





Research Article
Animal Genetics

Genetic structure of the endangered Irrawaddy dolphin (*Orcaella brevirostris*) in the Gulf of Thailand

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Abstract

The Irrawaddy dolphin (*Orcaella brevirostris*) is an endangered, small cetacean species which is widely distributed in rivers, estuaries, and coastal waters throughout the tropical and subtropical Indo-Pacific. Despite the extensive distribution of this species, little is known of individual movements or genetic exchange among regions in Thailand. Here, we evaluate the genetic diversity and genetic structure of *O. brevirostris* in the eastern, northern and western Gulf of Thailand, and Andaman Sea. Although phylogenetic relationships and network analysis based on 15 haplotypes obtained from 32 individuals reveal no obvious divergence, significant genetic differentiation in mitochondrial DNA (overall $F_{ST} = 0.226$, $P < 0.001$; $\Phi_{ST} = 0.252$, $P < 0.001$) is apparent among regions. Of 18 tested microsatellite loci, 10 are polymorphic and successfully characterized in 28 individuals, revealing significant genetic differentiation (overall $F_{ST} = 0.077$, $P < 0.05$) among the four sampling sites. Structure analysis reveals two inferred genetic clusters. Additionally, Mantel analysis demonstrates individual-by-individual genetic distances and geographic distances follow an isolation-by-distance model. We speculate that the significant genetic structure of *O. brevirostris* in Thailand is associated with a combination of geographical distribution patterns, environmental and anthropogenic factors, and local adaptations.

Keywords: Mitochondrial DNA, microsatellite, cross-amplification, genetic differentiation, cetacean.

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Introduction

The Irrawaddy dolphin (*Orcaella brevirostris*) is a small cetacean species which is widely distributed in rivers, estuaries, and coastal waters throughout the tropical and subtropical Indo-Pacific (Stacey and Arnold, 1999; Minton *et al.*, 2013). Estuarine and coastal populations of this species occur from the northwestern Bay of Bengal, east through the Gulf of Thailand to the Philippines, and south to the Indonesian Archipelago (Stacey and Arnold, 1999; Krützen *et al.*, 2018). Riverine subpopulations occur in three large rivers including Ayeyarwady in Myanmar, Mahakam in Indonesia, and Mekong in Cambodia and southern Lao People's Democratic Republic (Smith and Hobbs, 2002; Krebs, 2004; Beasley *et al.*, 2013). Other subpopulations occur in lagoons and marine appended lakes, such as Chilika in India, Songkhla in Thailand, and Malampaya Sound in the Philippines (Beasley *et al.*, 2002; Dolar *et al.*, 2002; Sutaria and Marsh, 2011).

Different habitat preferences are observed between the *O. brevirostris* populations in fresh waters and coastal waters.

In rivers, *O. brevirostris* occur in deeper waters (10–50 m) at the confluences of two rivers, or above and below rapids, while in estuaries and coastal waters they frequent shallower depths (generally < 10 m) within a few kilometers of shore (Dolar *et al.*, 2002; Smith and Hobbs, 2002; Smith *et al.*, 2006; Minton *et al.*, 2011; 2013; Jackson-Ricketts *et al.*, 2018). Because of their patchy and fragmented distribution in rivers and coastal waters, Irrawaddy dolphins are particularly vulnerable to disturbance (Minton *et al.*, 2017). This species is currently classified as Endangered by the International Union for Conservation of Nature (IUCN) in their Red List, and is listed in Appendix I of the Convention on International Trade in Endangered Species of Wild Flora and Fauna (CITES). Because of their small population sizes and declining number, and increasing anthropogenic threats, the three riverine subpopulations (Ayeyarwady, Mahakam, Mekong) and two lacustrine subpopulations (Songkhla, Malampaya) are classified as Critically Endangered by the IUCN (Minton *et al.*, 2017).

In Thailand's waters, *O. brevirostris* occur along the Andaman Sea coast, Songkhla Lake, and many locations along the coast surrounding the Gulf of Thailand (Chantrapornsyl *et al.*, 1996; 1999; Anderson and Kinze, 1999; Beasley *et al.*, 2002; Smith *et al.*, 2003; Kittiwattanawong *et al.*, 2007; Hines *et al.*, 2015). Recently, field observations and acoustic

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studies targeting *O. brevirostris* were conducted at locations such as Trat Bay, Bangpakong Estuary, and Donsak, in the eastern, northern, and western Gulf of Thailand, respectively (Tongnunui *et al.*, 2011; Hines *et al.*, 2015; Jutapruet *et al.*, 2017; Niu *et al.*, 2019; Jackson-Ricketts *et al.*, 2020). Additionally, significantly different isotope values in *O. brevirostris* teeth, indicating strong geographic variation and potential subpopulation structure in the Gulf of Thailand and Andaman Sea coast, have been reported (Jackson-Ricketts *et al.*, 2018). Despite this, little is known about individual movements and genetic exchange of *O. brevirostris* among regions in Thailand.

Understanding genetic diversity and the genetic structure of populations is important to conserve and manage species, especially for inshore dolphins whose isolated populations are highly affected by human activities (Mace and Lande, 1991; Jefferson *et al.*, 2009; Brown *et al.*, 2014). Mitochondrial DNA (mtDNA) has been used as a genetic marker for reconstructing patterns of population demography, admixture, biogeography and speciation, because of its high mutation rate, maternal inheritance and lack of recombination, and high intracellular copy number (Wilson *et al.*, 1985; Hurst and Jiggins, 2005). Recently, there are a few mtDNA-based studies on the genetic diversity and population structure of *O. brevirostris* throughout its distribution range. The mtDNA evidence shows that the Chilika population in India does not share any haplotypes with those of Thailand, Cambodia and Indonesia populations (Jayasankar *et al.*, 2011). Another molecular genetic-related study suggests a long-standing isolation of the Mekong dolphin population from other Orcaella populations, with the remaining population now experiencing low genetic diversity (Krützen *et al.*, 2018). Moreover, significant levels of genetic differentiation are found among three *O. brevirostris* populations, Chilika Lagoon in India, the eastern Gulf of Thailand, and the Mekong River in Cambodia, indicating strong genetic differentiation between coastal and riverine populations and among different geographic locations (Caballero *et al.*, 2019).

Microsatellite markers are short tandem repeats of 1–6 nucleotides that are widespread in the nuclear genomes of most taxa. Due to their high mutation rates, microsatellites are also considered to be robust and informative DNA markers for solving ecological and molecular genetic-related questions, such as those pertaining to bottlenecks, kinship, population structure, and migration (Selkoe and Toonen, 2006; De Barba *et al.*, 2017). There are few known microsatellite sequences for *O. brevirostris*, although several molecular genetic-related studies have been undertaken on this species (Jayasankar *et al.*, 2011; Krützen *et al.*, 2018; Caballero *et al.*, 2019; Kundu *et al.*, 2019). Fortunately, microsatellites with highly conserved flanking regions often allow cross-species amplification from congeneric or confamilial taxa, especially for vertebrates, including mammals (Selkoe and Toonen, 2006; Barbara *et al.*, 2007).

Our research aims were to: (1) assess transferability and polymorphism of microsatellite markers for *O. brevirostris* from closely related dolphin species; (2) evaluate genetic diversity of *O. brevirostris* using samples obtained from throughout the Gulf of Thailand and Andaman Sea coast using mitochondrial control region sequences and nuclear

microsatellite markers; and (3) estimate levels of genetic differentiation among these *O. brevirostris* individuals in Thailand for both marker types.

Material and Methods

Sampling

All fieldwork was under permits from the Ministry of Agriculture and Rural Affairs of China, and with approval from the Department of Marine and Coastal Resources of Thailand. The relevant CITES Permits (No. 2018CN/IC000475/CH) were obtained for import of samples. There was no issue on ethics in this study.

Sample collection and DNA extraction

From 2010 to 2018, muscle or skin samples were collected from 37 dead (stranded) *O. brevirostris* individuals recovered from nine sites in Thailand: Chachengsao (n = 1), Chonburi (n = 1), Chumporn (n = 2), Mueang Trat (n = 11), Phetchaburi (n = 1), Samut Sakorn (n = 9), Satun (n = 1), Surattani (n = 7), and Trang (n = 4). Geographic coordinates for all except one individual from Phetchaburi were recorded at the time of collection. Samples were grouped into four general regions as described in Jackson-Ricketts *et al.* (2018), each region encompassing specific sites: (1) Mueang Trat, representing the eastern Gulf of Thailand (EG, n = 11), (2) Chonburi, Chachengsao, Samut Sakorn, and Phetchaburi, making up the northern Gulf (NG, n = 12), (3) Chumporn and Surattani, making up the western Gulf (WG, n = 9), and (4) Trang and Satun, representing Andaman Sea sites (AS, n = 5) (Figure 1). Genomic DNA from minced tissue samples was extracted using DNeasy blood and tissue extraction kits (QIAGEN, Valencia, USA) according to the manufacturer instructions.

