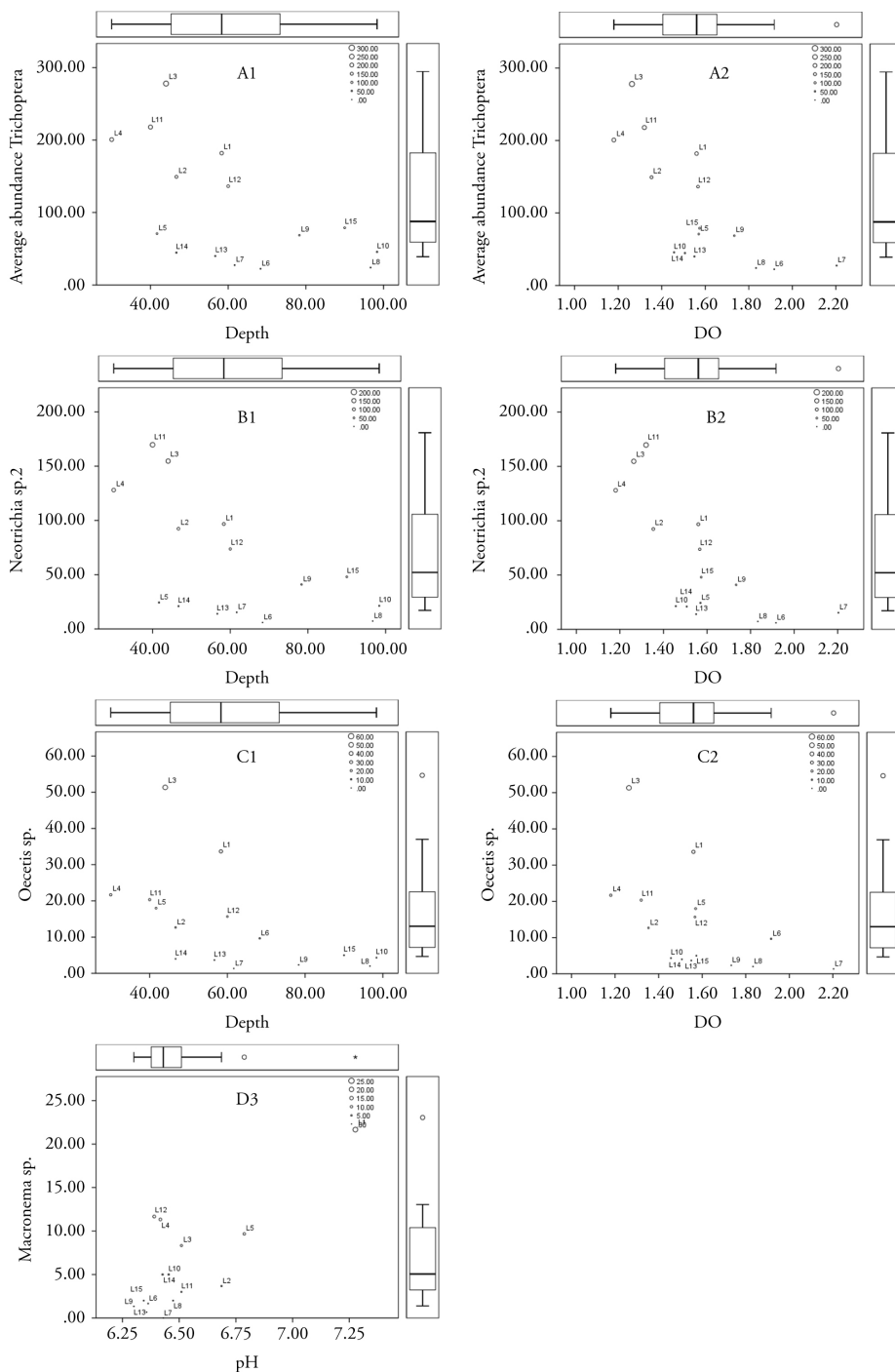


Figure 5. Distribution of the average abundance of Trichoptera (A) and average abundance of genera *Neotrichia* (B), *Oecetis* (C) and *Macronema* (D) associated to *E. crassipes* in the gradient of depth (1), dissolved oxygen (2) and pH (3) of the in the 15 marginal lakes along the Paraguay River, MT, sampled in September 2005. (1), dissolved oxygen (2) and pH (3) of the in the 15 marginal lakes along the Paraguay River, MT, sampled in September 2005.

