



## How climate change can affect the distribution range and conservation status of an endemic bird from the highlands of eastern Brazil: the case of the Gray-backed Tachuri, *Polystictus superciliaris* (Aves, Tyrannidae)

Diego Hoffmann<sup>1,4</sup>, Marcelo Ferreira de Vasconcelos<sup>2</sup> & Rogério Parentoni Martins<sup>3</sup>

<sup>1</sup>Departamento de Ciências Agrárias e Biológicas, Universidade Federal do Espírito Santo, Rod. BR 101 Norte, KM 60, Bairro Litorâneo, CEP 29932-540, São Mateus, ES, Brazil.

<sup>2</sup>Museu de Ciências Naturais, Pontifícia Universidade Católica de Minas Gerais, Avenida Dom José Gaspar, 290, Prédio 41, Coração Eucarístico, CEP 30535-610, Belo Horizonte, MG, Brazil.

<sup>3</sup>Departamento de Biologia, Universidade Federal do Ceará, Centro de Ciências, Avenida da Universidade, 2853, CEP 60020-181, Fortaleza, CE, Brazil.

<sup>4</sup>Corresponding author: Diego Hoffmann, e-mail: [diego\\_hoffmann@hotmail.com](mailto:diego_hoffmann@hotmail.com)

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**Abstract:** The Gray-backed Tachuri (*Polystictus superciliaris*) is a Tyrannidae restricted to eastern Brazilian highlands. Its population and range are still thought to be declining mainly due to habitat loss, caused by land use. We evaluated the impacts on its range (increase or decrease, displacement in latitude and altitude), considering possible effects caused by inappropriate land use and climatic changes. For this purpose, we modeled its current range and estimated its overlap between the predicted area and future scenarios. We also analyzed the efficiency of the Brazilian reserves to the species' protection. The range established by the Maxent model was 76.1% lower than that available in the literature and presented a decrease in the future. Range contraction for the periods under consideration ranged from 22.5% to 77.3%. The distribution center shifts to the southwest (from 102.5 km to 275.4 km) and to higher elevations (1,102 m to 1,428 m). From the predicted range, 20% are under some kind of human occupation. The Brazilian reserves partially protect the species, with only 12% of its range, from the 44.2% expected. Based on these predictions, and depending on the species' dispersal ability and adaptation, it may become vulnerable in the period of 2080. This study demonstrates that the most important areas for maintaining populations of *P. superciliaris* now and in the future are those located in the 'Iron Quadrangle', especially the Serra da Gandarela, where it was partially protected by the recently decreed Gandarela National Park. This region should receive special attention because it is strongly threatened by iron mining in areas relevant for conservation located outside this reserve.

**Keywords:** Climate change, Evaluation reserve efficiency, Espinhaço range, Species distribution model.

HOFFMANN, D., VASCONCELOS, M.F., MARTINS, R.P. Como as mudanças climáticas podem afetar a área de distribuição e o status de conservação de uma ave endêmica das montanhas do leste do Brasil: o caso do papa-moscas-de-costas-cinzentas, *Polystictus superciliaris* (Aves, Tyrannidae). Biota Neotropica. 15(2): e20130075. <http://dx.doi.org/10.1590/1676-06032015007513>

**Resumo:** O papa-moscas-de-costas-cinzentas (*Polystictus superciliaris*) é uma espécie de Tyrannidae endêmica dos topos de montanha do leste do Brasil. Sua população e área de distribuição estão em declínio principalmente devido à perda de habitat causada pelo uso inadequado da terra. Nós avaliamos os impactos sobre sua distribuição (aumento ou diminuição, deslocamento em latitude e altitude), considerando os possíveis efeitos causados pelo uso do solo e pelas mudanças climáticas. Para esse propósito, modelamos a distribuição atual e avaliamos a sobreposição entre a área predita e os cenários futuros. Também avaliamos a eficiência das Unidades de Conservação brasileiras para a proteção da espécie. A área de distribuição estimada por meio do programa Maxent foi 79,1% menor que a área encontrada na literatura, apresentando uma tendência a diminuição no futuro. Foi projetada uma contração de 22,5% a 77,3% da área de distribuição atual para o futuro. O centro da área de distribuição deslocou-se para o sudeste (de 102,5 a 275,4 km) e para áreas com alta elevação (1.102 m a 1.428 m). Da área projetada, 20% encontra-se sob algum tipo de ocupação humana. As Unidades de Conservação brasileiras protegem parcialmente a espécie, com somente 12% de sua distribuição quando o esperado seriam 44,2%. Baseado nas projeções e, dependendo das habilidades de dispersão e adaptação da

espécie, ela pode se tornar vulnerável no período de 2080. Este estudo demonstrou que as áreas mais importantes para a manutenção de *P. superciliaris* na atualidade e em períodos futuros situam-se no Quadrilátero Ferrífero, especialmente na Serra da Gandarela, onde parte foi recentemente decretada como Parque Nacional da Gandarela. Essa região deve receber atenção especial por ser altamente ameaçada pela exploração do minério de Ferro e por áreas relevantes continuarem desprotegidas fora da abrangência do novo Parque.

**Palavras-chave:** Mudanças climáticas, Avaliação da eficiência de Reservas, Cadeia do Espinhaço, Modelos de distribuição de espécies.

## Introduction

Current climatic models predict global warming during the present century (IPCC 2007b). It is likely that the ranges of many species will change drastically (IPCC 2007a). Recent studies have shown that several species are responding to these changes and this pattern tends to increase in the future (e.g. Hughes 2000, Parmesan & Yohe 2003). Thus, it is expected that these changes will also increase the threat to some species, especially those living at high latitudes and elevations (Millennium Ecosystem Assessment 2005). In this case, largest impacts of climate change are predicted to occur in the high mountain ranges, for example, in the Alps (Beniston et al. 1997, Hughes 2000, Sérgio 2003). This would occur because species living in mountains tend to be more sensitive to climatic changes and present limited ability to adjust their ranges because of the decreasing of available surface with the increasing of elevation (Thuiller et al. 2005, Rull & Vegas-Vilarrúbia 2006, Pauli et al. 2007).