Mitochondrial DNA amplification

A fragment of the mitochondrial control region was amplified using forward primer Ce-CRF 5'-GAATCCCCGGTCTTGTAACC-3' and reverse primer Ce-CRR 5'-TCTCGAGATTTTCAGTGTCTTGCTTT-3' (Hoelzel *et al.*, 1991). Total PCR volume of (50 µL) contained approximately 50 ng of genomic DNA, 1×Easy Taq Mix (Takara, Dalian, China), 1 µL of each forward and reverse primer, and ddH₂O to make up the final volume. The PCR profile comprised an initial denaturation step at 94 °C for 3 min, 35 cycles of 94 °C for 30 s, annealing at 50 °C for 1 min, extension at 72 °C for 1 min, and a final extension at 72 °C for 10 min. All PCR products were sequenced in both directions on an ABI 3730 automated DNA sequencer (Applied Biosystems).

Microsatellites crossamplification

We selected a total of 18 microsatellite markers (Table S1, data unpublished), initially isolated from the Indo-Pacific humpback dolphin *Sousa chinensis*, for this study. These loci were all 'perfect-type' and long tandem repeat motifs (e.g., tetra- or pentanucleotide), and demonstrated high polymorphism in Thailand *S. chinensis* populations (data unpublished). Microsatellites were allocated into 6 multiplex

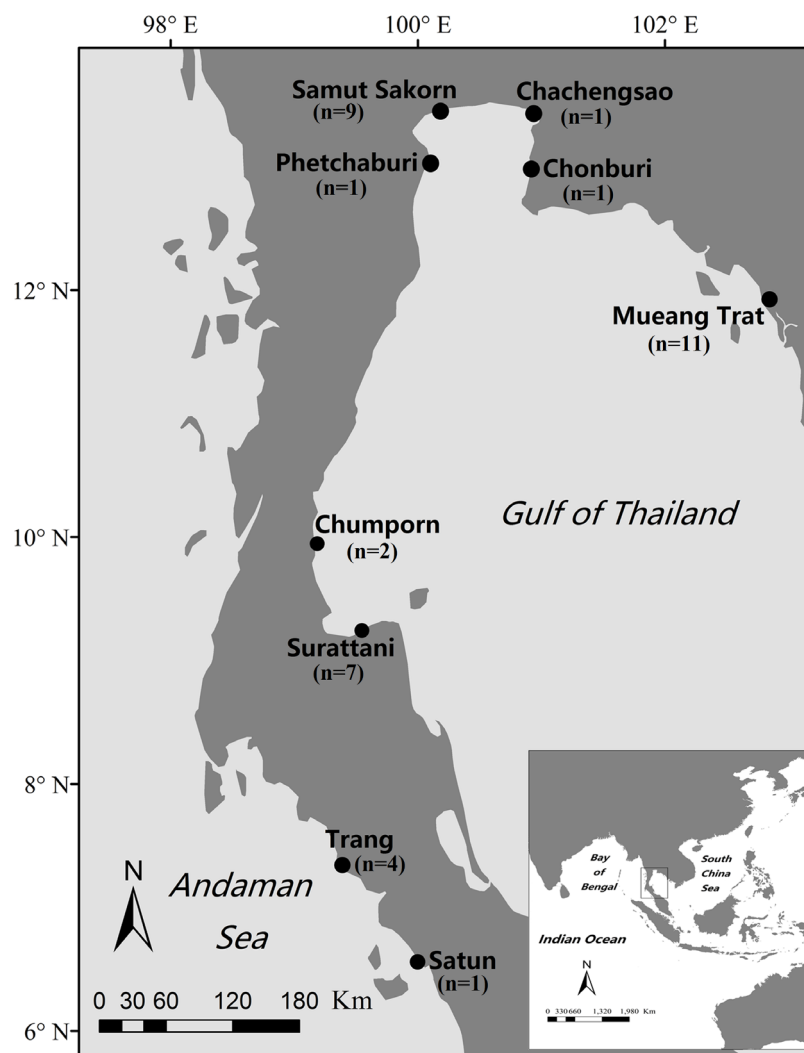


Figure 1 – Locations of sample collection in the Gulf of Thailand and Andaman Sea sites. Mueang Trat represents the eastern Gulf of Thailand (EG); Chonburi, Chachengsao, Samut Sakorn, and Phetchaburi make up the northern Gulf (NG); Chumporn and Surattani make up the western Gulf (WG); Trang and Satun represent the Andaman Sea site (AS).

PCR panels using software MPprimer (Shen *et al.*, 2010), based on annealing temperature, complementarity of primer pairs, and allele size range. The 5' end of each forward primer was labeled with a fluorescent dye (6-FAM, VIC or NED). The total PCR volume (20 μ L) consisted of approximately 50 ng of genomic DNA, 1 \times Multiplex PCR Kit (Takara, Dalian, China), the optimal dosage (Table S1) of each forward and reverse primer, and ddH₂O added to make up the final volume. PCR conditions involved an initial denaturation step at 94 °C for 3 min, followed by 32 cycles of 94 °C for 30 s, the specific annealing temperature (Table S1) for 90 s, extension at 72 °C for 60 s, and a final extension for 30 min at 60 °C. PCR products were run on an ABI 3730XL automated DNA sequencer (Applied Biosystems) using GeneScan LIZ 500 as the internal size standard. Allele sizes were automatically scored with GeneMapper version 4.1 (Applied Biosystems) and manually checked.

Detecting genotyping errors

Genomic DNA obtained from dead individuals may be low in quantity and/or quality, be at increased risk of

contamination, and more susceptible to genotyping errors such as allele dropouts and false alleles (Taberlet *et al.*, 1996; Bonin *et al.*, 2004; Pompanon *et al.*, 2005). Therefore, in each multiple PCR batch we used one individual as a positive control for genotyping each locus separately to ensure consistent amplification of alleles. A negative control with no DNA template was also used in each PCR batch to detect possible of contamination during PCR amplification. Moreover, we randomly retested genotypes from four individuals (> 10 %) to estimate genotyping error rates for across all 18 loci.

Statistical analyses

Mitochondrial DNA data analysis

Raw mtDNA sequences were aligned using the Clustal-W algorithm implemented in the program MEGA version 5.0 (Tamura *et al.*, 2011), then manually checked and edited. MEGA was also used to calculate nucleotide composition, and variable and conserved sites. Diversity measures including numbers of haplotypes (H), specific haplotypes (SH), and nucleotide (Hd) and haplotype diversities (π) for each

region, were analyzed using DnaSP version 5 (Librado and Rozas, 2009). Phylogenetic relationships among haplotypes were constructed using the neighbor-joining (NJ) method implemented with 1000 bootstrap replicates in MEGA. A haplotype network was constructed using the software PopART (Leigh and Bryant, 2015). We also used Arlequin version 3.0 (Excoffier *et al.*, 2005) to investigate genetic variation among regions by calculating both F_{ST} and Φ_{ST} values with 1000 random permutations.

Microsatellite data analysis

GenAIEx version 6.501 (Peakall and Smouse 2006, 2012) was used to estimate the number of alleles per locus (Na), effective number of alleles (Ne), observed heterozygosity (Ho), expected heterozygosity (He), unbiased expected heterozygosity (uHe), Shannon's information index (I), and Fixation Index (F). Genepop version 4.0.7 (Rousset, 2008) was used to test departure from Hardy-Weinberg equilibrium (HWE) and linkage disequilibrium (LD) among all pairs of loci. Myriads version 1.1 (Carvajal-Rodriguez, 2018) was performed to correct the p-value-based multiple testing by Bonferroni sequential correction procedures. To check for signs of genetic bottlenecks, which cause an heterozygosity excess in populations, BOTTLENECK version 1.2.02 (Cornuet and Luikart, 1996; Piry *et al.*, 1999) was used under the infinite allele model (IAM), stepwise mutation model (SMM), and two-phased model (TPM), based on 1000 iterations. We used the Wilcoxon sign-rank test to estimate significance because only 10 microsatellite loci were included in analysis (Wilcoxon, 1945). Micro-Checker version 2.2.3 (Van Oosterhout *et al.*, 2004) was used to detect occurrences of null alleles, allele dropout, or scoring error for each locus, with 95 % confidence intervals.

Genetic structure and number of genetic clusters (K) were determined using Structure 2.3.4 (Pritchard *et al.*, 2000) based on genotyping data generated from the 6 multiplex PCR panels in this study. For all analyses, the length of the burn-in period was set to 10^5 iterations, followed by 10^6 in the number of MCMC repetitions. Then, the Admixture model was used with correlated allele frequencies. The Locprior model was chosen to infer possible weak population structure with the assistance of sample group information (Hubisz *et al.*, 2009). The number of inferred K was set between 1 and 6, and 20 independent replicates were run for each K value. Subsequently, the web program Structure Harvester version 0.6.94 (Earl and VonHoldt, 2012) was used to calculate the Delta K value and determine the best number of K clusters (Evanno *et al.*, 2005). We used CLUMPP version 1.1.2 (Jakobsson and Rosenberg, 2007) to summarize the optimal alignment of the 20 replicates for the same K value. Final results were displayed graphically with Distruct version 1.1 (Rosenberg, 2004).

FSTAT version 2.9.3.2 (Goudet, 1995, update in Feb. 2002) was used to estimate genetic differentiation F_{ST} (Weir and Cockerham, 1984) among the four geographic regions based on 1000 permutations. A principal coordinate analysis (PCoA) was performed in GenAIEx based on the standardized covariance of the individual-by-individual genetic distance matrix (Peakall *et al.*, 1995; Smouse and Peakall, 1999). In addition, Mantel analysis (Mantel, 1967) was used in GenAIEx

to test isolation by distance (IBD) by performing the correlation between matrices of individual-by-individual genetic distances and geographic distances measured as the distance between two individuals calculated from sampling location coordinates. We conducted a second IBD analysis excluding the genetic and geographical data for AS samples because of the large distance between AS and other regions. Both tests were run with 999 random permutations in GenAIEx.

