



Environmental Microbiology

Biomethane production from vinasse in upflow anaerobic sludge blanket reactors inoculated with granular sludge



Valciney Gomes de Barros^a, Rose Maria Duda^{b,c}, Roberto Alves de Oliveira^{a,b,*}

^a Post-Graduate Program in Agricultural and Livestock Microbiology, Faculty of Agricultural and Veterinary Sciences, Univ Estadual Paulista, Campus of Jaboticabal, Jaboticabal, SP, Brazil

^b Laboratory of Environmental Sanitation, Department of Rural Engineering, Faculty of Agricultural and Veterinary Sciences, Univ Estadual Paulista, Campus of Jaboticabal, Jaboticabal, SP, Brazil

^c Faculty of Technology Jaboticabal, Jaboticabal, SP, Brazil

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ABSTRACT

The main objective of this study was to evaluate the anaerobic conversion of vinasse into biomethane with gradual increase in organic loading rate (OLR) in two upflow anaerobic sludge blanket (UASB) reactors, R1 and R2, with volumes of 40.5 and 21.5 L in the mesophilic temperature range. The UASB reactors were operated for 230 days with a hydraulic detention time (HDT) of 2.8 d (R1) and 2.8–1.8 d (R2). The OLR values applied in the reactors were 0.2–7.5 g_{total COD} (L d)⁻¹ in R1 and 0.2–11.5 g_{total COD} (L d)⁻¹ in R2. The average total chemical oxygen demand (total COD) removal efficiencies ranged from 49% to 82% and the average conversion efficiencies of the removed total COD into methane were 48–58% in R1 and 39–65% in R2. The effluent recirculation was used for an OLR above 6 g_{total COD} (L d)⁻¹ in R1 and 8 g_{total COD} (L d)⁻¹ in R2 and was able to maintain the pH of the influent in R1 and R2 in the range from 6.5 to 6.8. However, this caused a decrease for 53–39% in the conversion efficiency of the removed total COD into methane in R2 because of the increase in the recalcitrant COD in the influent. The largest methane yield values were 0.181 and 0.185 (L) CH₄ (g_{total COD removed})⁻¹ in R1 and R2, respectively. These values were attained after 140 days of operation with an OLR of 5.0–7.5 g_{total COD} (L d)⁻¹ and total COD removal efficiencies around 70 and 80%.

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Introduction

Ethanol is the world's most widely used biofuel.¹ The global production of ethanol in 2011 was 86.1 billion liters, of which the US and Brazil contributed 62.7% and 24.4%, respectively.²

Vinasse is the final residue obtained during ethanol production by fermentation of sugarcane.³ For one liter of ethanol produced from sugarcane is estimated to output 8–18 L of vinasse.⁴ The vinasse leaves the distillation column with a temperature of about 90 °C and pH between 3 and 4. It constitutes 94–97% water, Mg²⁺, Ca²⁺, K⁺, melanoidins, and

* Corresponding author.

E-mail: raoder@fcav.unesp.br (R.A. de Oliveira).

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residual amounts of sugar, alcohol, and volatile components such as chloroform, pentachlorophenol, phenol, and methylene chloride, and the amount of these substances depends on the feedstock and the process of ethanol production.^{4,5} The España-Gamboa et al.,⁶ reported the presence of antibacterial components and heavy metals in vinasse. Approximately 75% of suspended solids present in vinasse are organic and biodegradable,⁷ which provides a high chemical and biochemical oxygen demand (COD and BOD) of up to 100 gL⁻¹ and 60 gL⁻¹, respectively.^{3,8}

In Brazil, vinasse is mainly used in fertigation of sugarcane in adjacent areas of ethanol production industries because of its high organic matter and nutrient contents.⁶ However, studies suggest that the application of vinasse indiscriminately in soil could contaminate surface water and groundwater.⁹

In addition, the anaerobic digestion of vinasse may be utilized for stabilization of organic material and methane production, which can be used to produce energy required for drying yeast in distillery.⁸ During the anaerobic digestion of vinasse, most of the organic matter is removed, leaving the recalcitrant organic compounds and most of the nutrients in the effluent.⁵

The higher temperature of vinasse promotes thermophilic anaerobic digestion. However, in some industries, the systems that utilize the thermal energy of vinasse are currently being installed. This makes the topic of mesophilic anaerobic digestion interesting,¹⁰ and studies have shown its advantage over thermophilic anaerobic digestion in terms of imparting greater stability.⁷

Currently, the upflow anaerobic sludge blanket (UASB) reactor is the most widely used reactor for the treatment of vinasse obtained from the ethanol industry.¹¹ Low sludge production and the conversion of approximately 50% of the total chemical oxygen demand (_{total}COD) of vinasse to biogas have couples of advantages of using the UASB reactor.⁶ The design of the reactor is quite simple and does not

require sophisticated equipment.¹² This enables the use of UASB technology for treating industrial waste such as vinasse.

In the UASB reactors, the microorganisms are mainly grouped into granules and flocs formed by self-aggregation of bacteria and archaea. These formations greatly depend on the upflow and composition of wastewater. The granules are compact clusters that possess high specific methanogenic activity and sedimentation. They accumulate in large quantities in the fermentation chamber of UASB reactor.¹²

For efficient methane production from vinasse, strategies need to be developed for startup and maintenance of the anaerobic microbiota. High concentrations of phenolic compounds, such as melanoidins, present in the vinasse,⁹ heavy metals, and antibacterials used in the treatment of yeast may decrease and even inhibit the microbial activity.¹³

The startup time of the anaerobic reactor without adapted sludge inoculum can be up to 40% higher than that of a reactor with the use of adapted sludge.⁶

Within this context, this study aims to assess the startup and stabilization of the anaerobic conversion of vinasse to methane with gradual increase of organic loading rate in UASB reactors.

Materials and methods

Reactor configuration

The experimental unit consisted of two bench-scale UASB reactors (R1 and R2) of capacity 40.5 L and 21.5 L, respectively (Figure 1). The reactors R1 and R2 have five and four sludge collection points, respectively, distributed along the height of the reactors (97.8 and 108.5 cm, respectively, Figure 1). In R1, the sludge collection points P1, P2, P3, P4, and P5 are located at 5, 23.5, 37.4, 51.9, and 65.7 cm from the base of the reactor. In