In this way, it is essential to predict effects and potential management strategies for these species by developing practices aiming to measure how they can be affected by these changes (Jiguet et al. 2010). Aiming the projection of the future suitable range of several species and possible impacts of climatic change on them, several modeling algorithms have been used to define the relationship between the current range of species and the current climatic variables (Thuiller et al. 2005, Heikkinen et al. 2006, Virkkala et al. 2008).

Several authors have assessed the potential impacts of climate change on Neotropical birds (e.g. Anciães & Peterson 2006, Nunes et al. 2007, Marini et al. 2009b, a, Marini et al. 2010). Nevertheless, studies focusing birds endemic to high-altitude areas are very scarce and restricted to the Andean region (e.g. Graham et al. 2010, Şekercioglu et al. 2012).

In this respect, eastern Brazilian mountains are located between two important global and highly threatened hotspots: the Atlantic Forest and the Cerrado (Silva & Bates 2002, Mittermeier et al. 2004). These mountains are also considered important centers of diversity and endemism of plants (Giulietti et al. 1997, Safford 1999) and animals (Rodrigues 1988, Stattersfield et al. 1998, Vasconcelos 2008). One of these endemic species is the Gray-backed Tachuri (*Polystictus superciliaris*), a bird from the family Tyrannidae, which occurs from central Bahia state to northern São Paulo state (Vasconcelos et al. 2003, Vasconcelos 2008). This species lives in the 'campos rupestres' (rocky fields) and 'campos de altitude' (high altitude grasslands), between 950 m and 1,950 m (Sick 1997, Stattersfield et al. 1998, Vasconcelos et al. 2003). Gray-backed Tachuri populations are thought to be declining due to its restricted range associated to habitat loss (Stattersfield et al. 1998, Vasconcelos et al. 2003, BirdLife International 2014). Nevertheless, the species' known range presented in the basic literature and used to infer its current conservation status (see Ridgely & Tudor 2009) does not

include several localities where it has recent confirmed records (Vasconcelos et al. 2003, Vasconcelos 2008), and includes many areas where conditions are not propitious to its occurrence (e.g., lowlands, forests, and degraded areas), based on our field observations. This is due to the methodology employed by those authors, which basically consists on the union of localities (points) where the species has been recorded, which creates a range. This usually includes deviations of species' ranges, because it is restricted to mountaintops, without records in adjacent lowlands. For an example in the case of the Gray-backed Tachuri, compare the figure with species' range presented by Ridgely & Tudor (2009) to figure 5 of Vasconcelos (2008).

For species with restricted ranges, including most of the mountaintops endemics, habitat loss is the main factor influencing populations decline and their vulnerabilities (Davies et al. 2009). Knowledge on the impacts of climatic changes and habitat loss on such species is essential to understand how they may persist in the future, specially in regions facing rapid socio-economic development (Pressey et al. 2007), such as Eastern Brazil. Thus, range assessment is important for endemic species, such as the the Gray-backed Tachuri, for which there is little information and, therefore, is likely to become threatened in the future. Furthermore, this approach is also important to assess its conservation status (Anderson & Martínez-Meyer 2004, Hernandez et al. 2006) and also to determine its actual level of threats and protection (Fuller et al. 2006). In this context, such procedures may contribute to the correct management and successful application of resources for conservation, which are usually scarce.

Under this perspective, we inferred the current and the future ranges with suitable conditions for the Gray-backed Tachuri, in order to assess possible impacts of climatic changes and find new areas with favorable conditions for its occurrence outside its known range, which can be used to guide the application of resources for its conservation. We also evaluated the percentage of its range that coincides with current protected areas (based on available database of reserves decreed until 2011), to determine the need of investments in its protection and to identify possible new areas for its conservation, since it is possible that the species is not fully protected by not occurring in several reserves. Thus, we expect to provide new information for the protection of the Gray-backed Tachuri and its habitat preservation, in order to prevent that this endemic bird will be among the endangered species in the near future.

## Materials and Methods

### 1. Species occurrence locations

In order to determine the localities of occurrence for the Gray-backed Tachuri, we used records obtained by specimens deposited in museums, literature review and our own field

records (see Vasconcelos 2008, Vasconcelos & Rodrigues 2010 for sources). Geographic coordinates from museum specimens were obtained from the collection sites listed in their original labels or based on Paynter Jr. & Traylor Jr. (1991) and Vanzolini (1992). All records with geographic coordinates were plotted within a grid of 30" cells ( $\sim 0.833 \times 0.833 \text{ km}^2$ ). Duplicate records within a single cell were removed.

## 2. Environmental dataset and climatic change scenarios

We discarded variables highly correlated and used the six remainder bioclimatic variables of temperature and precipitation for projections of current and future ranges with suitable conditions. These variables were: minimum temperature of the coldest month, maximum temperature of the warmest month, mean annual temperature, annual precipitation, precipitation of the driest month and precipitation of the wettest month. The dataset was obtained from WorldClim (Hijmans et al. 2005), based on interpolations of observed data, representative of 1950-2000. We also used three topographic variables (slope, aspect and compound topographic index), obtained from the database of the US Geological Survey's Hydro1k base (<http://eros.usgs.gov/>). Soil data was obtained from FAO (1998). Vegetation layers were not included in the models because they are not available for future scenarios. All variables used for species distribution model (current and future periods) were standardized to the resolution of 30" over into the Brazilian's political limits.