Data availability

The 18 tested microsatellite sequences and 32 obtained mtDNA control region sequences (including 15 mtDNA haplotypes) reported in this paper have been deposited into the GenBank database under the accession numbers of MK766845–MK766870 (Table S1) and MT738330–MT738361 (Table S2), respectively.

Results

Available genetic data for analysis

No contamination was detected during PCRs, and no genotyping errors were observed when randomly retesting four individuals. Among the 18 loci tested, two (Sch5878 and Sch5685) were excluded because of unsuccessful amplification, and six (Sch7427, Sch193, Sch4657, Sch5373, Sch974, Sch5094) were excluded because they were monomorphic. The remaining 10 loci were polymorphic in the four sampling locations.

Due to the low quality of some DNA samples, we obtained mtDNA sequences for 32 individuals (Table S2) only, including two from AS, eight from WG, 10 from EG, and 12 from NG. For microsatellite data analysis, only DNA samples for which at least eight of the 10 polymorphic loci were successfully genotyped were included. Therefore, nine samples (two from EG, three from AS, and four from WG) were removed because of poor amplification success. Finally, we generated genotypes of 10 microsatellites for 28 individuals only, including two individuals from AS, five from WG, nine from EG, and 12 from NG. Allele frequency distributions in the four geographic regions based on 10 polymorphic loci are graphically represented in Figure 2.

Genetic diversity

Based on mtDNA data, 15 haplotypes were found in the 32 individuals by a consensus 883 bp sequences from the four geographic locations (Figure 3). Only one haplotype was shared between NG and WG. Genetic diversity estimates of H, SH, Hd, and π values for each region are summarized in Table 1. NG showed the most haplotypes ($n = 6$), while AS exhibited the least ($n = 2$). Estimated Hd was high (mean value = 0.925 ± 0.025), but observed π was relatively low (mean 0.009 ± 0.001). Although AS exhibited the highest Hd and π values among the four geographic regions, only 2 individuals were included in genetic analysis.

Based on microsatellite DNA data, genetic diversity estimates of Na, Ne, Ho, He, uHe, I and F values for each sampling site are summarized in Table 2. Average Na and Ne were highest in NG, with values of 4.000 ± 0.683 and 2.581 ± 0.519 , respectively. The highest mean He and I were in EG,

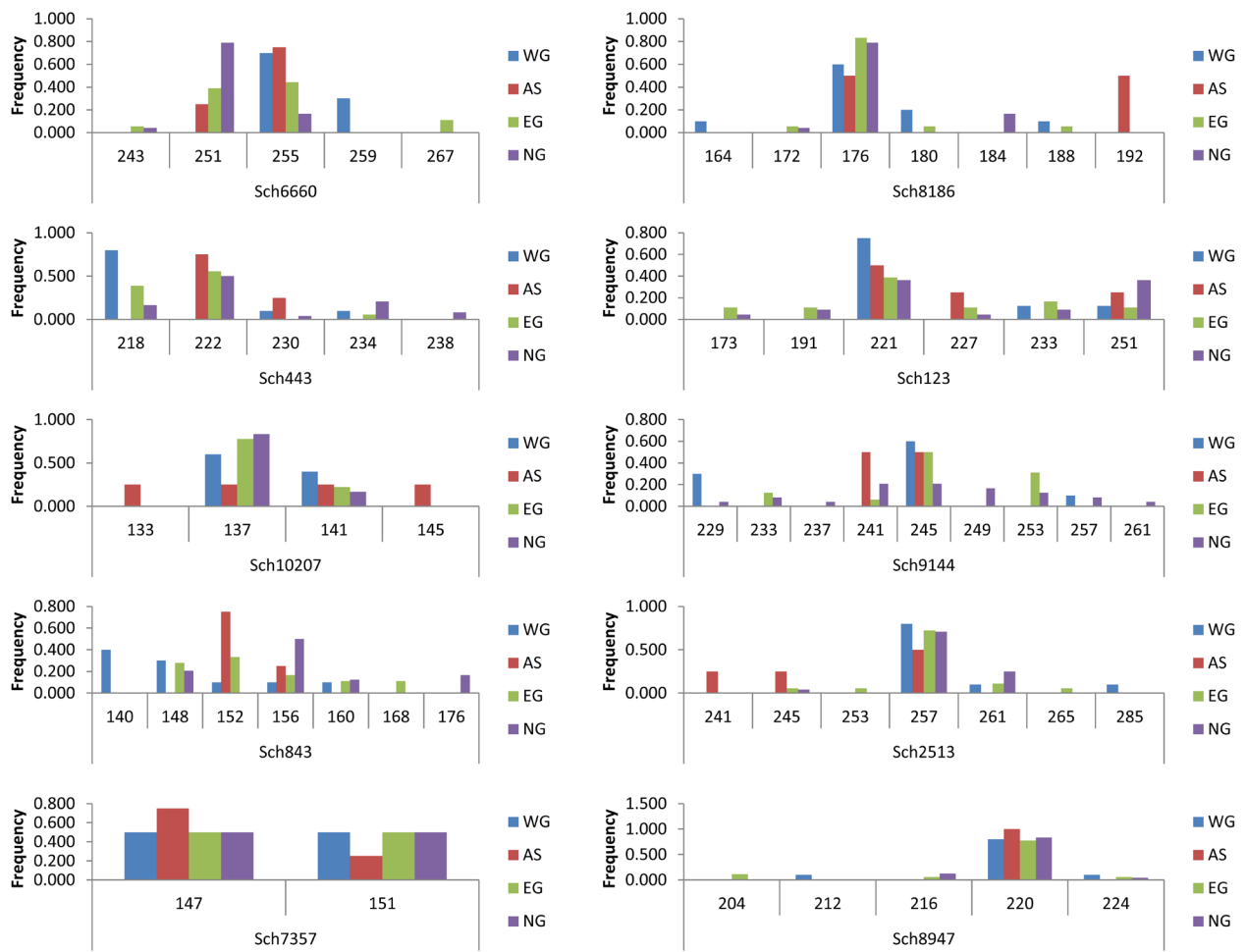


Figure 2 – Allele frequencies with graphs by each sampling location and each locus for microsatellite data. Allele frequencies are displayed for 10 polymorphic microsatellites, and different colors represent different regions in Thailand.

Haplotype	Variable sites (883bp)															Geographic location																		
	1	1	1	2	2	2	3	3	3	3	3	3	4	5	5	5	5	6	6	7	EG	WG	NG	AS										
Hap 1	C	C	G	C	T	A	T	T	T	T	C	C	G	A	T	T	C	C	C	G	T	A	C	G	A	C	T	-	3	1	-			
Hap 2	G	.	-	-	6	-		
Hap 3	.	.	C	T	-	-	1	-		
Hap 4	2	-	-	-		
Hap 5	G	.	.	C	C	.	T	T	T	-	3	-	-		
Hap 6	G	G	T	A	T	2	-	-	-
Hap 7	T	G	G	T	A	T	-	-	-	1
Hap 8	G	.	.	C	G	T	A	T	-	1	-	-
Hap 9	G	C	T	A	T	-	-	2	-
Hap 10	G	C	T	A	T	-	-	1	-
Hap 11	G	T	A	T	1	-	-	-
Hap 12	.	.	T	.	T	T	A	T	-	-	-	1
Hap 13	G	T	A	T	-	-	1	-
Hap 14	C	G	T	A	T	5	-	-	-
Hap 15	C	G	T	A	T	-	1	-	-
Total sample size																										10	8	12	2					

Figure 3 – Polymorphic sites of mtDNA haplotypes identified among the four geographic regions in Thailand. Nucleotide position and geographic location are at the top, respectively.

with values 0.531 ± 0.053 and 0.986 ± 0.117 , respectively (Figure 4). Although AS had the lowest I (0.710 ± 0.119) and He (0.450 ± 0.065), it had the highest Ho (0.600 ± 0.125) and uHe (0.633 ± 0.096). One locus significantly deviated from HWE ($P < 0.05$) in each of WG, EG, and NG (Table S3). Except for AS (because the sample size was too small to analyze), the other three locations had no significant bottleneck under any model of mutation (Wilcoxon test: $P > 0.05$). No significant LD was found in any pairs of the 10 tested loci after

Bonferroni sequential correction. No evidence of large allele dropout or scoring errors was detected by Micro-Checker.

ANOVA results revealed Na values differed significantly from each other among sampled regions (ANOVA: $F_{3,36} = 3.279$, $P = 0.032$), but Ne values did not ($F_{3,36} = 0.657$, $P = 0.584$). There were no significant differences among He ($F_{3,36} = 0.442$, $P = 0.725$) and uHe ($F_{3,36} = 0.559$, $P = 0.646$) values. Additionally, no significant differences were found among Ho ($F_{3,36} = 0.213$, $P = 0.887$) and I ($F_{3,36} = 1.151$, $P = 0.342$) values from any of the four geographic regions.

Table 1 – Information and molecular indices for *O. brevirostris* based on mtDNA control region sequences.