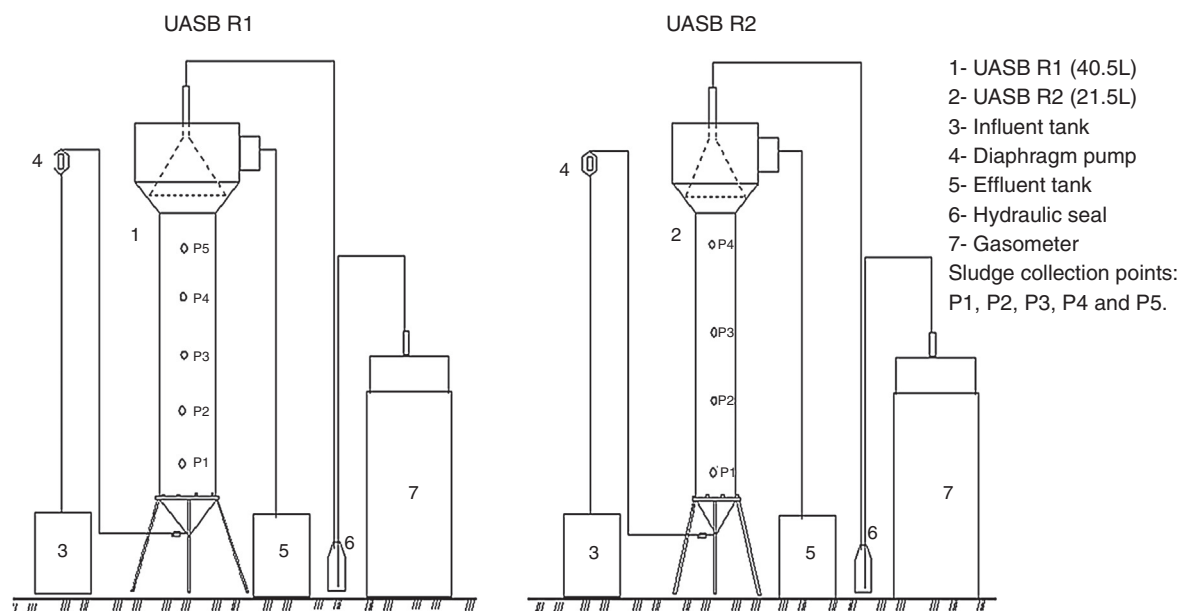


Figure 1 – Schematic representation of the treatment system with upflow anaerobic sludge blanket (UASB) reactors (R1 and R2).

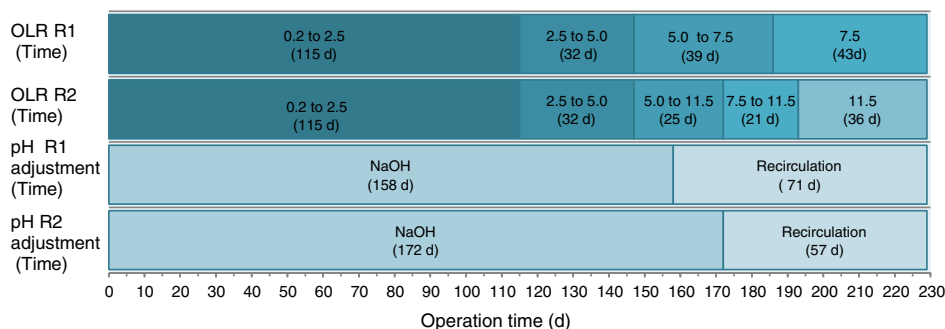


Figure 2 – Schematic diagram with the operating conditions of organic loading rate (OLR in $\text{g}_{\text{total COD}} (\text{L d})^{-1}$) and pH adjustment in UASB reactors (R1 and R2) for vinasse treatment.

R2, the sludge collection points P1, P2, P3, and P4 are located at 5, 25, 40.9, and 71.8 cm from the base of the reactor.

Inoculum and influent characteristics

For the UASB reactor startup, the inoculum was granulated sludge from the UASB reactor used to treat swine wastewater. The inoculum sludge had total (TS) and volatile (VS) solid concentrations of 45.6 g L^{-1} and 30.4 g L^{-1} , respectively. The volume of the sludge used was sufficient to occupy 30% of the volume of each reactor.

The influent used to feed the UASB reactors was obtained from *in natura* vinasse from a sugarcane plant in Ribeirão Preto, SP. The vinasse was collected weekly, from April to December 2012, after distillation in the wine columns, cooled down, and kept chilled.

The concentrations of total COD , TS, VS, Kjeldahl N, total P , K, Ca, Mg, Na, Fe, Mn, Zn, and Cu were determined for characterization of vinasse.¹⁴ The concentrations of total COD , ST, and SV were 45,000; 41,300 and 31,800 mg L^{-1} , respectively. The Kjeldahl N and total P, K, Ca, Mg, Na, Fe, Mn, Zn, and Cu concentration were 470, 170, 88, 3.2, 1.4, 20.4, 24.6, 2.44, 0.78, and 0.21 mg L^{-1} , respectively.

A total COD of about 45,000 mg L^{-1} was necessary to dilute the vinasse and gradually increase the total COD of the influent, consequently increasing the OLR in the UASB reactors. Initially, vinasse was diluted with water, and subsequently, the effluent from the UASB reactors was recirculated. Recirculation of the effluent allows limited use of dilution water and alkalizing as well as reuse of nutrients remaining in the effluent.

Operating conditions of the reactors

The R1 was operated with a hydraulic detention time (HDT) of 2.8 d, while R2 was first operated with an HDT of 2.8 d for 219 days and then decreased to 1.8 d. The HDT was decreased to obtain a gradual increase in OLR. The OLR was calculated by dividing the total COD of the influent by HDT.

The upflow velocity in the reactors R1 and R2 were similar, 0.019 m h^{-1} and 0.018 m h^{-1} , respectively. The surface loading rates in the settlers of the UASB reactors were 0.011 m h^{-1} (R1) and 0.014 m h^{-1} (R2). With the decrease in HDT for R2,

the upflow velocity and the surface loading rate in the settler of the UASB reactor increased to 0.028 m h^{-1} and 0.022 m h^{-1} , respectively.

Although the reactors had different dimensions, they were assumed to be identical, and the same operational conditions (HDT, OLR, alkalizing, and recirculation) were applied. This consideration allowed us to assess the effects of the variables by maintaining one reactor in a normal stable condition while the other was being subjected to new operational conditions.

The OLR was increased from 0.2 to $7.5 \text{ g}_{\text{total COD}} (\text{L d})^{-1}$ in R1 and from 0.2 to $11.5 \text{ g}_{\text{total COD}} (\text{L d})^{-1}$ in R2 (Figure 2). The OLR was gradually increased to adapt the inoculum sludge and to obtain stability with higher OLR.

The mean pH value of the *in natura* vinasse was 4.5. Therefore, it was necessary to correct the pH of the influent to approximately 7.0. Until 158 and 172 days of operation of R1 and R2, respectively, pH of the influent was corrected by adding a solution of 12 M NaOH. After this period, use of NaOH was discontinued and the effluent was recirculated, utilizing the alkalinity generated in the reactors for pH correction (Figure 2). Effluent recirculation in R1 and R2 was started with an OLR of 6 and $8 \text{ g}_{\text{total COD}} (\text{L d})^{-1}$, respectively.

The total COD , Kjeldahl nitrogen (KN), and total P found in *in natura* vinasse were 45,000; 470 mg L^{-1} and 62 mg L^{-1} , respectively. These values did not correspond to the recommended minimum proportion of $\text{COD:N:P} = 350:5:1$ for proper microbial growth.¹⁵ For supplemental phosphorus and nitrogen, potassium phosphate monobasic (KH_2PO_4) and urea ($\text{CH}_4\text{N}_2\text{O}$) were added to vinasse.¹⁶

Analytical methods

Table 1 shows the physical examinations and organic and inorganic constituents determination methods adopted for the samples of influents, effluents, sludge, and biogas of the reactors. The frequency and bibliographic references of the methodologies used are also listed in the table. The air temperature near R1 and R2 was measured daily with a thermometer, and the mean values ranged from 20°C to 30°C . Therefore, the reactors were operated predominantly in the mesophilic temperature range. The daily volume of methane produced in the

Table 1 – Determination and examination, frequency and bibliographic reference of the methodologies used for influent, effluent, sludge, and biogas.