The depth gradient and the concentration of dissolved oxygen in the lakes waters ordered the composition and abundance of genera affecting the Trichoptera fauna in the dry season (Table 2; Figure 5A1 and A2). Of the 13 taxa collected, 11 occurred at lower oxygen concentrations, being that *Neotrichia* sp2., *Polycentropus*, *Cyrnellus*, *Macronema*, *Oecetis* and *Nectopsyche* occurred in the whole gradient of the concentration and depth (Table 2), being the more abundant genera (*Neotrichia* sp2. and *Oecetis* sp.) negatively related to both variables (Figure 5B and 5C) and the others with one of them (Table 2). *Smicridea* sp., Leptoceridae and *Flintiella* sp. were only found within this gradient of DO (1.1 to 1.5 mg.l⁻¹) and depth (25 to 50 cm). Hydroptilinae e *Hydroptila* occurred only in intermediary DO concentration (1.5 to 1.9 mg.l⁻¹) and depth above 50cm. The occurrence of species in the extreme of the gradient concentration (1.9-2.3 mg.l⁻¹) coincided with those that occur in the entire gradient (Table 2).

4. Discussion

The abundance of exotic golden mussels does not affect the structure of caddisfly community associated to the roots of *E. crassipes* in the lakes of Paraguay River, but we found significant relationship between caddisfly and the environment. The absence of effects in the density of Trichoptera by golden mussels can be expressed by the inherent characteristics of the natural habitat and species (Strayer et al., 2011; Marçal & Callil, 2012; Uhde et al., 2012; Boltovskoy, 2015). Trichoptera presents varying degrees of tolerance to natural and anthropogenic disturbances, making them excellent indicators of ecosystem health (Myers et al., 2011), but the taxa sampled are naturally favored in shallow lakes. Considering the environmental data, our results demonstrate that the community of Trichoptera living in association with *E. crassipes* has a successful adaptability to environmental conditions in the water of the Pantanal floodplain. Dissolved oxygen is a structuring factor for this insect community, determining the abundance and occurrence of the genera in the lakes. It is associated with an unknown factor in this study related to the quality of microhabitat for the larvae of Trichoptera. The most abundant genera, such as *Neotrichia* sp2. and *Oecetis* sp. were found throughout the dissolved oxygen gradient, but they had their abundance decreased with the increase of this gradient. The genera with slighter occurrence, and therefore

less representative in abundance, were restricted to a range of gradient of DO and depth, such as *Smicridea* sp., Leptoceridae and *Flintiella*, which occurred only in the lowest oxygen concentrations.

The most marginal shallow lakes of the Pantanal (Wantzen et al., 2005; Oliveira et al., 2006) are characterized by low values of oxygen in the water, and this is strongly influenced by the depth, as in deep lakes little oxygen is consumed faster (Cremona et al., 2008). In the Paraguay River, during the dry season, the hydrological connectivity is reduced and the lakes become more susceptible to the influence of local processes such as rain, wind and decomposition of macrophytes (Marchese et al., 2005; Thomaz et al., 2007), but they do not present a clear pattern of deoxygenating due to the flow speed variation in different places (Poi de Neiff, 2003). All of these factors may indirectly influence the micro distribution of invertebrate, including Trichoptera larvae in lakes of seasonally floodplains. Thus, we observed a depth gradient pattern downstream-upstream of the lakes for Trichoptera and *L. fortunei*. The lakes L1 until to L5 (downstream) were the shallowest with highest density and richness of Trichoptera and less of *L. fortunei* that prefer deeper lakes, occurring at higher density at upstream lakes. Apparently, the depth and the oxygen dissolved in the water interfere in different ways in caddis and mussels density in macrophytes.