Population	Number	H	SH	Hd \pm SD	$\pi \pm$ SD
WG	8	4	3	0.786 +/- 0.113	0.009 +/- 0.002
AS	2	2	2	1.000 +/- 0.500	0.014 +/- 0.007
EG	10	4	4	0.733 +/- 0.120	0.009 +/- 0.001
NG	12	6	5	0.758 +/- 0.122	0.005 +/- 0.001
Total	32	15	14	0.925 +/- 0.025	0.009 +/- 0.001

Note: H is the number of haplotypes, SH is the number of specific haplotypes, Hd is the haplotype diversity, π is the nucleotide diversity.

Table 2 – Genetic diversity parameters of *O. brevirostris* in the four geographic regions in Thailand.

Sampling region		N	Na	Ne	I	Ho	He	uHe	F
WG	Mean	4.900	3.000	2.000	0.804	0.530	0.467	0.520	-0.102
(n = 5)	S.E.	0.100	0.298	0.198	0.083	0.084	0.039	0.043	0.120
AS	Mean	1.900	2.300	2.073	0.710	0.600	0.450	0.633	-0.319
(n = 2)	S.E.	0.100	0.260	0.268	0.119	0.125	0.065	0.096	0.172
EG	Mean	8.900	3.900	2.456	0.986	0.507	0.531	0.563	0.064
(n = 9)	S.E.	0.100	0.407	0.334	0.117	0.066	0.053	0.056	0.070
NG	Mean	11.900	4.000	2.581	0.966	0.522	0.509	0.531	-0.026
(n = 12)	S.E.	0.100	0.683	0.519	0.162	0.071	0.064	0.067	0.050

Note: N is the number of sample size, Na is the number of alleles, Ne is the number of effective alleles, I is the Shannon's information index, Ho is the observed heterozygosity, He is the expected heterozygosity, uHe is the unbiased expected heterozygosity, F is the fixation index, S.E. is the standard error.

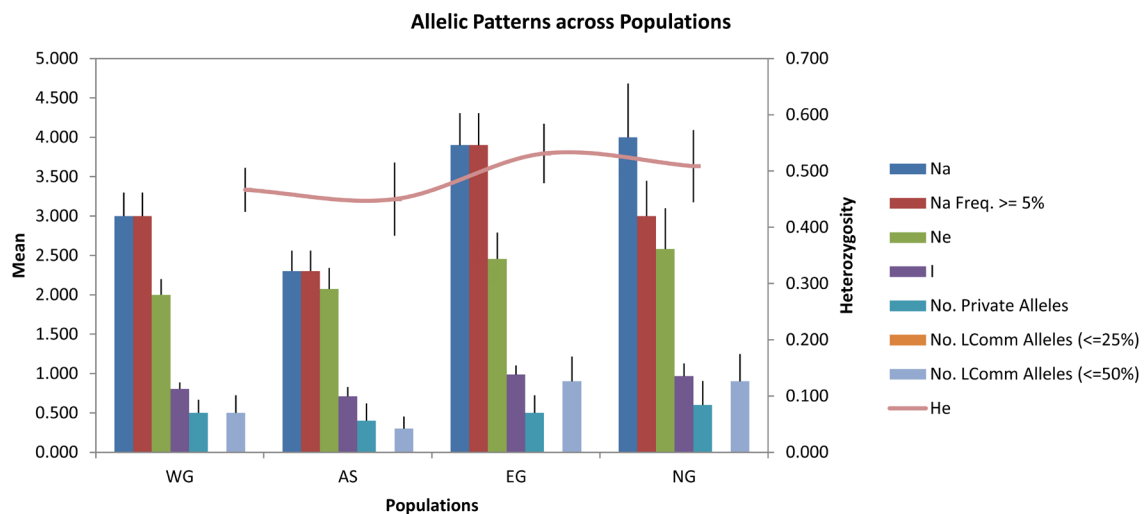


Figure 4 – Allelic patterns across regions in the geographic distribution of *O. brevirostris* in Thailand. Na is the number of different alleles, Na Freq. is the number of common alleles with a frequency $\geq 5\%$, Ne is the effective number of alleles, I is the Shannon's information index, No. LComm Alleles are the number of locally common alleles (Freq. $\geq 5\%$) found in 50% or 25% or fewer populations, and He represented by the curve is the expected heterozygosity.

Genetic differentiation

The NJ tree was constructed based on the 15 haplotypes of *O. brevirostris* and the two outgroup mtDNA sequences from *Orcinus orca* and *Steno bredanensis*. Haplotypes from different regions were randomly distributed on the NJ tree, with no phylogenetic structure corresponding to geography apparent (Figure 5A). The network also showed similar results with the NJ tree, without obvious divergence of haplotypes from different sites. Levels of divergence among most haplotypes were low. However, divergence from Hap_14 in EG and Hap_15 in WG to other haplotypes was extremely high. Both Hap_7 and Hap_12 (defined by two AS individuals) were also highly divergent from other haplotypes (Figure 5B).

AMOVA results revealed significant high genetic differentiation in mtDNA (overall $F_{ST} = 0.226$, $P < 0.001$; $\Phi_{ST} = 0.252$, $P < 0.001$) among Thailand *O. brevirostris* individuals. Pairwise F_{ST} and Φ_{ST} values revealed moderate to high levels

of genetic differentiation between different region pairs, but some cases showed no significant differentiation between AS and other region pairs (Table 3). For microsatellites, the estimated F_{ST} value revealed low but significant genetic differentiation (overall $F_{ST} = 0.077$, $P < 0.05$) among the four geographic regions. Low to moderate levels of genetic differentiation were detected between different region pairs. Estimates of pairwise F_{ST} are presented in Table 4. Pairwise differentiations were all significant ($P < 0.05$), except for the pair of EG and AS, which had the lowest genetic differentiation (pairwise $F_{ST} = 0.008$, $P > 0.05$). Some pairwise F_{ST} values were moderately differentiated, with the highest genetic differentiation found between WG and NG (pairwise $F_{ST} = 0.166$, $P < 0.01$). AMOVA results for the degree of variance in *O. brevirostris* individuals are summarized in Table 5. There was 7.706 % genetic variance among the four geographic regions, 6.676 % variance among individuals within region, and 85.618 % variance within individuals.

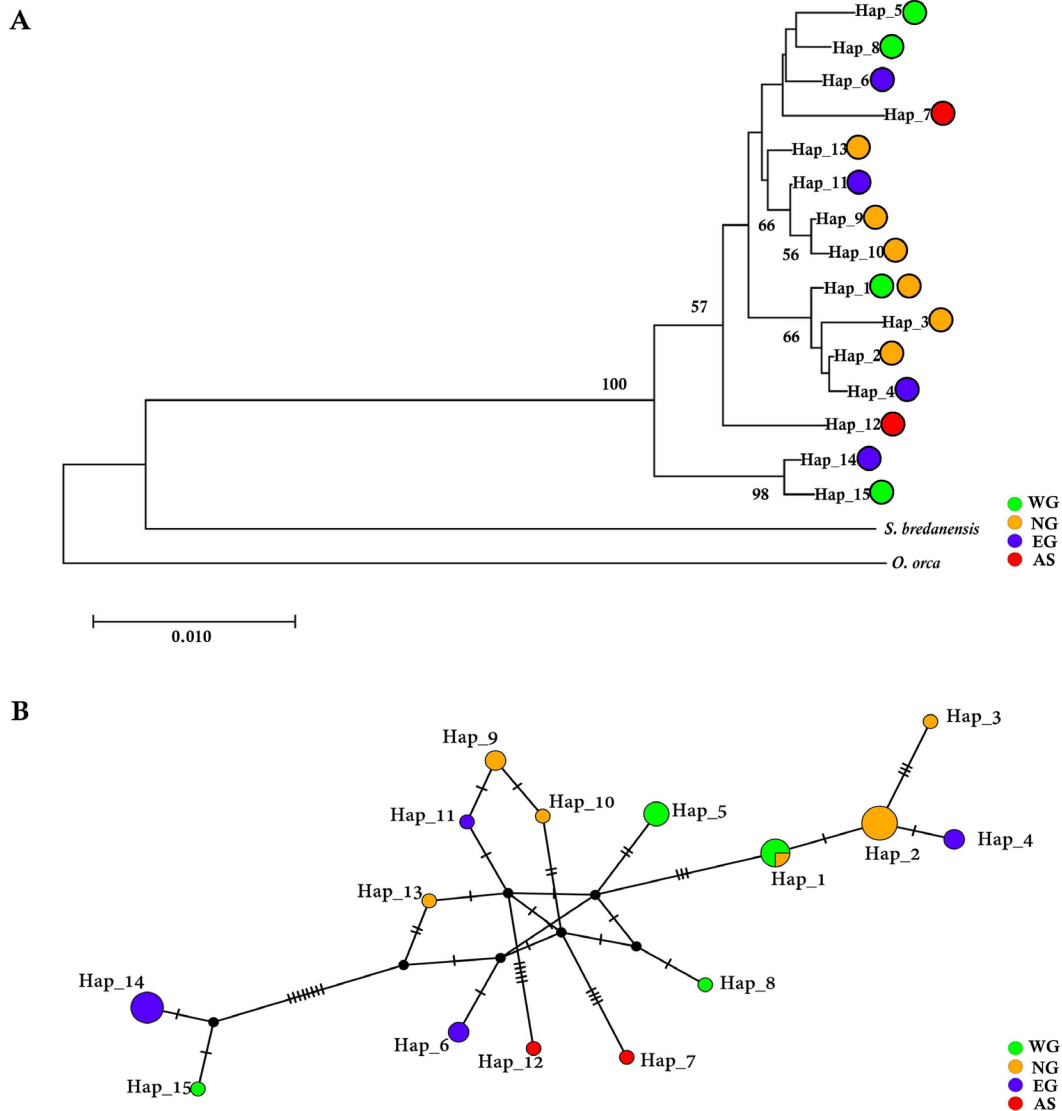


Figure 5 – Genetic structures of all 15 mtDNA haplotypes of *O. brevirostris* in Thailand. (A) Neighbour-joining tree of mtDNA haplotypes. *Orcinus orca* and *Steno bredanensis* are used as the outgroup species. Only bootstraps values $> 50\%$ are shown. (B) Network of mtDNA haplotypes. Size of the circles represents the frequency of haplotype, and dash lines represent the number of mutational steps between haplotypes. Colors indicate the sampling location(s) for each haplotype.