Examination and determination Influent and effluent	Frequency	Bibliographic Reference
pH	Twice a week	(Method: 4500 – B) ¹⁴
Total (totalCOD), dissolved (dissCOD), and suspended (ssCOD) chemical oxygen demand	Twice a week	(Method: 5220 – B) ¹⁴
total (TA), partial (PA), and intermediary (IA) alkalinity	Twice a week	14,17
Total (TSS), volatile (VSS), and fixed (FSS) suspended solids	Twice a week	(Method: 2540 – C e 2540 – E) ¹⁴
Total volatile acids (TVA)	Twice a week	18
Kjeldahl nitrogen (KN)	Twice a week	(Method: 4500-N-C) ¹⁴
Total phosphorus (totalP)	Twice a week	(Method: 4500-P-C) ¹⁴
Sludge		
Total solids (TS) and volatile solids (VS)	Biweekly	(Method 2540 – B and 2540 – E) ¹⁴
Biogas		
Production	Daily	(Method: gasometer) ¹⁹
Composition	Weekly	14 (Method: gas chromatography)

reactors was corrected to standard temperature and pressure (0 °C and 1 atm) (STP).

Results and discussion

pH, alkalinity, and total volatile acids

In the first 158 (R1) and 172 (R2) days of the operation, the pH values of the influent when OLR was 0.2–7.5 g_{totalCOD} (L d)⁻¹ were between 6.5 and 7.0 (Figure 3) because the influent was corrected with NaOH solution. The ratio of total alkalinity (TA)/COD in the influent of the reactor was 0.07–0.11. After this period and when recirculation of the effluent began, the pH values of the influent increased to 7.0–7.5 in R1 and 6.0–7.0 in R2 (Figure 3). This increase was due to the increase in the TA/COD ratio in the influent to 1.5 because of the recirculation. The TA/COD ratio greater than 0.2 in the influent suggests that the alkalinity in the reactor is sufficient to be operated with stability.²⁰

The effluent pH in both the reactors ranged from 7.0 to 8.0. With the recirculation of the effluent, the pH remained steady around 7.8 in R1, which was operated with a smaller OLR (Figure 3). These values are close to the range 6.7–7.8 considered ideal for the development of methanogenic archaea.²¹

The average concentrations of TA in the effluents of both reactors increased from 532 and 558 mgL⁻¹ to approximately 4280 and 3394 mgL⁻¹, respectively, with the increase in OLR from 0.2 to 7.5 g_{totalCOD} (L d)⁻¹ (Table 2 and Figure 4).

In R2, after 193 days of operation, with an OLR of 11.5 g_{totalCOD} (L d)⁻¹ and HDT of 2.8 d, the influent pH decreased to below 6.0. Approximately 40% of the effluent was being recirculated to correct the pH. Souza et al.¹⁶ used at least 50% effluent recirculation rate to maintain an influent pH of 7.0 in a thermophilic UASB treatment of vinasse. Therefore, after 218 days, HDT in R2 was reduced from 2.8 to 1.8 d, which increased the volume of the recirculated effluent and thus contributed to pH correction to approximately 7.0 and to subsequent TA increase (Figures 3 and 4).

The average concentrations of total volatile acids (TVA) in the influent and effluent of the reactors increased with gradual increase in OLR. Maximum TVA was 1728 mgL⁻¹ with an OLR of 7.5 g_{totalCOD} (L d)⁻¹ in R1 and 2722 mgL⁻¹ with an OLR of 11.5 g_{totalCOD} (L d)⁻¹ in R2. The accumulation of TVA in the treatment of vinasse was mentioned by Souza et al.¹⁶ and Espinosa et al.²² The typical reactor response to rapid changes in OLR could lead to massive TVA concentrations, drop in pH, and consequent failure of the process.²³

With the methods used to control the influent pH, the average intermediate alkalinity (IA)/partial alkalinity (PA) ratios in the effluent were low (0.18–0.35) in R1 and higher (0.23–0.86) in R2 because of a higher OLR of 11.5 g_{totalCOD} (L d)⁻¹ applied (Table 2). According to Ripley et al.,²⁴ an IA/PA ratio of above 0.3 indicates the occurrence of disorders in the anaerobic digestion process.

The average TVA/TA ratios in the effluents were 0.13–0.31 in R1 and 0.12–0.50 in R2 (Table 2) with increased OLR. The largest TVA/TA ratios were obtained for an OLR of 11.5 g_{totalCOD} (L d)⁻¹ with increasing TVA concentration from 2596 to 4663 mgL⁻¹. This is because of the decrease in the total COD removal efficiency from 82% to 60%.

The TVA/TA ratio above 0.8 may inhibit methanogenic archaea, of 0.3–0.4 indicates an unstable system, and a ratio of 0.1–0.2 is appropriate.²⁵ Following this finding, instabilities existed only when R2 was operated with an OLR of 11.5 g_{totalCOD} (L d)⁻¹, and stable operation of both UASB reactors was possible with an OLR of up to 7.5 g_{totalCOD} (L d)⁻¹.

COD and suspended solids

The values of _{totalCOD} of the influent were 1866–21,971 mgL⁻¹ in R1 and 1866–28,543 mgL⁻¹ in R2 (Table 3) with the gradual increase in OLR. The average values of dissolved COD (dissCOD) were 84–89% of _{totalCOD}, indicating that most of the organic matter of the influent can be found mainly soluble.

The concentrations of total and volatile suspended solids (TSS and VSS) in the influent were 117–1910 mgL⁻¹ and 96–1556 mgL⁻¹, respectively (Table 3). VSS was 69–85% of TSS, thus indicating that organic suspended solids were predominant.

The maximum average efficiencies of _{totalCOD} and _{dissCOD} removal in the UASB reactors of 81% and 82%, respectively, were achieved with an OLR of 2.5–5.0 g_{totalCOD} (L d)⁻¹ in R1 and R2. With the increase in average OLR from 5.0 to 7.5 g_{totalCOD} (L d)⁻¹ in R1 and R2, and 7.5–11.5 g_{totalCOD} (L d)⁻¹ in R2, the removal efficiencies of _{totalCOD} and _{dissCOD} remained similar, 70–82%, respectively (Table 4 and Figure 5). Therefore, the strategy used (inoculated with granular sludge,

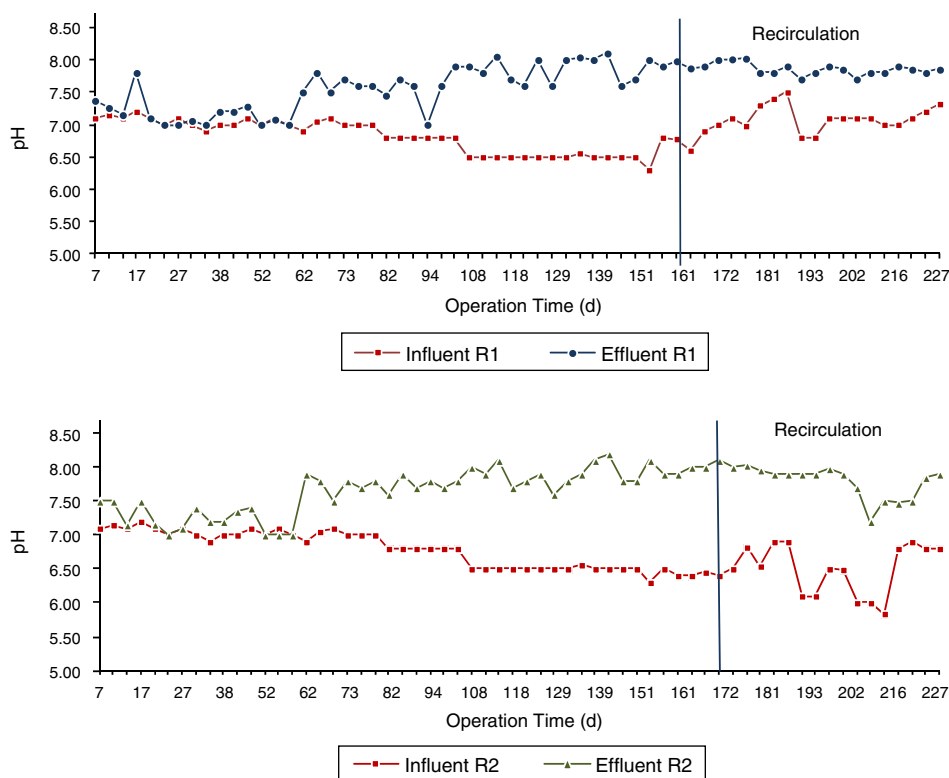


Figure 3 – pH values of the influents and effluents of UASB reactors (R1 and R2) for vinasse treatment.