For future projections, we used the same variables for each period (2020, 2050 and 2080), by considering three General Circulation Models of gases (GCMs) (CCMA, HADCM3 and CSIRO) and two scenarios of the Special Report Emissions Scenarios (SRES) of the IPCC (2007a) (A2a e B2a), which reflect the potential impacts of different possible demographic, socio-economic and technological developments in the emission of greenhouse gases.

## 3. Distribution modeling

We used Maxent version 3.2.19 with the default setting (Phillips et al. 2006) to model current and future potential suitable ranges for the Gray-backed Tachuri. The projection of current distribution was generated using 75% of occurrence records randomly selected and evaluated with the remaining 25%. The models were evaluated using an area under the relative operating characteristic curve (AUC/ROC) (Fielding & Bell 1997). The curve considers the rate of failure in relation to the predicted fractional area. The data were divided this way five times and it was calculated the average AUC with a more robust estimate of the predictive performance. We used all available data to obtain the final distribution range projection.

For the projection of the suitable range for the Gray-backed Tachuri, in each of the three future periods, we obtained a consensus among six projections (three GCMs x two SRES scenarios), maximizing the accuracy and achieving a central trend. To transform the probabilistic ranges (present and future) in a binary map of presence/absence, we selected a threshold value that would encompass all sites (Pearson et al. 2007). Thus, we considered the possibility of 5% error in the forecast, because, occasionally, some records could represent individuals outside of its optimal niche or in transitional areas. All cells below the threshold value were set to '0' and kept the

probability of occurrence of the above. The area was calculated by summing all pixels that presented the average probability of conditions higher than '0' among the six projections of presence and absence. Finally, we removed from the predicted areas the overlapping lands already converted into intensive agriculture, urban areas and water bodies, using a land cover map presented by Eva et al. (2002).

In order to analyze the responses of the Gray-backed Tachuri to the climatic changes based on its range, we evaluated the extension and central point (centroid of its range) of predicted range in future periods. The distribution range was calculated by summing all pixels above the threshold value. To assess whether there was range shift, we calculated the distance between the centers of current and future ranges.

For each period, we estimated the ratio between the size of the future and current range, as a measurement of potential impact of climatic changes. This estimate assumes that the species is able to disperse from its present range to a predicted range in the future, without any environmental or physiological impairment (hypothesis of total dispersion). We also calculated the percentage of overlap of the range of each future period in relation to the present, which assumes that the species is unable to disperse into new areas and may occupy only those areas with favorable current conditions and that are predicted in the future as climatically favorable (hypothesis of null dispersion).

To assess whether there was an altitudinal shift, we calculated the difference between the average elevation of the future range projected for each period and the current projection. In this case, the altitude was extracted from all pixels of the predicted range by overlapping a relief map from the US Geological Survey's Hydro1k base (<http://eros.usgs.gov/>) for subsequent calculation of its average.

## 4. Reserves data set and species representation

Data on the distribution of protected areas in Brazil were obtained from the databases available in ICMBio (<http://www4.icmbio.gov.br>). This database consists of about 1,000 federal, state and municipal reserves, implemented until 2006 (equivalent to reserves of category II of IUCN) (MMA 2006). We performed gap analyzes, considering the projection of the current suitable range for the Gray-backed Tachuri and projections for future periods. For each period, an analysis was performed with all reserves. The predicted suitable range was overlapped with reserves distribution and we calculated the percentage of this range found in the protected area. We followed the methods proposed by Rodrigues et al. (2004) and Catullo et al. (2008) in order to identify a representation target (Scott et al. 1993), depending on the extent of the predicted suitable range for the species. Representation is the amount of the range of a species within protected areas, a parameter commonly used in conservation plans to evaluate the need for protection. Analyses were performed considering the following representation targets: i) species with restricted range (area of potential presence less than 1,000  $\text{km}^2$ ), which should be protected in 100% of its range; ii) widespread species (area of potential presence larger than 250,000  $\text{km}^2$ ), which should be protected in 10%; and iii) species with an area of potential presence of more than 1,000  $\text{km}^2$  and less than 250,000  $\text{km}^2$ , which should have a representation target obtained by interpolation between the two extremes, using a linear regression of log-transformed area of potential presence

(Rodrigues et al. 2004). A total gap occurs when a species is not represented in any reserve, a partial gap occurs when a species is represented only partially, and the species is considered covered when the representation target is fully met (Catullo et al. 2008). Representation under climate change was evaluated considering total and null capacity of dispersion (for each scenarios, present, 2020, 2050 and 2080).

## Results

To model the potential distribution, we used 41 spatially independent points of occurrence of the Gray-backed Tachuri. The accuracy of the distribution model in predicting favorable conditions for the presence of the Gray-backed Tachuri was better than random. The AUC mean, based on test data generated from five models, was 0.993 (SD 0.007) and the selected cumulative threshold was 57. The omission range of the test data was 0.08 (SD 0.11) and the fractional predicted area was 0.003.

### 1. Current distribution

The predicted current distribution of the Gray-backed Tachuri, with favorable conditions, was 30,745 km<sup>2</sup> (Figure 1). This range was 76.1% lower than that presented by the literature (129,012 km<sup>2</sup>) (Ridgely & Tudor 2009) (Figure 2a) and 90.2% lower than that used to its conservation status assessment (314,000 km<sup>2</sup>) (BirdLife International 2014). The range with natural vegetation currently available for the species was 25,240 km<sup>2</sup>, not considering those areas converted by intensive agriculture, urban areas and water bodies (5,505 km<sup>2</sup>).

### 2. Distribution in future scenarios

The suitable range for the Gray-backed Tachuri exhibited a contraction for future periods in relation to present range with favorable conditions. Under the null hypothesis of dispersion for the 2080 period, only a fifth of the current range remains (Figures 2b-d). Considering the hypothesis of total dispersion, the decrease in the current distribution for 2080 was of 64.7% (Table 1). Under the null hypothesis of dispersion, the range decrease was of 77.3%.