PCoA plots of individuals based on the standardized covariance of genetic distance matrix revealed that the large genetic distance was detected between individuals from WG and NG, indicating the high level of genetic differentiation.

Axis coordinates 1 and 2 accounted for 17.901 % and 12.988 % of the total variance, respectively (Figure 6).

Structure Harvester analysis revealed there were most likely only two clusters (Delta K = 2.0, Figure 7). The modal

Table 3 – Matrix of pairwise Φ_{ST} (above diagonal) and F_{ST} (below diagonal) values among the four geographic regions based on mtDNA control region sequences.

Φ_{ST} F_{ST}	WG	AS	EG	NG
WG	–	0.118	0.197*	0.160*
AS	0.158	–	0.291*	0.363*
EG	0.242**	0.202	–	0.348**
NG	0.205**	0.184	0.254**	–

Note: * P < 0.05, ** P < 0.01.

Table 4 – Pairwise F_{ST} estimates among the four geographic regions based on microsatellites.

Population	WG	AS	EG	NG
WG	–			
AS	0.121*	–		
EG	0.060*	0.008	–	
NG	0.166**	0.101*	0.034*	–

Note: * P < 0.05, ** P < 0.01.

Table 5 – Analysis of molecular variance of *O. brevis* in the four sampled regions.

Source of variation	df	SS	MS	Est. Var.	%
Among populations	3	17.679	5.893	0.231	7.706
Among individuals with population	24	71.339	2.972	0.201	6.676
Within individuals	28	72.000	2.571	2.571	85.618
Total	55	161.018		3.003	100

Note: df is the degrees of freedom, SS is the sums of squares, MS is the mean squares, Est. Var. is the estimated variance within and among populations.

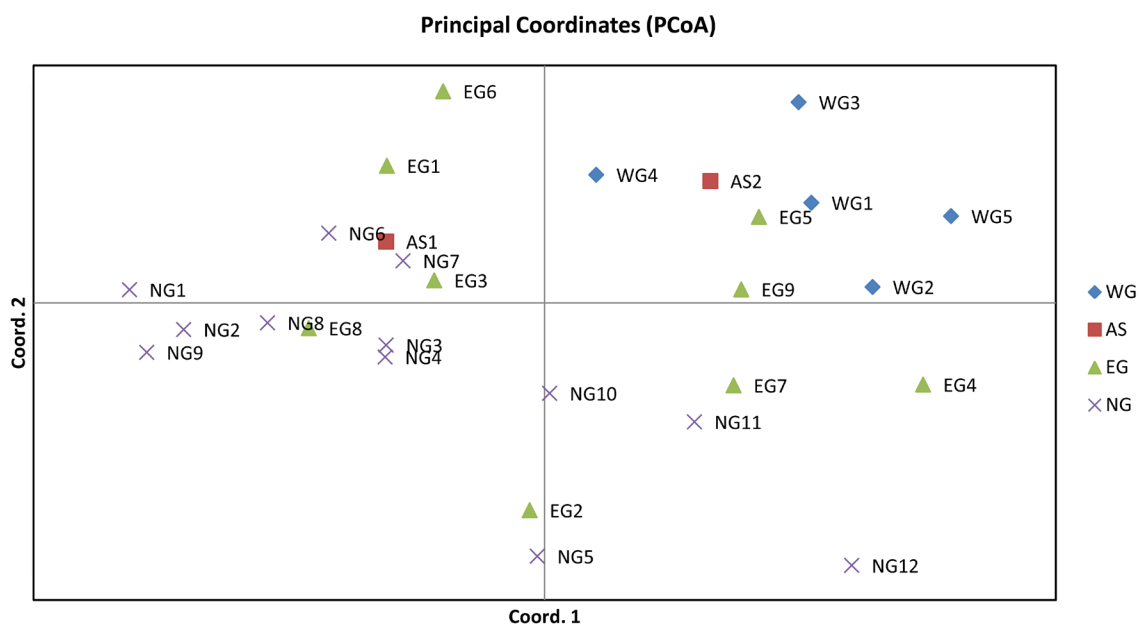


Figure 6 – Principal coordinate analysis plots of individuals based on the standardized covariance of genetic distance matrix. Different shapes and colors represent different geographic regions in Thailand. Axis coordinates 1 and 2 account for 17.901 % and 12.988 % of the total variance, respectively.

value of Delta K can best show the real number of groups, based on the second order rate of change with respect to K of the likelihood function (Evanno *et al.*, 2005). However, the graphical output by Distruct indicated that all sampled individuals exhibited admixture in the possible genetic clusters,

with K ranging 2 to 5 (Figure 8). Even at K = 2, a weak genetic structure for the two inferred genetic clusters remained apparent. Cluster I was predominant in three locations (WG, AS, and EG), with assignment probabilities were 74.5 %, 66.9 %, and 57.9 %, respectively. Cluster II was predominant in NG,

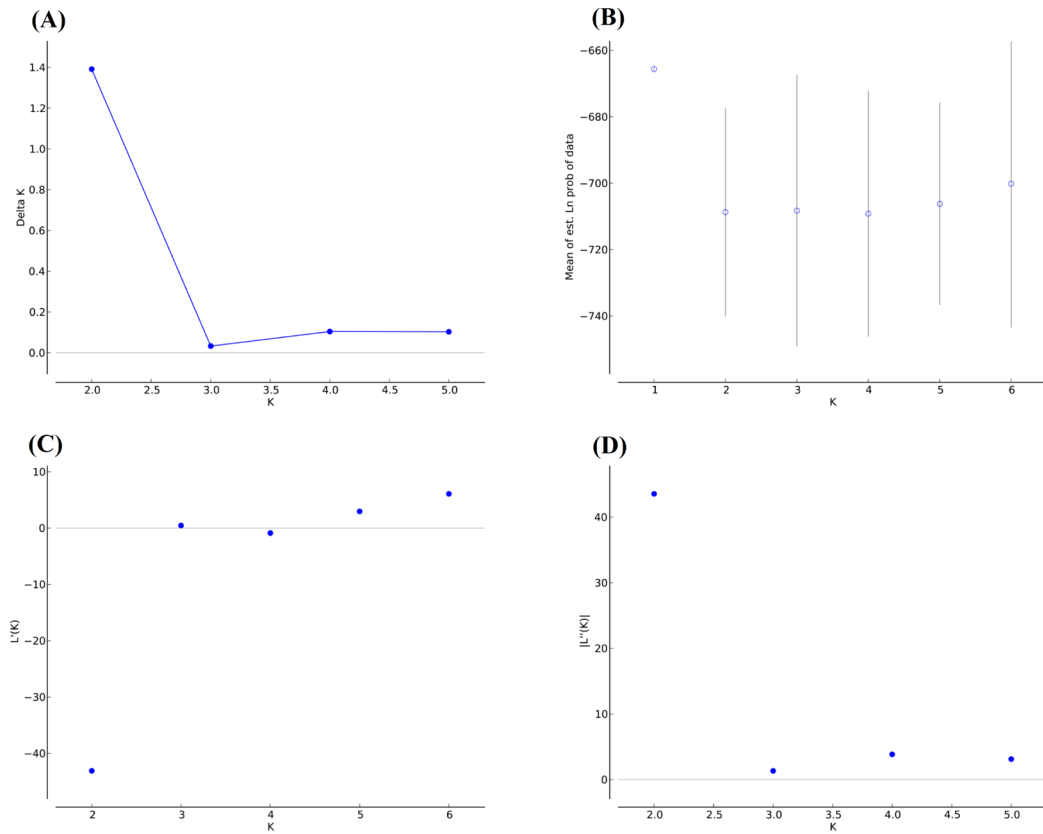


Figure 7 – Optimal K value determined by Structure Harvester online program. (A) Delta K; (B) L(K) (mean ± SD); (C) Rate of change of the likelihood distribution; and (D) Absolute value of the 2nd order rate of change of the likelihood distribution.