Table 2 – Average values of the concentrations of total alkalinity (TA), total volatile acids (TVA) of the influents and effluents; ratios IA/PA and TVA/TA of the effluents; and OLR applied during the operation of the UASB reactors (R1 and R2) for vinasse treatment.

Attributes	R1				R2				
	0.2–2.5	2.5–5.0	5.0–7.5	7.5	0.2–2.5	2.5–5.0	5.0–7.5	7.5–11.5	11.5
OLR ($\text{g}_{\text{totalCOD}} \text{L d}^{-1}$)	0.2–2.5	2.5–5.0	5.0–7.5	7.5	0.2–2.5	2.5–5.0	5.0–7.5	7.5–11.5	11.5
Operation days	(0–115)	(116–147)	(148–186)	(187–229)	(0–115)	(116–147)	(148–172)	(173–193)	(193–229)
pH									
Influent	6.94	6.51	6.80	6.80	6.94	6.51	6.80	6.80	6.48
vc	2.7	0.3	4.4	2.8	2.7	0.3	4.4	2.9	8.8
Effluent	7.41	7.88	7.89	7.82	7.52	7.88	7.93	7.98	7.71
vc	4.5	2.8	1.8	7.8	4.5	2.4	1.4	1.0	3.3
TA ($\text{mg L}^{-1} \text{CaCO}_3$)									
Influent	216	781	2349	4828	216	781	873	2284	4386
vc	38	31	46	17	38	31	25	40	28
Effluent	532	2529	4280	6100	558	2517	3394	4090	6288
vc	64	20	20	29	57	34	24	18	20
TVA ($\text{mg L}^{-1} \text{CH}_3\text{COOH}$)									
Influent	158	943	1599	3050	158	943	1262	2596	4663
vc	77	44	37	6	78	44	30	11	10
Effluent	60	388	623	1728	61	328	499	454	2722
vc	48	67	23	23	43	56	15	16	48
IA/PA									
Effluent	0.32	0.18	0.24	0.35	0.32	0.16	0.36	0.23	0.86
vc	76	28	16	26	69	31	30	33	70
TVA/TA									
Effluent	0.13	0.15	0.14	0.31	0.12	0.13	0.15	0.11	0.50
vc	34	47	13	45	36	39	15	8	58

OLR, organic loading rate; IA, intermediary alkalinity; PA, partial alkalinity; vc, variation coefficient (%).

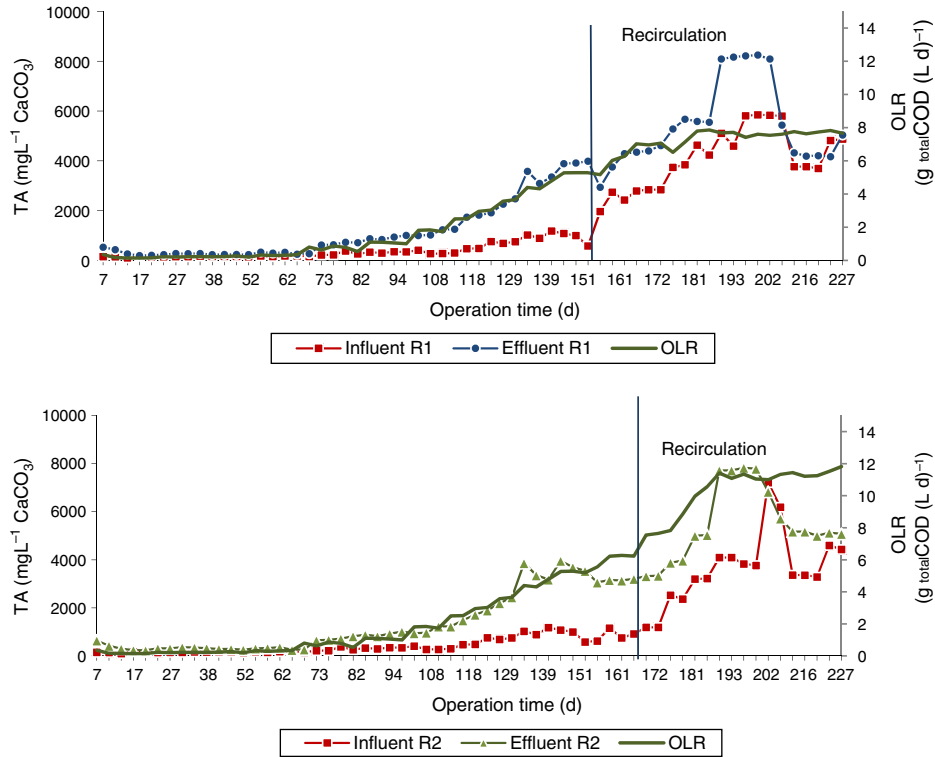


Figure 4 – Concentrations of total alkalinity (TA) as a function of the organic loading rate (OLR) applied in UASB reactors (R1 and R2) for vinasse treatment.

Table 3 – Average values of $total\ COD$, $diss\ COD$, TSS, and VSS of the influents and effluents and the OLR applied during the operation of UASB reactors (R1 and R2) for vinasse treatment.

Attributes	R1				R2				
	0.2–2.5	2.5–5.0	5.0–7.5	7.5	0.2–2.5	2.5–5.0	5.0–7.5	7.5–11.5	11.5
OLR ($g_{total\ COD} (L\ d)^{-1}$)	0.2–2.5	2.5–5.0	5.0–7.5	7.5	0.2–2.5	2.5–5.0	5.0–7.5	7.5–11.5	11.5
Operation days	(0–115)	(116–147)	(148–186)	(187–229)	(0–115)	(116–147)	(148–172)	(173–193)	(193–229)
$total\ COD^a$									
Influent	1866	10,377	17,554	21,971	1866	10,377	16,239	24,800	28,543
vc	94	22	13	3	94	22	9	15	17
Effluent	415	2037	5637	10,904	378	1888	3748	4404	10,540
vc	59	40	37	10	58	45	14	7	11
$total\ COD^a$									
Influent	1568	8897	15,344	18,809	1568	8897	14,569	21,430	24,103
vc	91	27	13	7	91	27	10	13	19
Effluent	335	1705	4852	9689	308	1579	3110	3765	9122
vc	75	33	37	13	74	34	13	10	16
TSS ^a									
Influent	139	480	1001	1869	117	480	775	1582	1910
vc	129	36	53	11	89	36	50	51	17
Effluent	30	201	731	1150	28	143	427	668	1100
vc	109	46	48	9	121	50	32	35	19
VSS ^a									
Influent	96	355	785	1385	88	355	663	1182	1556
vc	113	38	49	13	102	38	51	41	20
Effluent	19	137	428	696	14	102	285	339	706
vc	98	48	38	16	69	58	40	45	24

^a Unit: mgL^{-1} .