We observed that in marginal regions of shallower lakes covered by aquatic plants have lower concentrations of oxygen in the water. The oxygen deficit is usually a stress factor for aquatic invertebrates, causing a decline in total abundance and an increase in the number of species adapted for survival under hypoxic conditions (Irmeler, 1975; Poi de Neiff & Carignan, 1997). However, the depth and oxygen are covariant factors (Cremona et al., 2008), that contribute to the advancement of seasonal succession process in water bodies of the Pantanal. Thus, the succession process and progress of senescence of aquatic plants is dependent on the quantity and quality of water in the lakes, as observed in lakes connected to Paranapanema River (Stripari & Henry, 2002; Silva & Henry, 2013). During the dry season, dissolved oxygen in water presented negative correlations between decomposition rates of macrophytes (Stripari & Henry, 2002), which explains higher density of Trichoptera in lakes with less depth and OD. Nevertheless, aquatic plants that form floating mats, such as water hyacinth (*E. crassipes*), can

decrease macroinvertebrate abundance by blocking light transmission and decreasing photosynthesis by phytoplankton and others plants, leading to anoxic condition in shore lakes (Masifwa et al., 2001).

We believe that low concentrations of oxygen in the marginal regions of the lakes leads to both decomposition and low photosynthesis, causing periodic fluctuations in abundances of adapted communities, such as Trichoptera. The presence of regular hydrological phases interferes in the abundances and in the life-cycle of the populations, so that the life-history adaptations typically involve the synchronization of life-cycle events, such as reproduction and growth, in relation to the long-term average dynamics of the flow regime (Lytle & Poff, 2004). This implies, higher productivity with extraordinary concentration of individuals, greater resilience in process cyclic succession, and higher functional redundancy (Walker, 1992).

The depth gradient of marginal lakes with low water level causing oxygen depletion in dry season affects the quality of the habitat (*E. crassipes*), and consequently the abundance of Trichoptera and *L. fortunei*. Similar condition of greater density of Trichoptera, as *Oecetis inconspicua* (Walker, 1852) (Leptoceridae), in vegetated lakes with shallow marginal region (Strayer & Smith, 2001), is explained by the high primary production of periphyton. The invertebrate density presents an inverse relationship with the remaining biomass of macrophyte during the decomposition (Bruquetas de Zozaya & Neiff, 1991), because during decomposition, bacteria and fungus and later of periphytic algae produced an increase in nitrogen and protein, and the detritus became attractive as food source of the invertebrates (Masifwa et al., 2001; Weatherhead & James, 2001; Stripari & Henry, 2002). Thereby, the stage of succession of *E. crassipes* apparently interferes in the richness and abundance of genera and trophic structure of the community, that was characterized by occurrence of micro-caddisfly, favored *E. crassipes* by the food provided in their roots (Stripari & Henry, 2002; Mormul et al., 2006; Poi de Neiff & Neiff, 2006; Silva & Henry, 2013).

The effect of an invasive species is more assertive in species directly connected via a chain of consumption than those that have indirect relations but, our results demonstrate greater abundance of collectors; detritus eaters and periphyton scrapers as *Neotrichia* and moderate amount of filterers, like *Cyrrnellus*, *Polycentropus* and *Macronema* sp.

In the floodplain, river and lake areas, *Neotrichia* comprise the fauna of collectors in the first stages of decomposition of aquatic macrophytes (Poi de Neiff & Bruquetas de Zozaya, 1991). Thus, it is clear that the succession in these plants for the invertebrate species has a relationship with time of decomposition expressed by the loss of biomass, and probably the individuals of this taxon are later colonizer in terms of accumulation of debris on the roots. The loss of planktonic food via mussel's filtration activity is compensated by increased production of aquatic macrophytes, periphyton and detritus amortizing any effect on food webs of the littoral zone (Strayer et al., 2011). Thus, the influence on filter feeders as Polycentropodidae and Hydropsychidae would not be observed depending on the amount of resources available in macrophytes.