The range shift in future periods showed a gradual increase until 2080. This displacement was of 275.4 km under the total dispersion hypothesis and of 250.7 km considering the null hypothesis of dispersion. In addition, we obtained an altitudinal shift average of 325.5 m and 275.1 m, considering the two hypotheses, respectively (Table 1). For all future periods, the distribution center showed a shift toward the southwest (Figures 2b-d).

### 3. Coverage by the National System of Conservation Units (SNUC)

The predicted area with favorable conditions for the presence of the Gray-backed Tachuri reached 27 to 42 reserves (current and future scenarios). Those reserves protect a total area that ranges from 7,248 km<sup>2</sup> to 9,635 km<sup>2</sup>, depending on the specific scenario (Table 2). Nevertheless, only 49% of this area presented suitable environments for the species' occurrence (Table 2) and represented only 11.7% of the predicted current range, from an expected range of at least, 44.2% (Table 1).

Similarly, in future periods, it was observed the loss of favorable conditions for up to eight reserves, considering the null hypothesis of dispersion, and the addition of up to 17 new reserves, if we consider the hypothesis of total dispersion (Table 2). Among the reserves that had favorable current conditions (27), 19 remained, at least partially, favorable in all future scenarios. Among the 13 reserves predicted by the model that have records for the species, 12 partially remained suitable conditions for all future periods (Table 3). The majority of the reserves showed a decrease in their areas of potential occurrence for the species in relation to the current scenario. Most affected areas in future scenarios are located in the northern Minas Gerais and Bahia, as is the case of the Morro do Chapéu State Park (Bahia), Biribiri State Park and Sempre Vivas National Park (Minas Gerais), which showed extinction conditions for 2080 (Table 3).

## Discussion

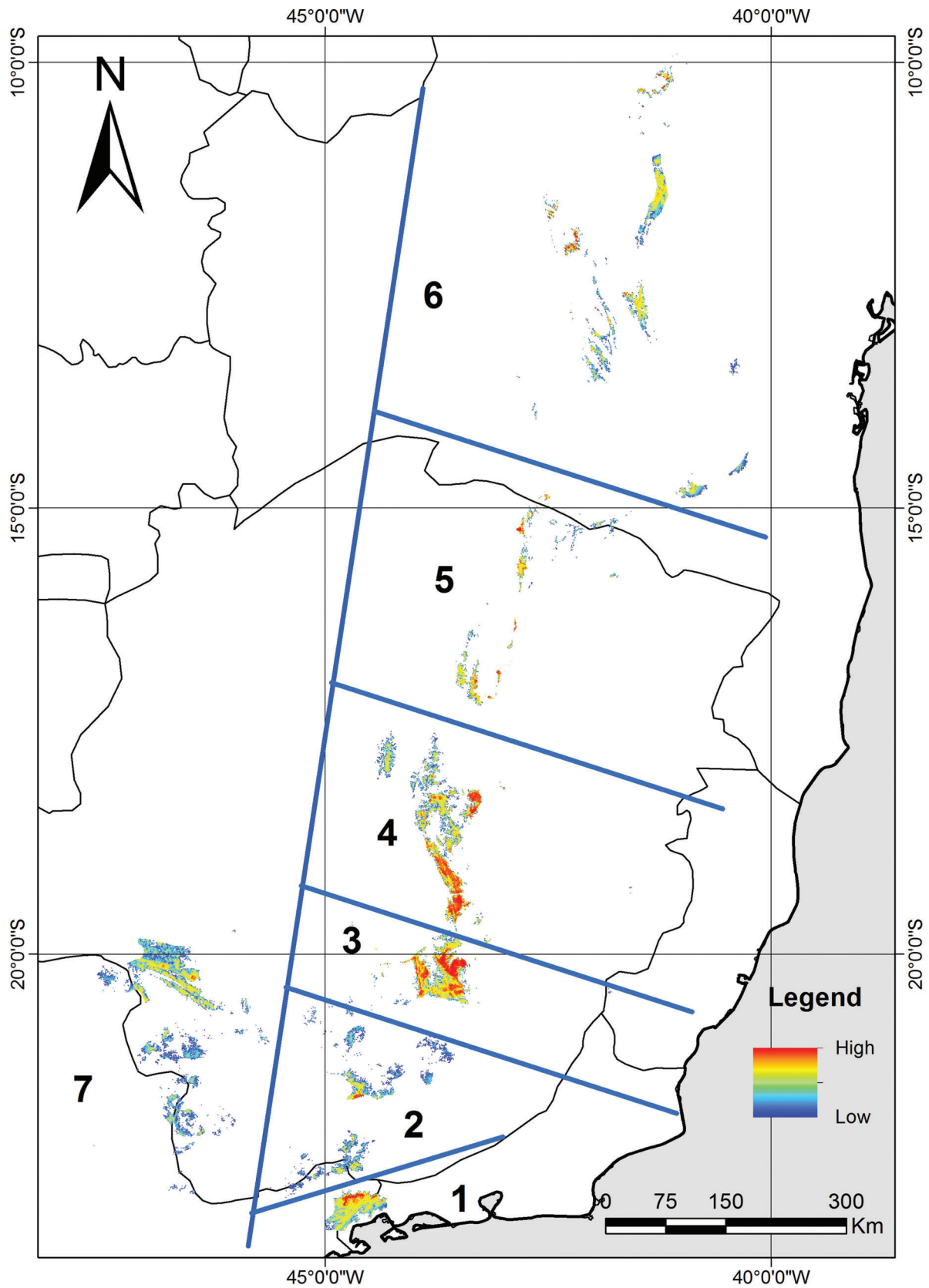
### 1. Current predicted distribution

Recent records support our projections of current suitable range for the Gray-backed Tachuri, considering that these new records made on Caparaó National Park (Costa 2011) and Serra dos Órgãos National Park (Rennó 2012) were not used in the models, and were predicted by these (Figures 2b-d, fine and isolated areas).