OLR, organic loading rate; $total\ COD$, total chemical oxygen demand; $diss\ COD$, dissolved chemical oxygen demand; TSS, total suspended solids; VSS, volatile suspended solids; vc, variation coefficient (%).

Table 4 – Average values of OLR and the removal efficiencies (in %) of $_{\text{total}}\text{COD}$, $_{\text{diss}}\text{COD}$, TSS, and VSS during the operation of the UASB reactors (R1 and R2) for vinasse treatment.

Attributes	R1				R2				
	0.2–2.5	2.5–5.0	5.0–7.5	7.5	0.2–2.5	2.5–5.0	5.0–7.5	7.5–11.5	11.5
Operation days	(0–115)	(116–147)	(148–186)	(187–229)	(0–115)	(116–147)	(148–172)	(173–193)	(193–229)
$_{\text{total}}\text{COD}$	67	81	67	49	69	82	77	82	60
vc	32	5	13	10	30	5	4	4	15
$_{\text{diss}}\text{COD}$	72	81	68	47	73	82	78	82	60
vc	26	3	13	13	22	3	5	3	17
TSS	64	59	24	38	65	73	46	54	41
vc	40	38	80	25	39	31	122	20	24
VSS	71	50	41	49	65	60	54	70	54
vc	44	85	51	10	48	72	21	12	12

OLR, organic loading rate; $_{\text{total}}\text{COD}$, total chemical oxygen demand; $_{\text{diss}}\text{COD}$, dissolved chemical oxygen demand; TSS, total suspended solids; VSS, volatile suspended solids; vc, variation coefficient (%).

pH correction with NaOH and application of increasing OLR) allowed the startup and stabilization of COD removal in UASB reactor with OLR up to $11.5 \text{ g}_{\text{total}}\text{COD} (\text{L d})^{-1}$.

The removal efficiencies of $_{\text{total}}\text{COD}$ decreased to approximately 50% and 60% with the recirculation of the effluent when OLR of 7.5 and $11.5 \text{ g}_{\text{total}}\text{COD} (\text{L d})^{-1}$ was applied in R1 and R2, respectively. This was due to the increase in the amount of compounds that cannot be easily degraded with the subsequent recirculation of the effluent.

For a successful startup of the anaerobic treatment of vinasse from wine distillery, Wolmarans and Villers²⁶ recommended the UASB reactor to be operated at an OLR of $4.0\text{--}8.0 \text{ g}_{\text{total}}\text{COD} (\text{L d})^{-1}$ until 90% $_{\text{total}}\text{COD}$ removal is

achieved, which will initiate the gradual increase of OLR. This result was similar to that of our work. However, other studies on sugarcane vinasse treatment in UASB reactors attained successful with COD removal at an efficiency below 90%,^{16,27} which resulted in an OLR up to $30 \text{ g}_{\text{total}}\text{COD} (\text{L d})^{-1}$.¹⁶ This confirms that it is possible to increase the OLR, but the COD removal decreases, as occurred in UASB reactors (R1 and R2) with OLR above $7.5 \text{ g}_{\text{total}}\text{COD} (\text{L d})^{-1}$.

The average values of TSS removal efficiency decreased from 64% to 38% in R1 and from 65% to 41% in R2 when OLR was increased from 0.2 to $7.5 \text{ g}_{\text{total}}\text{COD} (\text{L d})^{-1}$ in R1 and from 0.2 to $11.5 \text{ g}_{\text{total}}\text{COD} (\text{L d})^{-1}$ in R2 (Table 4). This decrease in TSS removal efficiency is due to wash out of the sludge in the reactors. This issue was solved by discarding 10% of the

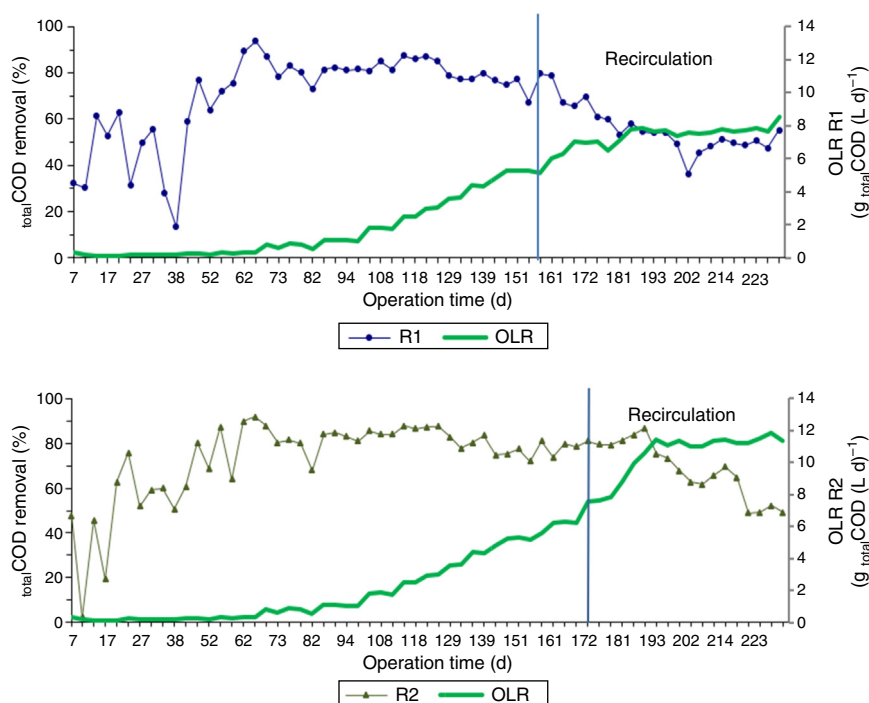


Figure 5 – Removal efficiency of the total chemical oxygen demand ($_{\text{total}}\text{COD}$) as a function of the organic loading rate (OLR) applied in the UASB reactors (R1 and R2) for vinasse treatment.