The consequences of the presence of invasive species to the invertebrate fauna can be variable, depending of the community's characteristics, physicochemical conditions, structure of habitats, and the density potentiality of invasive species (Stachowicz & Tilman, 2005; Bruno et al., 2005). In general, in shallower lakes the density of *L. fortunei* was lower, and despite not being the hypothesis testing, the density of the species showed a relationship with the gradient downstream-upstream of depth of the lakes, with a trend of higher mussel density in deep lakes (upstream) and less density at downstream shallow lakes, connected to oxygen suppression at shallower depths. The reduction in population size of *L. fortunei* in Pantanal was cited by Oliveira et al. (2010a), resulting from the oxygen depletion events caused by increasing the decomposition of organic matter, during the rising water phase; a phenomenon locally known as "dequada". The seasonal depletion of oxygen in the water in the Paraguay River and other rivers of the Pantanal, during some months of the year, combined with other factors (low calcium, low pH, low chlorophyll a concentrations, high suspended solids and water velocity) explain the low density of *L. fortunei* in some areas of the Pantanal (Oliveira et al., 2010b, 2011). However, the dissolved oxygen has been found as a factor that regulates the reproductive activity and survival of the golden mussel, impeding its uncontrolled proliferation (Oliveira et al., 2010b, 2011; Eilers et al., 2011; Bonel et al., 2013), limiting the abundance increase, but not interfering in the establishment of the species (Oliveira et al., 2010b). Perepelizin & Boltovskoy (2015) consider that oxygen deprivation may be a viable alternative for the control of mussel fouling in

industrial installations, because they observed total mortality at dissolved oxygen levels < 0.16 mg/L after 10-12 days (at 27°C), in addition to higher mortality of smaller mussels (7mm) when oxygen deprivation occurs in lower temperature (at 20°C). The data presented in our study, show that during the dry season, the lowest percentage of oxygen in shallow lakes, also acts as controller for population *L. fortunei*, associated with other covariates factors, as water temperature, typically high in the Paraguay River (occasional cold fronts 19.4 to 26.1°C). Thus, population growth of *L. fortunei* no Pantanal is associated with the hydroperiod, which makes habitat conditions and environmental variables unfavorable at different times of seasonal cycle (dry season; rising season see Oliveira et al., 2010a, b, 2011).

The gregarious habit and the scarcity of stable substrates on the bottom of the rivers and lakes also contribute to the control of the *L. fortunei* population. The instability of *E. crassipes* as a substrate and its brief life cycle makes it a refuge, a place with food resources and suitable setting for epibenthic species with seasonal life cycles. As this species of Mytilidae have indirect development, while stage veliger is easily dispersed by water column and after become pediveliger and post-larvae, it needs to fix to complete their development, thereby increasing the filtration, and thus, reach sexual maturity (Santos et al., 2005). Among other strategies the golden mussels assume, the most viable in the Pantanal, is the association of young roots epiphytic macrophytes, initially attracted by periphyton, source of oxygen and food. Because it is a sedimentation basin (Junk et al., 2006), the availability of hard substrates to the successful installation of the golden mussel is limited in the Pantanal. Thus, low densities of *L. fortunei* in the Pantanal, when compared with Argentina and southern Brazil showing amounts of about 145,000 ind.m⁻² (Boltovskoy et al., 2006; Mansur et al., 2003), also show the limitations of the environment – ‘ecosystem resistance’ for the development of the population of the invader.