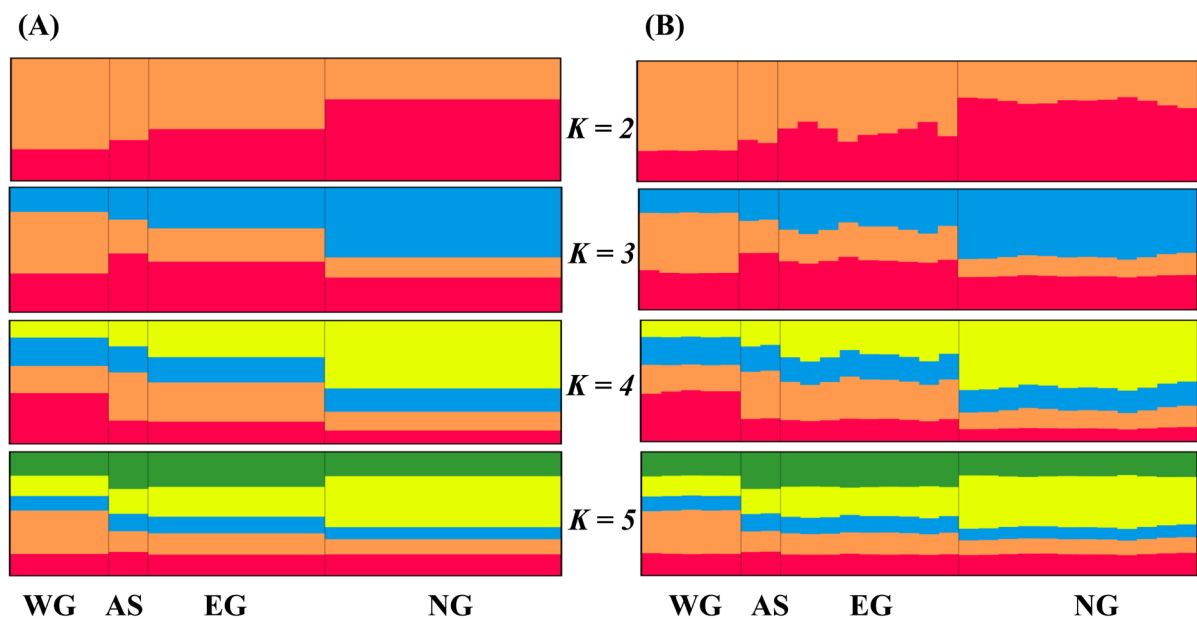


Figure 8 – Results of clustering (K varied from 2 to 5) calculated by Structure program based on (A) the population Q-matrix, and (B) the individual Q-matrix. Black vertical lines separate the four geographic regions in Thailand, and different colors represent the possible genetic clusters.

with an assignment probability of 66.7 %. We also tried to identify possible partitions within cluster I and cluster II, but none was apparent.

Mantel tests revealed a positive and weak correlation ($r^2 = 0.0699$, $P = 0.004$) between the individual-by-individual

genetic distances and geographic distances, indicating a pattern of IBD among the four geographic regions (Figure 9A). The result remained positive and significant when geographic coordinates and microsatellite data for two AS individuals were removed ($r^2 = 0.0665$, $P = 0.001$; Figure 9B).

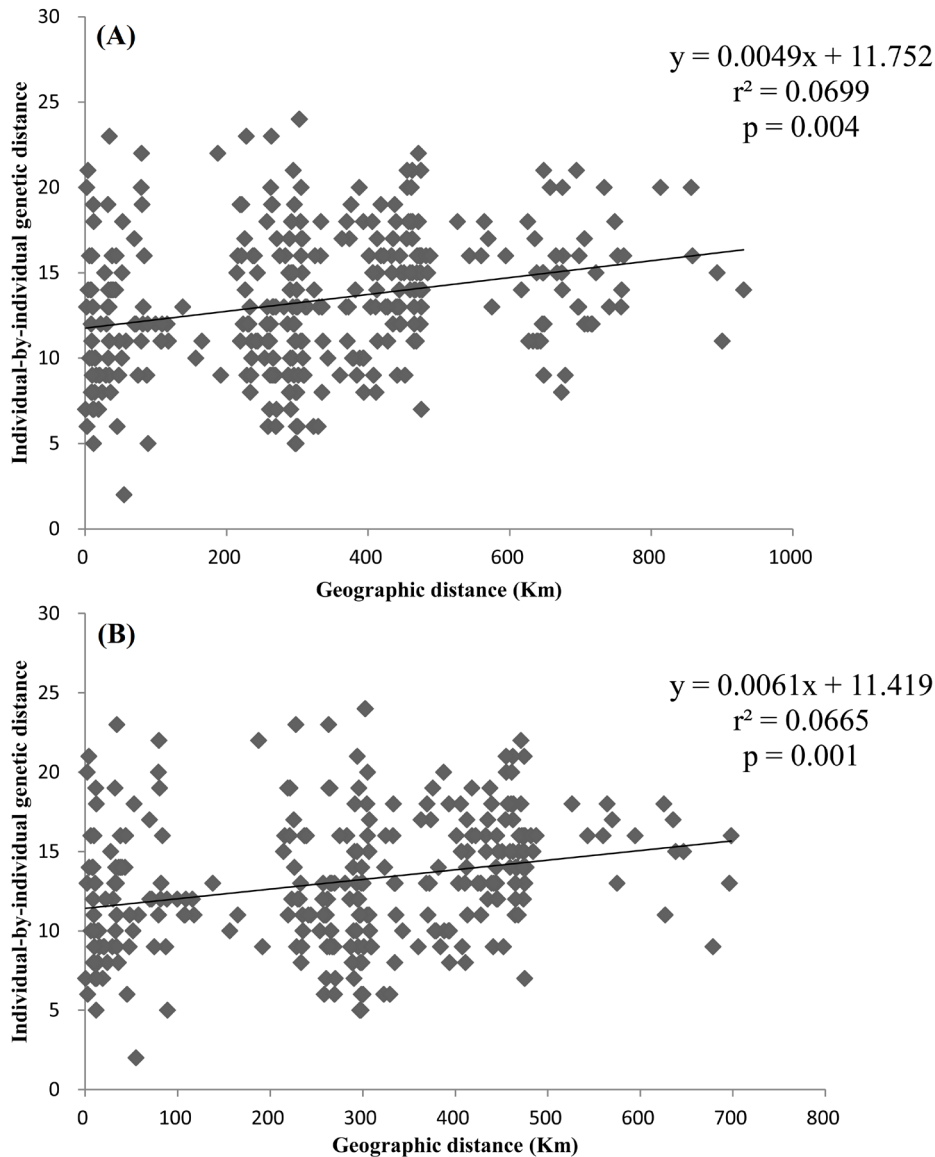


Figure 9 – Isolation by distance plots using individual-by-individual genetic distances and geographic distances (km) among (A) the four (WG, AS, EG, and NG), and (B) the three (WG, EG, and NG) sampling sites in Thailand. Geographic and microsatellite data of 27 individuals were included in analysis because of the missing location coordinates of one individual in Phetchaburi (NG).

Discussion

Some molecular markers with highly conserved primers have strong versatility during PCR. They can be invariant within species and sometimes even across broad taxonomic groups (Selkoe and Toonen, 2006). Microsatellites with highly conserved primer regions may be transferred from closely related species, but they rarely work across broad taxonomic groups (Glenn and Schable, 2005). Ten loci cross-amplified from other cetaceans (including *Tursiops aduncus*, *T. truncatus*, and *O. orca*) proved polymorphic in Mekong River *O. brevirostris* (Krützen *et al.*, 2018). We

report 10 of 18 tested loci isolated from *S. chinensis* to be polymorphic among *O. brevirostris* individuals from the four geographic locations in Thailand. These polymorphic markers will be useful for future molecular genetics studies on *O. brevirostris* and related species.

Based on microsatellite data, levels of genetic diversity we detected for *O. brevirostris* in Thailand waters (combined $N_a = 5.90$, $N_e = 2.87$, $uHe = 0.59$, and $Ho = 0.51$) are higher than those from Mekong River (combined $N_a = 3.40$, $N_e = 1.91$, $uHe = 0.45$, and $Ho = 0.41$) reported by Krützen *et al.* (2018). Differences in genetic diversity among *O. brevirostris*

populations may be associated with the different polymorphic microsatellite markers used in these two studies. However, our analysis of mtDNA data also reveals that the haplotype and nucleotide diversities of *O. brevirostris* in Thailand ($Hd = 0.925$, $\pi = 0.900\%$) are much higher than those in Mekong River ($Hd = 0.418$, $\pi = 0.120\%$) and Chilika Lagoon ($Hd = 0.345$, $\pi = 0.078\%$) reported by Caballero *et al.* (2019). Even if we remove the two individuals from AS, the genetic diversity of *O. brevirostris* within the Gulf of Thailand remains high ($Hd = 0.915$ and $\pi = 0.860\%$), which is comparable to those in the Gulf of Thailand ($Hd = 0.845$ and $\pi = 1.150\%$) reported by Caballero *et al.* (2019). A similar situation is reported for the similarly small Indo-Pacific finless porpoise *Neophocaena phocaenoides*, for which the genetic diversity of a riverine population was lower than coastal populations (Jia *et al.*, 2014). It is likely that decreased genetic diversity in *O. brevirostris* in Mekong River and Chilika Lagoon is related to a small effective population size and long-term isolation.