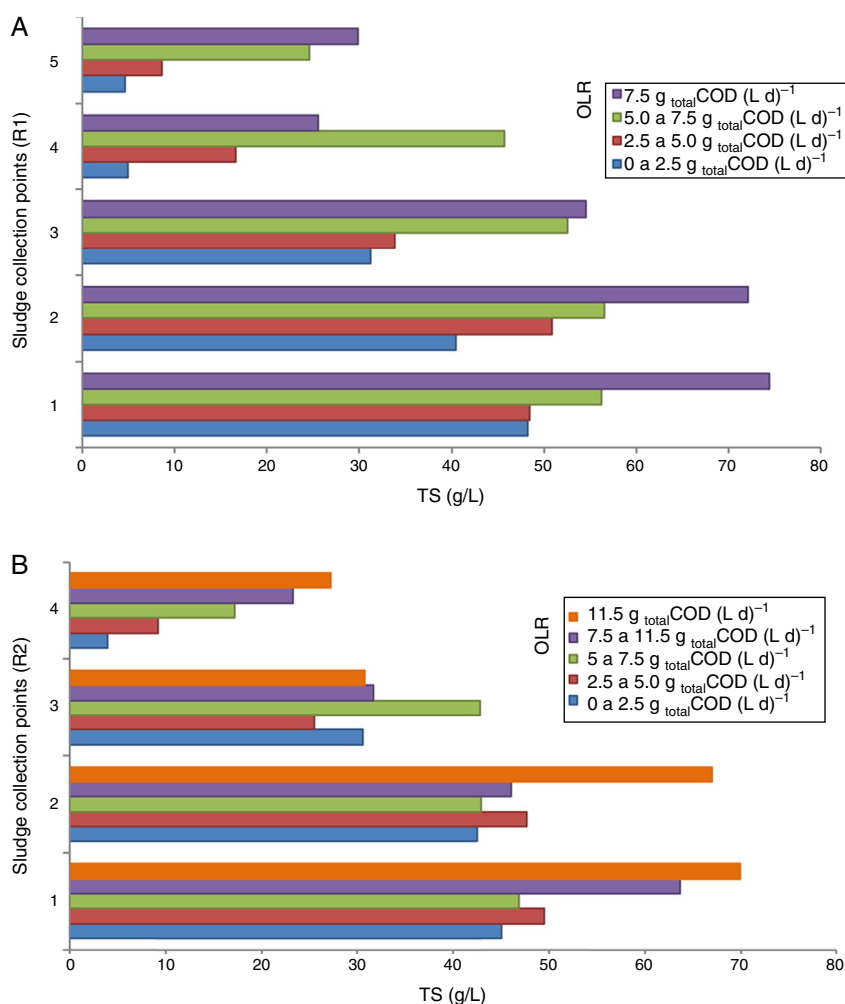


Figure 6 – Concentration of total solids (TS) in the sludge collected at the points shown in Figure 1 as a function of the organic loading rate (OLR) applied in the UASB reactors (R1 and R2) treating vinasse. (A) R1 and (B) R2.

volume of sludge blanket of both reactors with an OLR of $5.0\text{--}7.5\text{ g}_{\text{total COD}}(\text{L d})^{-1}$.

Total and volatile solids in the sludge

TS concentration in the sludge of the UASB reactors (R1 and R2) increased with OLR (Figure 6A and B), indicating that there was an increase in the sludge blanket of the reactors.

In R1, TS concentration in the sludge increased from 49 to 75 g L^{-1} , 41 to 72 g L^{-1} , 31 to 55 g L^{-1} , 5 to 28 g L^{-1} and 5 to 30 g L^{-1} at collection points P1, P2, P3, P4, and P5, respectively, and with an OLR of $0.2\text{--}7.5\text{ g}_{\text{total COD}}(\text{L d})^{-1}$ (Figure 6A).

In R2, TS concentration in the sludge increased from 45 to 70 g L^{-1} , 42 to 68 g L^{-1} , 31 to 45 g L^{-1} and 5 to 25 g L^{-1} at collection points P1, P2, P3, and P4, respectively, and with an OLR of $0.2\text{--}11.5\text{ g}_{\text{total COD}}(\text{L d})^{-1}$ (Figure 6B).

The decrease in TS concentration in the sludge from both reactors at collection points P3 and P4 is due to the disposal of excess sludge (10% of the volume of the sludge blanket). This prevented VSS wash out with the effluent and was performed with an OLR of $5.0\text{--}7.5\text{ g}_{\text{total COD}}(\text{L d})^{-1}$.

The ratio of volatile and total solids (VS/TS) in the sludge from the UASB reactors ranged from 0.54 to 0.76 (Table 5). These values indicate the predominance of organic matter in the sludge, and thus the presence of microorganisms, as confirmed by intensive conversion of total COD removed into methane (Table 6).

According to Brazilian legislation (Resolution n° 375²⁸), the sewage sludge or derived product is considered stable for agriculture use if $\text{VS/TS} < 0.70$. Therefore, it was observed that the sludge has been stabilized, especially in the top of the sludge blanket (collection points, P3, P4, and P5 from R1 and P3 and P4 from R2). Thus, when necessary, sludge disposal should be performed from P3 because at this point $\text{VS/TS} < 0.7$.

The organic load in the sludge (OLS) ranged from 0.16 to $0.42\text{ g}_{\text{total COD}}(\text{g VS d})^{-1}$ in R1 and 0.15 to $0.67\text{ g}_{\text{total COD}}(\text{g VS d})^{-1}$ in R2, with an increase in OLR (Table 5). The recommendation of Chernicharo¹² was followed. The OLS during the startup of the UASB reactors was maintained between 0.05 and $0.15\text{ g}_{\text{total COD}}(\text{g VS d})^{-1}$. It was gradually increased to a value lower than $2.0\text{ g}_{\text{total COD}}(\text{g VS d})^{-1}$ depending on the removal efficiencies.

Table 5 – Average values of the ratio of VS/TS in the sludge, OLS and OLR during the operation of the UASB reactors (R1 and R2) for vinasse treatment.

Attributes	R1				R2				
	0.2–2.5	2.5–5.0	5.0–7.5	7.5	0.2–2.5	2.5–5.0	5.0–7.5	7.5–11.5	11.5
OLR ($\text{g}_{\text{totalCOD}} (\text{L d})^{-1}$)	0.2–2.5	2.5–5.0	5.0–7.5	7.5	0.2–2.5	2.5–5.0	5.0–7.5	7.5–11.5	11.5
Operation days	(0–115)	(116–147)	(148–186)	(187–229)	(0–115)	(116–147)	(148–172)	(173–193)	(193–229)
P1									
VS/TS	0.67	0.74	0.68	0.75	0.65	0.75	0.73	0.69	0.74
vc	19	2	14	4	20	1	0	2	2
P2									
VS/TS	0.58	0.76	0.68	0.71	0.56	0.71	0.74	0.66	0.76
vc	13	0	15	2	1	6	0	5	1
P3									
VS/TS	0.63	0.73	0.68	0.69	0.58	0.69	0.74	0.61	0.65
vc	1	3	13	11	3	8	0	7	0
P4									
VS/TS	0.58	0.64	0.68	0.66	0.57	0.55	0.63	0.54	0.63
vc	18	18	4	11	24	6	0	2	4
P5									
VS/TS	0.58	0.56	0.59	0.53	–	–	–	–	–
vc	14	4	7	28	–	–	–	–	–
OLS ($\text{g}_{\text{totCOD}} (\text{g VS d})^{-1}$)	0.16	0.27	0.38	0.42	0.15	0.27	0.34	0.68	0.67

OLR, organic loading rate; OLS, organic load in the sludge; VS, volatile solids; TS, total solids; vc, variation coefficient (%). P1, point 1 (bottom); P2, point 2; P3, point 3; P4, point 4; P5, point 5 (top), as shown in Figure 1.

Methane

Methane percentage in the biogas decreased from 83% to 69% and from 85% to 64% in R1 and R2, respectively, when OLR was increased (Table 6). However, the volumetric methane production reached up to $0.8 \text{ L CH}_4 (\text{L reactor d})^{-1}$ in R1 and $1.3 \text{ L CH}_4 (\text{L reactor d})^{-1}$ in R2 when higher OLR values of 7.5 and $11.5 \text{ g}_{\text{totalCOD}} (\text{L d})^{-1}$ were applied, respectively (Figure 7). The highest average values of the volumetric methane production were 0.597 and $0.989 \text{ L CH}_4 (\text{L d})^{-1}$ with OLR values of $5.0\text{--}7.5 \text{ g}_{\text{totalCOD}} (\text{L d})^{-1}$ in R1 and $7.5\text{--}11.5 \text{ g}_{\text{totalCOD}} (\text{L d})^{-1}$ in R2.