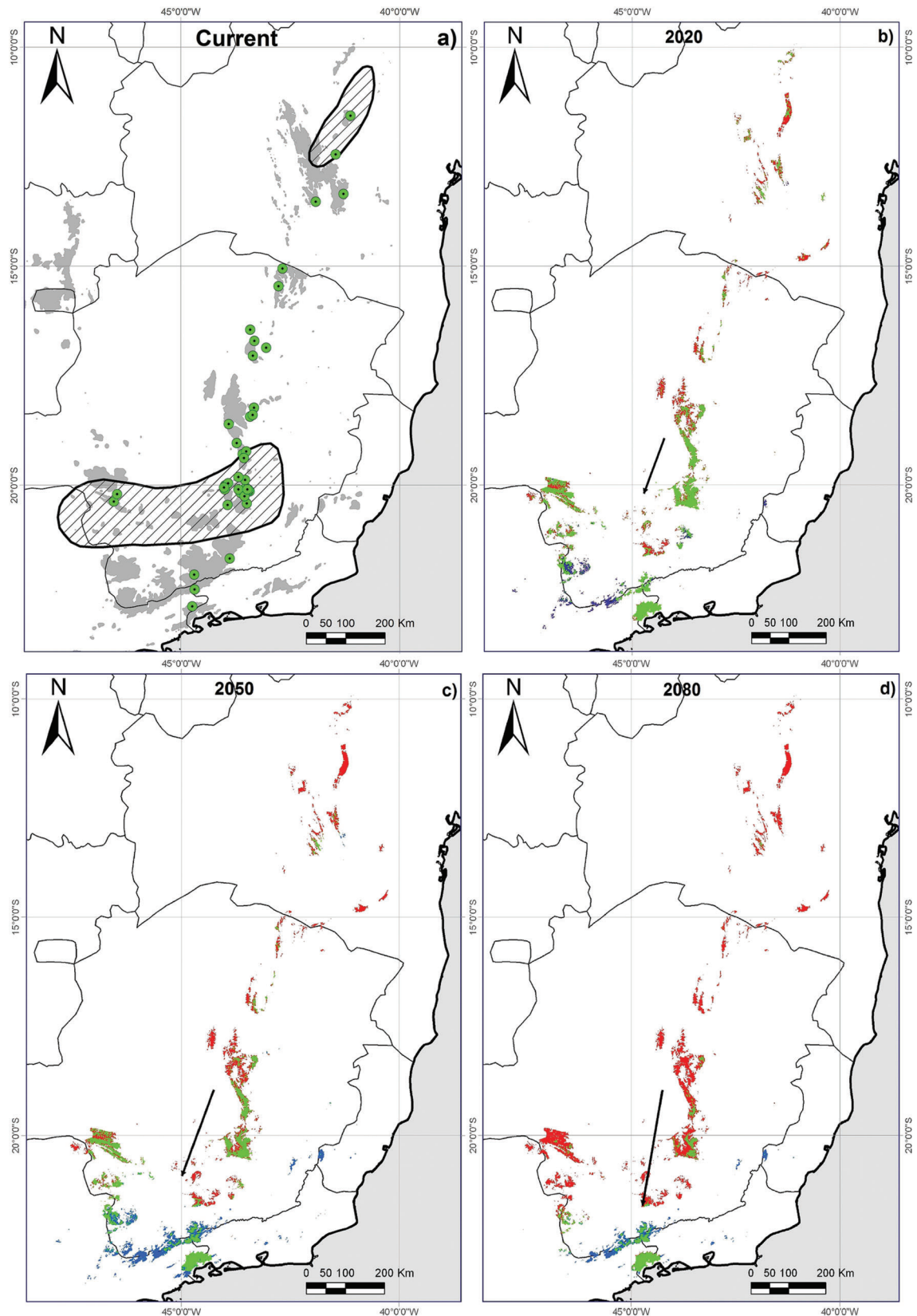
The predicted current suitable range for the Gray-backed Tachuri showed to be 75% lower than that presented by the literature (Ridgely & Tudor 2009) (Figure 2a). Probably the range of this species will prove to be even lower if vegetation is considered. This is because some sites where the soil is deeper and more fertile are covered by forest patches, and because deforestation is a dynamic process that can negatively affect the extent of projected occurrence for future scenarios. This difference may be related to the methods employed by these authors, which are similar to the minimum convex polygon (Odum & Kuenzler 1955), that consists in the union of all points of occurrence in a polygon, not taking into account the habitat preferences of species. Thus, many areas with unfavorable conditions may be considered favorable. Both hatched areas presented in Figure 2a can be explained because there were only two isolated sets of points at the time of delimitation of the specie's range, which suggested two disjunct populations. Nevertheless, recent surveys have shown that this species is distributed throughout the entire Espinhaço Range, from its southern portion 'Iron Quadrangle' to its northern end, known as the Chapada Diamantina, and also in the Serra da Canastra, to the west, and part of Serra da Mantiqueira and Serra do Mar Ranges, to the south (Vasconcelos et al. 2003, Vasconcelos & D'Angelo Neto 2007, Vasconcelos 2008, Vasconcelos & Rodrigues 2010). Costa et al. (2010) reported a genetic isolation between specimens of two areas that in current predicted range are isolated by physical barriers corresponding to lowland areas. In this context and based on the current predicted range we were able to identify seven possible populations may be isolated in the eastern Brazilian highlands by lowland barriers (Figure 1). Therefore, further phylogeographic studies and information on the ability of transposition of unfavorable areas by the species are necessary to assess if there is interruption of gene flow by these other possible barriers.