ANOVA based on microsatellite data reveals significantly different N_a values among the four sampling sites in Thailand, with the mean N_a in AS the lowest. However, N_e values from these locations did not significantly differ. N_e is an estimate of the number of equally frequent alleles in an ideal population, which enables meaningful comparisons of allele diversity across loci with diverse allele frequency distributions (Peakall and Smouse, 2012, Appendix 1). There is also no significant difference in other genetic diversity estimators, such as mean values for H_e , uH_e , H_o and I values. Compared with H_e and uH_e , the I value may be a better measure of allelic and genetic diversity, because it is not bounded by 1 (Peakall and Smouse, 2012, Appendix 1). Additionally, analysis of mtDNA data reveals Hd and π values in AS are much higher than those in the other three regions. Results of genetic diversity indices in AS may be controversial because we could include only two individuals in genetic analysis. No significant differences were found in genetic diversity among WG, EG, and NG in the Gulf of Thailand.

Overall, genetic differentiation among geographic regions was statistically significant based on mtDNA and microsatellite analyses. Our results are consistent with statistical differences in average stable isotope values reported from *O. brevirostris* teeth in the same regions by Jackson-Ricketts *et al.* (2018), indicating distinct geographic groups (Eastern Gulf, Northern Gulf, Western Gulf, and Andaman Sea) may exist in Thailand's waters. However, no significant difference was detected for average tooth stable isotope values between northern and western Gulf regions. We report, to the contrary, significant genetic differences between NG and WG, suggesting *O. brevirostris* in these two regions are unlikely to be components of one large subpopulation, as stable isotope analysis is able to detect ecological populations, which may not correspond to genetic populations. Therefore, a lack of observed isotope differences might be due to similarities in habitat biochemistry in these two regions.

Our microsatellite data reveals the genetic structure of *O. brevirostris* in Thailand follows an IBD model, suggesting individual movements and exchanges are more likely to occur between adjacent regions. Therefore, the significant genetic differentiation detected between WG and NG regions might be

explained by geographic distance (the mean distance between WG and NG individuals is 494.481 km), which is larger than individual dispersal distances. Although smaller geographic distances (mean 228.072 km) between EG and NG, significant genetic differentiation (both mtDNA and microsatellite data) remained between them. No significant genetic differentiation was found between AS and other regions based on different types of molecular markers, despite considerable distances between them (mean 653.853 km) and obvious environmental breaks, such as the Strait of Malacca. Perhaps this is an artefact of small sample sizes in AS, and that more genetic data are needed to accurately determine population structure between this region and elsewhere.

Environmental factors and coastal development have likely played important roles in driving genetic differentiation of *O. brevirostris* in Thailand. Geographical barriers formed from oceanographic variables, such as ocean currents, upwelling, bathymetry, sea surface temperature, primary productivity, and salinity, can affect cetacean genetic structure (Bilgmann *et al.*, 2007; Mendez *et al.*, 2010; 2011; Amaral *et al.*, 2012; 2017). In Thailand, *O. brevirostris* have been reported from along the coast of Trat province and the Bang Pakong Estuary adjacent to the Chonburi province (Tongnunui *et al.*, 2011; Hines *et al.*, 2015; Jackson-Ricketts *et al.*, 2020). However, *O. brevirostris* today appears to be absent from approximately 250 km of coastline neighboring Chanthaburi and Rayong provinces, between Chonburi and Trat provinces. Shallow depths, high water turbidity, and short distances to river mouths may be linked to the distribution of *O. brevirostris* in the northern and eastern Gulf of Thailand (Tongnunui *et al.*, 2011; Jackson-Ricketts *et al.*, 2020). Anthropogenic barriers to individual dispersal caused by coastal development may also contribute to population fragmentation of inshore dolphins (Brown *et al.*, 2014). Given anecdotal historical evidence for *O. brevirostris* occurring in Chanthaburi, its absence today may be also be related to higher levels of fishing pressure and industrial development in Chanthaburi and Rayong provinces (Jackson-Ricketts *et al.*, 2020).

In addition to geographic distribution patterns, and environmental and anthropogenic factors, adaptations caused by the complex social behaviors of cetaceans may contribute to genetic divergence (Mendez *et al.*, 2013; Fruet *et al.*, 2014). Some social and specialized foraging behaviors have been reported to be associated with significant levels of population structure in *T. aduncus* (Krützen *et al.*, 2004; Frère *et al.*, 2010) and *O. orca* (Ford *et al.*, 1998; Matkin *et al.*, 2007; Riesch *et al.*, 2012). We speculate that local adaptations related to foraging strategies may have contributed to genetic differentiation of *O. brevirostris* in Thailand, because isotopic dietary analysis reveals *O. brevirostris* forages on different proportions of prey in neighboring locations where geographic barriers in the eastern Gulf of Thailand are lacking (Jackson-Ricketts *et al.*, 2018). Therefore, *O. brevirostris* in Thailand may have high site fidelity because of local prey resources and specialized foraging behavior.

Conclusion

In this study, 10 microsatellites were successfully cross-amplified and demonstrated to be polymorphic in *O.*

brevirostris from the eastern, western, and northern Gulf of Thailand and Andaman Sea coast. Based on analyses of both mtDNA and microsatellites, genetic diversity does not differ significantly among sampling locations in the Gulf of Thailand, but significant genetic differentiation is apparent between different region pairs. We speculate that significant genetic structure is associated with a combination of geographical distribution patterns, environmental and anthropogenic factors, and local behavioral adaptations.

Caution interpreting these results should be exercised because our sample size for each region was small, and samples were collected in an opportunistic manner, in that all individuals were dead (beach stranded). Our results, together with isotope variability in teeth (Jackson-Ricketts *et al.*, 2018), indicate that *O. brevirostris* may use specific habitats and have restricted home ranges throughout the life cycle in Thailand's waters. An improved understanding of the level of conservation and management for *O. brevirostris* in Thailand's waters will require more attention being paid to demographic, ecological and genetic-related issues regarding this endangered species.

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Conflict of interest

The authors declare no conflicts of interest.

Author Contributions

YD, KK, and XW conceived the study; YD and RC conducted the experiments, analyzed the data and wrote the manuscript; WS contributed to sample collection and DNA extraction; LZ, RA and FW contributed to data analysis and revising the manuscript; KK and XW supervised the whole project; all authors read and approved the final version.

References

- Amaral AR, Beheregaray LB, Bilgmann K, Boutov D, de Freitas LA, Robertson KM, Sequeira M, Stockin KA, Coelho MM and Moeller LM (2012) Seascape genetics of a globally distributed, highly mobile marine mammal: The short-beaked common dolphin (Genus *Delphinus*). *PLoS One* 7:e31482.
- Amaral AR, Smith BD, Mansur RM, Brownell RL and Rosenbaum HC (2017) Oceanographic drivers of population differentiation in Indo-Pacific bottlenose (*Tursiops aduncus*) and humpback (*Sousa spp.*) dolphins of the northern Bay of Bengal. *Conserv Genet* 18:371–381.
- Anderson M and Kinze CC (1999) Annotated checklist and identification key to the whales, dolphins, and porpoises (*Order Cetacea*) of Thailand and adjacent waters. *Nat Hist Bull Siam Soc* 47:27–62.
- Barbara T, Palma-Silva C, Paggi GM, Bered F, Fay MF and Lexer C (2007) Cross-species transfer of nuclear microsatellite markers: Potential and limitations. *Mol Ecol* 16:3759–3767.
- Beasley IL, Chooruk S and Piwpong N (2002) The status of the Irrawaddy dolphin, *Orcaella brevirostris*, in Songkhla Lake, Southern Thailand. *Raffles Bull Zool* 10:75–83.
- Beasley IL, Pollock K, Jefferson TA, Arnold P, Morse L, Yim S, Lor Kim S and Marsh H (2013) Likely future extirpation of another Asian river dolphin: The critically endangered population of the Irrawaddy dolphin in the Mekong River is small and declining. *Mar Mamm Sci* 29:226–252.
- Bilgmann K, Moller LM, Harcourt RG, Gibbs SE and Beheregaray LB (2007) Genetic differentiation in bottlenose dolphins from South Australia: Association with local oceanography and coastal geography. *Mar Ecol Prog Ser* 341:265–276.
- Bonin A, Bellemain E, Bronken Eidesen P, Pompanon F, Brochmann C and Taberlet P (2004) How to track and assess genotyping errors in population genetics studies. *Mol Ecol* 13:3261–3273.
- Brown AM, Kopps AM, Allen SJ, Bejder L, Littleford-Colquhoun B, Parra GJ, Cagnazzi D, Thiele D, Palmer C *et al.* (2014) Population differentiation and hybridisation of Australian Snubfin (*Orcaella heinsohni*) and Indo-Pacific Humpback (*Sousa chinensis*) Dolphins in North-Western Australia. *PLoS One*:e101427.
- Caballero S, Dove V, Jackson-Ricketts J, Junchompoo C, Cohen S and Hines E (2019) Mitochondrial DNA diversity and population structure in the Irrawaddy dolphin (*Orcaella brevirostris*) from the Gulf of Thailand and the Mekong River. *Mar Mamm Sci* 35:300–310.
- Carvajal-Rodriguez A (2018) Myriads: P-value-based multiple testing correction. *Bioinformatics* 34:1043–104.
- Chantrapornsy S, Adulyanukosol K and Kittiwathanawong K (1996) Records of cetaceans in Thailand. *Res Bull* 61:39–63.
- Chantrapornsy S, Adulyanukosol K and Kittiwattanawong K (1999) Stranded cetaceans from Thailand. *IBI Reports* 9:55–72.
- Cornuet JM and Luikart G (1996) Description and power analysis of two tests for detecting recent population bottlenecks from allele frequency data. *Genetics* 144:2001–2014.
- De Barba M, Miquel C, Lobréaux S, Quenette PY, Swenson JE and Taberlet P (2017) High-throughput microsatellite genotyping in ecology: Improved accuracy, efficiency, standardization and success with low-quantity and degraded DNA. *Mol Ecol Resour* 17:492–507.
- Dolar MLL, Perrin WF, Gaudio JP, Yaptinchay AASP and Tan JML (2002) Preliminary report on a small estuarine population of Irrawaddy dolphins *Orcaella brevirostris* in the Philippines. *Raffles Bull Zool* 10:155–160.
- Earl DA and VonHoldt BM (2012) Structure Harvester: A website and program for visualizing Structure output and implementing the Evanno method. *Conserv Genet Resour* 4:359–361.
- Evanno G, Regnaut S and Goudet J (2005) Detecting the number of clusters of individuals using the software STRUCTURE: A simulation study. *Mol Ecol* 14:2611–2620.
- Excoffier L, Laval G and Schneider S (2005) Arlequin (version 3.0): An integrated software package for population genetics data analysis. *Evol Bioinform* 1:47–50.
- Ford JK, Ellis GM, Barrett-Lennard LG, Morton AB, Palm RS and Balcomb KC (1998) Dietary specialization in two sympatric populations of killer whales (*Orcinus orca*) in coastal British Columbia and adjacent waters. *Can J Zool* 76:1456–1471.