The specific methane production increased from 0.133 to $0.181 \text{ L CH}_4 (\text{g}_{\text{totalCOD removed}})^{-1}$ in R1 and from 0.145 to $0.185 \text{ L CH}_4 (\text{g}_{\text{totalCOD removed}})^{-1}$ in R2 with the application of an OLR of $0.2\text{--}7.5 \text{ g}_{\text{totalCOD}} (\text{L d})^{-1}$. With higher values of OLR and effluent recirculation, the average values of specific methane production decreased to 0.172 and $0.115 \text{ L CH}_4 (\text{g}_{\text{totalCOD removed}})^{-1}$ in R1 and R2, respectively (Table 6). The methane yield was below the theoretical value of $0.35 \text{ L CH}_4 (\text{g COD removed})^{-1}$ calculated stoichiometrically. The methane yield obtained by Souza et al.,¹⁶ was $0.37 \text{ L CH}_4 (\text{g COD removed})^{-1}$ with an OLR of $26.5 \text{ g COD} (\text{L d})^{-1}$ and that obtained by España-Gamboa et al.,⁶ was $0.26 \text{ L CH}_4 (\text{g COD removed})^{-1}$.

Table 6 – Average values of OLR, volumetric and specific methane production, and mass balance for conversion of removed totalCOD into methane during operation of UASB reactors (R1 and R2) for vinasse treatment.

Attributes UASB reactor	OLR ($\text{g}_{\text{totalCOD}} (\text{L d})^{-1}$)	Volumetric methane production* ($\text{L CH}_4 (\text{L d})^{-1}$)		Specific methane production* ($\text{L CH}_4 (\text{g}_{\text{totalCOD removed}})^{-1}$)		Mass balance (removed totalCOD converted into CH_4) [†] (%)
			vc		vc	
R1	0.2–2.5	0.087	80	0.133	48	58
	2.5–5.0	0.440	43	0.175	30	51
	5.0–7.5	0.597	18	0.181	20	48
	7.5	0.554	4	0.172	35	48
R2	0.0–2.5	0.120	57	0.145	33	65
	2.5–5.0	0.550	34	0.179	22	52
	5.0–7.5	0.829	21	0.185	14	53
	7.5–11.5	0.989	15	0.138	25	39
	11.5	0.938	22	0.115	30	42

Values adjusted for standard temperature and pressure (STP) (0°C and 1 atm).

OLR, organic loading rate; totalCOD, total chemical oxygen demand; vc, variation coefficient (%).

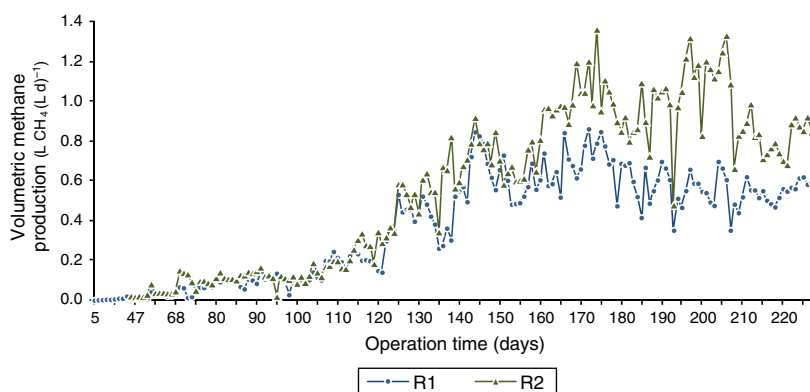


Figure 7 – Volumetric methane production during operation of UASB reactors (R1 and R2) for vinasse treatment.

removed)⁻¹ with an OLR of 17.0 g COD (L d)⁻¹ in the vinasse treatment in thermophilic UASB reactors. These results suggest that the vinasse anaerobic treatment conducted in the thermophilic phase allows application of larger OLR and yields higher specific methane production.

The average conversion rates of total COD removed into methane in R1 and R2 were 48–58% and 39–65%, respectively (Table 6). These values are higher by more than 50% compared with those observed by Harada et al.,²⁷

Assuming that 10% of removed COD was converted in the sludge, as indicated by Chernicharo,¹² methane loss of 25% and 51% was found in R1 and R2. A significant portion of gases generated in the anaerobic treatment can remain dissolved in the liquid and be expelled out with the treated effluent.

The methane loss in the effluent from the UASB reactors can vary from 20% to 50%.²¹ Therefore, the values assigned to the sludge production and methane losses in the effluent can result in up to 30–60% of COD removal. These values are within the range quoted for R1 and R2.

Nitrogen and phosphorus

The average concentration of total P and KN in the influent increased from 56 to 476 mg L⁻¹ and from 18 to 63 mg L⁻¹, respectively, due to the nutrient supplementation and the increase of OLR in R1 and R2 (Table 7). Supplementation of nitrogen, phosphorus, and potassium can reduce the effects of possible shock loads and prevent flotation of granules in UASB

Table 7 – Average concentrations (mg L⁻¹) of Kjeldahl nitrogen (KN), total phosphorus (total P) and potassium (K) in the influents and effluents; removal efficiency (E in %); and OLR during operation of UASB reactors (R1 and R2) for vinasse treatment.

Attributes	R1				R2				
	0.2–2.5	2.5–5.0	5.0–7.5	7.5	0.2–2.5	2.5–5.0	5.0–7.5	7.5–11.5	11.5
OLR (g _{total} COD (L d) ⁻¹)	0.2–2.5	2.5–5.0	5.0–7.5	7.5	0.2–2.5	2.5–5.0	5.0–7.5	7.5–11.5	11.5
Operation days	(0–115)	(116–147)	(148–186)	(187–229)	(0–115)	(116–147)	(148–172)	(173–193)	(193–229)
KN									
Influent	56	144	345	470	56	144	261	428	476
vc	56	28	48	22	56	28	29	25	15
Effluent	44	99	280	365	40	113	200	341	367
vc	39	33	53	20	40	18	25	43	14
E	33	40	36	27	36	33	32	34	30
vc	54	60	33	53	41	49	77	45	41
total P									
Influent	18	52	65	52	18	52	79	56	63
vc	26	27	37	49	26	27	12	43	48
Effluent	10	22	16	8	16	52	59	28	48
vc	96	5	47	43	109	7	22	53	45
E	61	30	41	58	49	15	41	48	38
vc	66	134	75	43	78	79	72	59	74
K									
Influent	4	17	25	8	4	17	26	5	9
vc	53	47	66	32	52	47	30	52	13
Effluent	4	13	22	9	4	14	27	5	8
Vc	52	69	81	23	56	36	33	23	30

OLR, organic loading rate; vc, variation coefficient (%).

reactors. Among several possible formulations for supplemental phosphorus and potassium, KH_2PO_4 is recommended owing to its buffer capacity.¹⁵

Although the vinasse from sugarcane contains high concentrations of potassium ion, approximately 5 g L^{-1} ,²⁰ it was used as the source of phosphorus KH_2PO_4 because, according to Chen et al.,²⁹ almost no reports of toxic effects of potassium on the microbiota of anaerobic reactors.

The ratio of COD:N:P in the influent varied from 350:4.8:0.8 to 350:7.4:1.7 in R1 and R2 for OLR greater than $2.5 \text{ g total COD (L d)}^{-1}$. These values are close to that suggested by Chernicharo¹² sufficient for satisfying the conditions of microorganisms for anaerobic digestion.

In the effluent of the reactors, totalP and KN were lower than those observed in the influent and ranged from 44 to 367 mg L^{-1} and 10 to 52 mg L^{-1} , respectively. The average removal efficiencies of KN and totalP were 27–40% and 30–61%, respectively, in R1 and 30–36% and 15–49%, respectively, in R2 (Table 7). According to Oliveira et al.,³⁰ one of the possible mechanisms for nitrogen and phosphorus removal is the formation of struvite ($\text{NH}_4\text{MgPO}_4 \cdot 6\text{H}_2\text{O}$) and vivianite ($\text{Fe}_3(\text{PO}_4)_2 \cdot 8\text{H}_2\text{O}$) in addition to immobilization in the microbiota from the sludge blanket.