Distribution and protection of *P. superciliaris*



**Figure 1.** Current predicted distribution of the Gray-backed Tachuri (*Polystictus superciliaris*). The color scale represents the probability of favorable conditions within each pixel above the threshold of sensitivity-specificity, from lower (light blue) to highest (red). White areas are below the favorable conditions. Lines represent possible geographical natural barriers to the dispersal of the species and probably dividing isolated subpopulations (numbers).



**Figure 2.** (a) Range of the Gray-backed Tachuri (*Polystictus superciliaris*) (shaded area) following Ridgely & Tudor (2009) and real distribution based on museum specimens, literature and fieldwork (points); altitudes above 1,000 m are represented in gray; (b-d) area with constant favorable conditions (current and future periods) for the presence of the Gray-backed Tachuri (green), potential area for future occupation/dispersal but not currently (blue), current area with the possibility of loss of favorable conditions (red) in future periods (2020, 2050 and 2080, respectively) and shift of distribution center in relation to the current distribution (black arrows).

**Table 1.** Estimation of the distribution of the Gray-backed Tachuri (*Polystictus superciliaris*) for current and future periods; predicted area currently impacted; expected level of effective protection in reserves; variation in size and loss of area; direction and displacement of the area centroid and altitudinal shift in future periods in relation to the present, under the assumptions of the hypothesis of total dispersion and the null hypothesis of dispersion.

	Projected range (km <sup>2</sup> )	Used area within range (km <sup>2</sup> )	Protected area expected (%)	Protected area in all reserves (%)	Loss of range (%)	Range shift (km)	Altitudinal shift (m)	Range shift direction
				With dispersion				
Current	30,745	5,505	44.2	11.7	*	*	*	*
2020	22,247	4,178	49.4	13.1	27.6	138	126	Southwest
2050	23,837	3,937	48.3	13.8	22.5	216	186	Southwest
2080	10,837	851	61.2	17.0	64.8	275	326	Southwest
				Without dispersion				
Current	30,745	5,505	44.2	11.7	*	*	*	*
2020	19,036	3,268	52.0	13.8	38.1	103	110	Southwest
2050	15,537	2,589	55.3	15.5	49.5	170	136	Southwest
2080	6,986	649	68.3	17.0	77.3	251	275	Southwest

**Table 2.** Number of reserves with high suitability for the occurrence of the Gray-backed Tachuri (*Polystictus superciliaris*) at present and in future periods, considering the variation in the extent of the area, the total area of reserves and percentage of area represented.

	Present	With dispersion			Without dispersion		
		2020	2050	2080	2020	2050	2080
Number of reserves	27	30	42	30	26	25	19
Reserves with decrease of range	0	16	16	10	16	16	10
Reserves with increase of range	0	5	7	7	5	7	7
Reserves that have lost conditions	0	1	2	8	1	2	8
Reserves added	0	4	17	11	0	0	0
Reserves in which the range remain constant	0	5	2	2	5	2	2
Reserves areas (km <sup>2</sup> )	7,248	7,558	9,635	7,494	7,230	6,828	5,216
% of suitable areas in the reserves	49%	39%	34%	25%	26%	24%	12%
Representation in the reserves (km <sup>2</sup> )	3,586	2,924	3,292	1,839	2,628	2,403	1,188

## 2. Projected distribution in future scenarios

Climatic change may be an important driver for range reduction of several species (IPCC 2007a). The conversion of natural vegetation into antropic areas is a dynamic process along time (see Silva et al. 2006), being responsible for the loss of approximately 1/5 of the current range projected. Even with the extinction of these activities, our results predicted a decrease of 38.1-77.3% of the current species range.

The center of distribution of the Gray-backed Tachuri showed a gradual latitudinal shift southwestward. It also presented an altitudinal shift. These results showed that this species responds similarly to other montane species (e.g., Bussche et al. 2008, Popy et al. 2010), which tend to occupy higher areas and/or to move to higher latitudes, with temperature increase. Parmesan & Yohe (2003) considered that each 1 km of latitudinal displacement in the range would be equivalent to the displacement of 1 m in altitude. The model of distribution of the Gray-backed Tachuri partially showed this pattern, or was close to it. The shift of the species' distribution Center toward southwest coincides with this predictable hypothesis, since displacement vectors were directed to the highland areas of Espinhaço and Mantiqueira Ranges (Figure 2b-d).

In a recent gap analysis, Marini et al. (2009a) found a larger area for the current distribution (55,000 km<sup>2</sup>) of the Gray-backed

Tachuri in comparison to that observed in the present study (30,745 km<sup>2</sup>). Similar discrepancies can also be observed for the extent of the impacted area (13000/5505 km<sup>2</sup>) and available range (420000/25240 km<sup>2</sup>). This difference may be related to the fact that Marini et al. (2009a) have used a larger number of algorithms and/or due a coarser resolution (of 4.5 x 4.5 km<sup>2</sup>), which covers more unfavorable areas that were considered in the present analysis. Despite these methodological differences, a decrease in the range of the Gray-backed Tachuri in future periods is clear in both studies, as the representation of the species range in protected areas, is lower than expected. The response of the species to future climate change will depend on the availability of shrubby environments where it feeds and nests, and from the shrubby vegetation response by climate change (Hoffmann et al. 2007, Hoffmann & Rodrigues 2011) and/or on its ability to adapt to new environmental conditions.

## 3. Conservation and protection by the Brazilian system of conservation units (SNUC)

Among all reserves where the Gray-backed Tachuri has been recorded (n = 14) (M.F. Vasconcelos, unpublished data), 85% (n = 12) were predicted by the model. Other 15 reserves were predicted to have favorable conditions in the present and in future periods and they present shrubby vegetation that is

Table 3. Reserves with favorable conditions predicted for the occurrence of *Polystictus superciliosus* today and in future periods.