Trichoptera larvae also have more gregarious habits (Wiggins et al., 1980) but, these immature insects usually have morphological and physiological adaptations (Holomuzki & Biggs, 2000), which are determinants for the increased tolerance or escape ability in environments susceptible to drought (Wiggins et al., 1980). The immature Leptoceridae and Hydroptilidae family have morphed-ecological adaptations that increase their respiratory efficiency and enable them to survive in lentic environments

with low concentrations of oxygen (Wiggins, 2000). In this context, the negative correlation of the most abundant genera (*Neotrichia* sp2, *Oecetis* sp.) with dissolved oxygen and depth may be related to the fact that the larvae of these genera of caddisflies, as well as some immature Lepidoptera and Diptera-Chironomidae, use body ventilation to pump water through their tubes (Merritt & Cummins, 1996). Whence, most caddis builders’ tubes are late colonizers in the succession process when compared to other invertebrates (Albertson et al., 2011).

The caddisfly associated with macrophytes could be negatively affected by dense populations of golden mussel settled in the Pantanal floodplains. In this study, however, the populational density this exotic mollusk was low. The magnitude of the impacts of the *L. fortunei* in the biota is related to the population size of invasive species and the invaded system susceptibility (Hicks, 2004). *L. fortunei* like *D. polymorpha* (Karatayev et al., 2010), has competitive advantages such as high reproductive capacity (Darrigran et al., 2009; Callil et al., 2012) and dispersion strategy (Oliveira et al., 2011) classifying it as a successful invasive species. However, after its spread in the habitat, the harmful effects to the environment are dependent on the level of stability and resilience of the ecosystem where the species was introduced (Stachowicz & Tilman, 2005; Strayer et al., 2011), and only long-term studies will show which groups will be affected by *L. fortunei* in the Pantanal aquatic habitats.

Caddisfly associated with water hyacinth roots showed adaptation to the environment characterized by hypoxic conditions in the dry season, while *L. fortunei* was limited by the depth and oxygen reduction, with lower densities in ponds with these characteristics. So, dissolved oxygen is a structuring factor for this insect community, determining the abundance and occurrence of genera in the lakes. Opportunistic species as *Neotrichia* sp. may be favored by an increase of detritus and recourses of animal and plant origin (periphyton, testate amoebas, bio-deposition) accumulated in the roots. It is important to distinguish the effects of grouping valves phytophilous fauna those observed for the benthic fauna. The macrofouling serves as an additional substrate available to a variety of organisms and its tridimensional structure increases the complexity of the habitat (Ricciardi et al., 1997; Boltovskoy et al., 2006, 2009; Sylvester et al., 2007b). The input of organic material from filtration and bio-deposition increase the density

and biomass of trophic groups (Burlakova et al., 2012) and the empty shells accumulate to form reefs that provide additional hard structures that act as refuges (Sardiña et al., 2008). So, *L. fortunei* as *D. polymorpha* act as a very effective ecosystem engineer, altering both ecosystem structure and function (Sylvester et al. 2007a; Darrigran & Damborenea, 2011). Macrofouling of mussel shells may not signify much to the structural complexity in macrophytes, but this needs to be properly tested. A study on the differences between substrates and colonized communities (epibenthic fauna associated with macrophytes and benthic fauna) by *L. fortunei* bring new elements of effectiveness of this species as ecosystem engineer.

We believe that the action of *L. fortunei* in invertebrate communities depends on the behavior and the niches occupied by components of the community – ‘biotic resistance’ as well as the limitations of the environment – ‘ecosystem resistance’ for the development of the population of the invader. We know that the equilibrium of a natural system depends of the structure of the environment, which reflects how healthy it is, the functioning, determined not only resistance but also the range of resilience. The action of an invasive species with a high power of interference in these attributes, integrity and stability, should receive continuous attention. Whence, the maintenance of the natural flooding cycle in the Pantanal is essential for the renewal of macrophytes and the fauna that thrive in this community. We also highlight the need to carry out continuous monitoring of the invasion of *L. fortunei* on the Upper Paraguay basin, as well as studies with sampling designs to identify the specific effects of this invader on different trophic niches.

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