The potassium concentration in the effluent of the reactors was similar to that in the influent (Table 7). Therefore, the KN, totalP , and K available in the effluent and retained in the sludge can be used for fertigation and fertilization, and thus can partly replace mineral fertilizers and reduce production costs.

Conclusions

The highest totalCOD conversion into methane of $0.19 \text{ L CH}_4 (\text{g totalCOD removed})^{-1}$ was achieved after 140 days of operation of the UASB reactors with totalCOD removal efficiencies of approximately 70% and 80%, and an OLR of $5.0\text{--}7.5 \text{ g totalCOD (L d)}^{-1}$. The highest totalCOD removal efficiencies were 81% and 82% in R1 and R2, respectively, with an OLR of $2.5\text{--}5.0 \text{ g totalCOD (L d)}^{-1}$. Recirculation of the effluent allowed adjustment of influent pH without the need to add sodium hydroxide. The UASB reactors produced methane with high efficiency, a better quality effluent, and stable sludge. The nutrients present in the vinasse and those obtained from the supplements in the anaerobic treatment can be recycled by using the effluent in fertigation and the sludge for plant fertilization.

Conflicts of interest

The authors declare no conflicts of interest.

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REFERENCES

- Wang W, Xie L, Luo G, Zhou Q, Lu Q. Optimization of biohydrogen and methane recovery within a cassava ethanol wastewater/waste integrated management system. *Bioresour Technol.* 2012;120:165–172.
- Sawin JL, Bhattacharya SC, Galan EM, et al. REN21 – *Renewables Global Status Report*. Ren21; 2012:1–172. <http://www.ren21.net/REN21Activities/GlobalStatusReport.aspx>.
- Laíme E, Fernandes P, Oliveira DCS, Freire EDA. Possibilidades tecnológicas para a destinação da vinhaça: uma revisão *Technological possibilities for the disposal of vinasse: a review*. *Rev Tróp – Ciênc Agrár Biol.* 2011;5(3):16–29.
- Ferreira LFR, Aguiar MM, Messias TG, et al. Evaluation of sugar-cane vinasse treated with *Pleurotus sajor-caju* utilizing aquatic organisms as toxicological indicators. *Ecotoxicol Environ Saf.* 2011;74(1):132–137.
- Wilkie AC, Riedesel KJ, Owens JM. Stillage characterization and anaerobic treatment of ethanol stillage from conventional and cellulosic feedstocks. *Biomass Bioenergy.* 2000;19(2):63–102.
- España-Gamboa EI, Mijangos-Cortés JO, Hernández-Zárate G, Maldonado JAD, Alzate-Gaviria LM. Methane production by treating vinasses from hydrous ethanol using a modified UASB reactor. *Biotechnol Biofuels.* 2012;5(1):82.
- Khemkhao M, Nuntakumjorn B, Techkarnjanaruk S, Phalakornkule C. UASB performance and microbial adaptation during a transition from mesophilic to thermophilic treatment of palm oil mill effluent. *J Environ Manage.* 2012;103:74–82.
- Kaparaju P, Serrano M, Angelidaki I. Optimization of biogas production from wheat straw stillage in UASB reactor. *Appl Energy.* 2010;87(12):3779–3783.
- Satyawali Y, Balakrishnan M. Removal of color from biometanated distillery spentwash by treatment with activated carbons. *Bioresour Technol.* 2007;98:2629–2635.
- Vaccari G, Tamburini E, Sgualdino G, Urbaniec K, Klemeš J. Overview of the environmental problems in beet sugar processing: possible solutions. *J Clean Prod.* 2005;13(5):499–507.
- Akarsubasi AT, Ince O, Oz NA, Kırdar B, Ince BK. Evaluation of performance, acetoclastic methanogenic activity and archaeal composition of full-scale UASB reactors treating alcohol distillery wastewaters. *Process Biochem.* 2006;41(1):28–35.
- Chernicharo CAL. *Reatores Anaeróbios Princípios Do Tratamento Biológico Em águas Residuárias*. 2nd ed. Belo Horizonte: DESA/UFMG; 2007.
- Parnaudeau V, Condom N, Oliver R, Cazevieille P, Recous S. Vinasse organic matter quality and mineralization potential, as influenced by raw material, fermentation and concentration processes. *Bioresour Technol.* 2008;99(6):1553–1562.
- APHA. *Standard Methods for the Examination of Water and Wastewater*. 21st ed. Washington: American Public Health Association; 2005.
- Singh RP, Kumar S, Ojha CSP. Nutrient requirement for UASB process: a review. *Biochem Eng J.* 1999;3(1):35–54.
- Souza ME, Fuzaro G, Polegato AR. Thermophilic anaerobic digestion of vinasse in pilot plant UASB reactor. *Water Sci Technol.* 1992;25(7):213–222.
- Jenkins SR, Morgan JM, Sawyer CL. Measuring anaerobic sludge digestion and growth by a simple alkalimetric titration. *J Water Pollut Control Fed.* 1983;55(5):448–453.

18. DiLallo R, Albertson OE. Volatile acids by direct titration. *J Water Pollut Control Fed.* 1961;33:356–365.
19. de Oliveira RA [thesis] *Efeito da concentração de sólidos suspensos do afluente no desempenho e características do lodo de reatores anaeróbios de fluxo ascendente com manta de lodo tratando águas residuárias de suinocultura.* São Carlos, Brasil: Escola de Engenharia, USP; 1997, 359 pp.
20. Doll MMR, Foresti E. Effect of the sodium bicarbonate in the treatment of vinasse in AnSBBR operated at 55 and 35 °C. *Eng Sanit E Ambient.* 2010;15:275–282.
21. Van Haandel AC. Influence of the digested COD concentration on the alkalinity requirement in anaerobic digesters. *Water Sci Technol.* 1994;30:23–34.
22. Espinosa A. Effect of trace metals on the anaerobic degradation of volatile fatty acids in molasses stillage. *Water Sci Technol.* 1995;32(12):121–129.
23. Leitao R, Van Haandel A, Zeeman G, Lettinga G. The effects of operational and environmental variations on anaerobic wastewater treatment systems: a review. *Bioresour Technol.* 2006;97(9):1105–1118.
24. Ripley LE, Boyle WC, Converse JC. Improved alkalimetric monitoring for anaerobic digestion of high-strength wastes. *J Water Pollut Control Fed.* 1986;58(5):406–411.
25. Zhao HW, Viraraghavan T. Analysis of the performance of an anaerobic digestion system at the Regina wastewater treatment plant. *Bioresour Technol.* 2004;95(3):301–307.
26. Wolmarans B, De Villiers GH. Start-up of a UASB effluent treatment plant on distillery wastewater. *Water SA.* 2002;28(1):63–68.
27. Harada H, Uemura S, Chen A-C, Jayadevan J. Anaerobic treatment of a recalcitrant distillery wastewater by a thermophilic UASB reactor. *Bioresour Technol.* 1996;55(3):215–221.
28. Conselho Nacional do Meio Ambiente. Resolução N. 375; 2006:1–32. <http://www.mma.gov.br/port/conama/legislacao/CONAMA.RES.CONS.2008.401.pdf>.
29. Chen Y, Cheng JJ, Creamer KS. Inhibition of anaerobic digestion process: a review. *Bioresour Technol.* 2008;99(10):4044–4064.
30. de Oliveira RA, Vazoller RF, Foresti E. Sludge bed characteristics of UASB reactors: growth, activity, microbial structure and chemical composition of granules. In: *8th International Conference on Anaerobic Digestion.* 1997: 524–531.