Reserve	State <sup>a</sup>	Municipality	Biome <sup>b</sup>	Area (km <sup>2</sup> )	Present	2020	2050	2080	1 <sup>c</sup>	2 <sup>d</sup>
Córrego dos Fehos Ecological Station	MG	Nova Lima	Ce/Ma	4.74	x	x	x	x	-	-
Cercadinho Ecological Station	MG	Belo Horizonte	Ce/Ma	1.94	x	x	x	-	-	x
Bananal Ecological Station	SP	Bananal	Ma	7.87	x	x	x	x	-	x
Tripuí Ecological Station	MG	Ouro Preto	Ce/Ma	3.85	x	x	x	x	-	-
Serra da Piedade Natural Monument	MG	Sabarará, Caeté	Ce/Ma	16.78	x	x	x	x	-	x
Biribiri State Park	MG	Diamantina	Ce	148.40	x	x	x	-	-	x
Campos do Jordão State Park	SP	Campos do Jordão	Ma	71.49	x	x	x	x	-	-
Furnas do Bom Jesus State Park	SP	Pedregulho	Ce	18.57	x	-	-	-	-	-
Pedra Azul State Park	ES	Domingos Martins	Ma	8.44	-	-	x	-	-	x
Forno Grande State Park	ES	Castelo	Ma	6.85	-	-	x	x	-	x
Desengano State Park	RJ	Campos de Goytacazes, Santa Maria Madalena, São Fidélis	Ma	187.00	-	-	x	-	-	x
Itambé State Park	MG	Santo Antônio do Itambé, Serro	Ce/Ma	55.76	x	x	x	x	-	-
Pico Marumbi State Park	PR	Morretes	Ma	21.43	-	-	x	x	-	-
Três Picos State Park	RJ	Cachoeiras de Macacu, Teresópolis	Ma	406.24	-	-	x	x	-	x
Grão-Mogol State Park	MG	Grão Mogol	Ca/Ce	294.97	x	x	x	-	-	x
Ibitipoca State Park	MG	Lima Duarte, Santa Rita	Ce/Ma	13.19	-	x	x	x	-	-
Itacolomi State Park	MG	Mariana e Ouro Preto	Ce/Ma	65.11	x	x	x	x	-	-
Mananciais de Campos do Jordão State Park	SP	Campos do Jordão	Ma	182.60	-	x	x	x	-	-
Morro do Chapéu State Park	BA	Morro do Chapéu	Ca	401.79	x	x	-	-	-	x
Rio Pardo State Park	MG	Rio Pardo de Minas	Ca/Ce/ Ma	106.70	x	x	x	-	-	-
Rio Preto State Park	MG	São Gonçalo do Rio Preto	Ce	103.97	x	x	x	x	-	-
Serra do Brigadeiro State Park	MG	Araçonga, Divino, Ervália, Fervedouro	Ce/Ma	130.17	-	x	x	x	-	x
Serra do Mar - N. Cunha-Indaia State Park	SP	Cunha	Ma	66.17	-	-	x	-	-	-
Serra do Mar - N. Picinguaba State Park	SP	Ubatuba	Ma	453.20	-	-	x	x	-	-
Serra do Mar - N. Santa Virgínia State Park	SP	Natividade da Serra, São Luís do Paraitinga, Cunha	Ma	143.88	-	-	x	-	-	-
Serra do Papagaio State Park	MG	Baependi, Aiuruoca, Itamonte	Ma	200.75	x	x	x	x	-	-
Serra do Rola Moça State Park	MG	Belo Horizonte	Ce/Ma	34.81	x	x	x	x	-	-
Serra Negra State Park	MG	Itamarandiba	Ce/Ma	119.26	-	-	x	-	-	x
Chapada Diamantina National Park	BA	Mucugê, Andaraí, Lençóis, Palmeiras, Ibicoara	Ca/Ce/ Ma	1246.70	x	x	x	x	-	-
Serra da Bocaina National Park	SP-RJ	Parati, Angra dos Reis, Ubatuba	Ma	862.19	x	x	x	x	-	-
Serra da Canastra National Park	MG	São Roque de Minas, Delfinópolis	Ce/Ma	1711.05	x	x	x	x	-	-
Serra do Cipó National Park	MG	Jaboticatubas, Morro do Pilar	Ce/Ma	272.07	x	x	x	x	-	-
Serra dos Órgãos National Park	RJ	Petrópolis, Guapimirim, Magé, Teresópolis	Ma	92.01	-	-	x	x	-	x
Sempre-Vivas National Park	MG	Bocaitiva, Olhos-D'água, Diamantina, Buenópolis	Ce	1058.06	x	x	x	-	-	x
Itatiaia National Park	RJ-MG	Itatiaia, Resende, Bocaina de Minas, Itamonte	Ma	247.63	x	x	x	x	-	-

Continued on next page



Table 3. Continued.

Reserve	State <sup>a</sup>	Municipality	Biome <sup>b</sup>	Area (km <sup>2</sup> )	Present	2020	2050	2080	1 <sup>c</sup>	2 <sup>d</sup>
Caparaó National Park	ES-MG	Ibitirama, Alto Caparaó	Ma	274.94	x	x	x	x	-	x
Brejo Grande Municipal Natural Park	MG	Paraisópolis	Ma	2.09	-	x	x	x	-	-
Ecológico M. Sagui da Serra Municipal Natural Park	MG	Manhumirim	Ma	3.76	-	-	x	-	-	-
Mangabeiras Municipal Natural Park	MG	Belo Horizonte	Ce/Ma	2.12	x	x	x	-	-	x
Ribeirão do Campo Municipal Natural Park	MG	Conceição do Mato Dentro	Ce/Ma	26.52	x	x	x	x	-	x
Araras Biological Reserve	RJ	Petrópolis, Miguel Pereira	Ma	18.70	-	-	x	x	-	-
Serra dos Toledos Biological Reserve	MG	Itajubá	Ma	9.38	x	x	x	x	-	-
Serra Pedra do Coração Biological Reserve	MG	Caldas	Ma	0.12	x	x	x	x	-	-

<sup>a</sup>Biome: Am = Amazônia, Ca = Caatinga (dry forest), Ce = Cerrado e Ma = Atlantic Forest; <sup>b</sup>State: Ba = Bahia; ES = Espírito Santo; MG = Minas Gerais; PR = Paraná; RJ = Rio de Janeiro and SP = São Paulo <sup>c</sup>Vasconcelos, 2009; <sup>d</sup>Phytophysiognomies with presence of shrub and rocky outcrops.

favorable to the species (M.F. Vasconcelos, unpublished data) (Table 3). However, the species has not been recorded in these locations because of the lack of avifaunal surveys, or even because of its inability to colonize these areas through lowland barriers (Chaves et al. 2014).

Currently, the species is found in reserves with more than 1,000 km<sup>2</sup>, such as the Serra da Canastra and the Chapada Diamantina National Parks (1,711 km<sup>2</sup> and 1,246 km<sup>2</sup>, respectively), but also in small reserves, with less than 20 km<sup>2</sup>, as in the case of Ibitipoca State Park and Serra da Piedade Natural Monument (13 km<sup>2</sup> and 16 km<sup>2</sup>, respectively). However, small reserves may not have enough area to maintain viable populations of the Gray-backed Tachuri, estimated as about 200 km<sup>2</sup> (Hoffmann 2011). Disregarding the reserve size, the Gray-backed Tachuri appeared to be less abundant in the north of its range (M.F. Vasconcelos, unpublished data), which includes the Chapada Diamantina National Park. Thus, one of the largest reserves with predicted occurrence has a low efficiency for the species' conservation.

The representation of the Gray-backed Tachuri in reserves is only one-quarter of the expected for the present and tends to decrease gradually in future periods due to range reduction. The percentage of the predicted range represented in the reserves, for considering the species protected, increased gradually in future periods. Thus, there was a decrease of the total predicted range and, consequently, the need for increasing the species' representation for, at least, 61.2 and 68.3%, considering the hypothesis of total dispersion and the null hypothesis of dispersion, respectively (Table 1). The decrease in the species' representation indicates that the reserves that currently present suitable habitat will become inefficient and inadequate for the species' conservation under future scenarios of environmental changes.

The distribution models as those generated by Maxent, could act as important tools for understanding many aspects of current and future predicted suitable ranges for the Gray-backed Tachuri, providing important information to re-evaluate its conservation status and ideas for possible conservation plans, as well as subsidizing important questions about the species, yet to be investigated. An example would be the potential barriers that probably isolate populations, in order to demonstrate the possibility of occupation of new areas in future. Knowledge on the species' ability to cross these potential barriers would be an important indicator of the possibility of colonization of new areas in the face of possible climatic changes, considering that there are no reports on the viability of translocation and reintroduction programs for small and exclusively insectivorous tyrant-flycatchers (Tyrannidae). The region where the environmental suitability for the Gray-backed Tachuri remained constant in future periods, with a great area extension, is located in the 'Iron Quadrangle' region, Minas Gerais. This region has been strongly impacted by iron mining with more than 50 iron-ore opencast mines that totalize c. 2000 ha (DNPM - Departamento Nacional de Produção Mineral 2006, Jacobi et al. 2007), which have caused high destruction of the 'campos rupestres' vegetation.

In order to review the species' conservation status it should be considered, for the estimation of its range, restrictive characteristics of its localities of occurrence (e.g. altitudinal limit and climatic variables), and not only the delineation of an area that incorporates all records of occurrence. The range of the Gray-backed Tachuri used in the assessment of its

conservation status in 2008 was 104,000 km<sup>2</sup> and tripled (314,000 km<sup>2</sup>) for the reevaluation in 2011, lowering the species' status from 'near-threatened' to 'least concern' (BirdLife International 2014). The present results contradict this new categorization, indicating that the conservation status of many species, mainly based on their ranges, may be influenced by factors not considered limiting of their occurrence. Thus, we consider that the species should remain as 'near-threatened' and, if the impacts of climatic changes really occur, the species should be included in the 'vulnerable' category according to the BI criterion (<20,000 km<sup>2</sup>) of IUCN (2001).

The species had a projected range on the political boundaries of five Brazilian states: São Paulo, Minas Gerais, Rio de Janeiro, Espírito Santo and Bahia. Nevertheless, it is not endangered in any regional threatened category in the first four states that present regional red lists (Bergallo et al. 2000, Biodiversitas 2007, Passamani & Mendes 2007, Bressan et al. 2009). The northern range of the species is located in Bahia, a state that does not have any red list. Furthermore, the fact that the species is not represented on a list of threatened species at a regional level should be viewed with caution, since the species was recorded in the states of Rio de Janeiro and Espírito Santo after the publication of the red lists of those states. In both states there is evidence that populations of Gray-backed Tachuri are very small and present restricted range. On the other hand, in the state of São Paulo there are historical records of the species and an absence of recent records. Thus, the extent of occurrence of the species presented in this paper is also important for the assessment and reassessment of the conservation status of the species at regional levels.

Finally, we suggest investing in research on autoecology and natural history of this and other species that are virtually unknown in the Neotropics. This information may help improve the models prediction and are essential for the conservation of the Gray-backed Tachuri and probably the entire highland biota from eastern Brazil.

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