- Frère CH, Krützen M, Mann J, Watson-Capps JJ, Tsai YJ, Patterson EM, Connor R, Bejder L and Sherwin WB (2010) Home range overlap, matrilineal and biparental kinship drive female associations in bottlenose dolphins. *Anim Behav* 80:481–486.
- Fruet PF, Secchi ER, Daura-Jorge F, Vermeulen E, Flores P, Simoes-Lopes P, Genoves R, Laporta P, DiTullio J and Möller L (2014) Remarkably low genetic diversity and strong population structure in common bottlenose dolphins (*Tursiops truncatus*) from coastal waters of the Southwestern Atlantic Ocean. *Conserv Genet* 15:879–895.
- Glenn TC and Schable NA (2005) Isolating microsatellite DNA loci. *Methods Enzymol* 395:202–222.
- Goudet J (1995) FSTAT (version 1.2): A computer program to calculate F-statistics. *J Hered* 86:485–486.
- Hines EM, Samantha S, Chalutip J, Ponnampalam LS, Ilangakoon AD, Jackson-Ricketts J and Monanunsap S. (2015) Line transect estimates of Irrawaddy dolphin abundance along the eastern Gulf Coast of Thailand. *Front Mar Sci* 2:63.
- Hoelzel AR, Hancock JM and Dover GA (1991) Evolution of the cetacean mitochondrial D-loop region. *Mol Biol Evol* 8:475–493.
- Hubisz MJ, Falush D, Stephens M and Pritchard JK (2009) Inferring weak population structure with the assistance of sample group information. *Mol Ecol Resour* 9:1322–1332.
- Hurst GD and Jiggins FM (2005) Problems with mitochondrial DNA as a marker in population, phylogeographic and phylogenetic studies: The effects of inherited symbionts. *Proc R Soc Lond B Biol Sci* 272:1525–1534.
- Jackson-Ricketts J, Ruiz-Cooley RI, Junchompoo C, Thongsukdee S and Costa DP (2018) Ontogenetic variation in diet and habitat of Irrawaddy dolphins (*Orcaella brevirostris*) in the Gulf of Thailand and the Andaman sea. *Mar Mamm Sci* 35:492–521.
- Jackson-Ricketts J, Junchompoo C, Hines E, Hazen EL, Ponnampalam LS, Ilangakoon AD and Monanunsap S (2020) Habitat modeling of Irrawaddy dolphins (*Orcaella brevirostris*) in the Eastern Gulf of Thailand. *Ecol Evol* 10:2778–2792.
- Jakobsson M and Rosenberg NA (2007) CLUMPP: A cluster matching and permutation program for dealing with label switching and multimodality in analysis of population structure. *Bioinformatics* 23:1801–1806.
- Jayasankar P, Patel A, Khan MZ, Das P and Panda S (2011) Mitochondrial DNA diversity and PCR-based sex determination of Irrawaddy dolphin (*Orcaella brevirostris*) from Chilika Lagoon, India. *Mol Biol Rep* 38:1661–1668.
- Jefferson TA, Hung SK and Würsig B (2009) Protecting small cetaceans from coastal development: Impact assessment and mitigation experience in Hong Kong. *Mar Policy* 33:305–311.
- Jia K, Lin W, Gui D, Karczmarski L and Wu Y (2014) Molecular evidence reveals the distinctiveness of Indo-Pacific finless porpoises (*Neophocaena phocaenoides*) in the Pearl River Estuary and insights into genus *Neophocaena*'s origin. *Mar Biol* 161:1919–1930.
- Jutapruet S, Intongcome A, Wang X, Kittiwattanawong K and Huang S (2017) Distribution of three sympatric cetacean species off the coast of the Central-Western Gulf of Thailand. *Aquat Mamm* 43:465–473.
- Kittiwattanawong K, Chantrapornsy S, Ninwat S and Chooruk S (2007) Review of the status and conservation of Irrawaddy dolphins *Orcaella brevirostris* in the Songkhla Lake of Thailand. In: Smith BD, Shore RG and Lopez A. Status and conservation of freshwater populations of Irrawaddy dolphins (Working paper 31). Wildlife Conservation Society, New York, pp 83–86.
- Kreb D (2004) Abundance of freshwater Irrawaddy dolphins in the Mahakam River in East Kalimantan, Indonesia, based on mark-recapture analysis of photo-identified individuals. *J Cetacean Res Manag* 6:269–277.
- Krützen M, Sherwin WB, Berggren P and Gales NJ (2004) Population structure in an inshore cetacean revealed by microsatellite and mtDNA analysis: Bottlenose dolphins (*Tursiops* sp.) in Shark Bay, Western Australia. *Mar Mamm Sci* 20:28–47.
- Krützen M, Isabel B, Ackermann CY, Dietmar L, Arne L, Ryan GE, Bejder L, Parra GJ, Wolfensberger R and Spencer PBS (2018) Demographic collapse and low genetic diversity of the Irrawaddy dolphin population inhabiting the mekong river. *PLoS One* 13:e0189200.
- Kundu S, Sharma G, Balakrishnan S, Tyagi K, Chandra K and Kumar V (2019) DNA barcoding identified two endangered dolphins: Threats on living aquatic mammals in India. *Mitochondrial DNA B Resour* 4:77–80.
- Leigh JW and Bryant D (2015) Popart: Full-feature software for haplotype network construction. *Methods Ecol Evol* 6:1110–1116.
- Librado P and Rozas J (2009) DnaSP v5: A software for comprehensive analysis of DNA polymorphism data. *Bioinformatics* 25:1451–1452.
- Mace GM and Lande R (1991) Assessing extinction threats: Toward a reevaluation of IUCN Threatened Species categories. *Conserv Biol* 5:148–157.
- Mantel N (1967) The detection of disease clustering and a generalized regression approach. *Can Res* 27:209–220.
- Matkin CO, Barrett-Lennard LG, Yurk H, Ellifrit D and Trites AW (2007) Ecotypic variation and predatory behavior among killer whales (*Orcinus orca*) off the eastern Aleutian Islands, Alaska. *Fish Bull* 105:74–87.
- Mendez M, Rosenbaum HC, Subramaniam A, Yackulic C and Bordino P (2010) Isolation by environmental distance in mobile marine species: Molecular ecology of franciscana dolphins at their southern range. *Mol Ecol* 19:2212–2228.
- Mendez M, Subramaniam A, Collins T, Minton G, Baldwin R, Berggren P, Särnblad A, Amir OA, Peddemors V, Karczmarski L *et al.* (2011) Molecular ecology meets remote sensing: Environmental drivers to population structure of humpback dolphins in the Western Indian Ocean. *Heredity* 107:349–361.
- Mendez M, Jefferson TA, Kolokotronis S-O, Krützen M, Parra GJ, Collins T, Minton G, Baldwin R, Berggren P, Särnblad A *et al.* (2013) Integrating multiple lines of evidence to better understand the evolutionary divergence of humpback dolphins along their entire distribution range: A new dolphin species in Australian waters? *Mol Ecol* 22:5936–5948.
- Minton G, Peter C and Tuen AA (2011) Distribution of small cetaceans in the nearshore waters of Sarawak, East Malaysia. *Raffles Bull Zool* 59:91–100.
- Minton G, Peter C, Zulkiffi Poh AN, Ngeian J, Braulik G, Hammond PS and Tuen AA. (2013) Population estimates and distribution patterns of Irrawaddy dolphins (*Orcaella brevirostris*) and Indo-Pacific finless porpoise (*Neophocaena phocaenoides*) in the Kuching Bay, Sarawak. *Raffles Bull Zool* 61:877–888.
- Minton G, Smith BD, Braulik GT, Kreb D, Sutaria D and Reeves R (2017) *Orcaella brevirostris* (errata version published in 2018), The IUCN Red List of Threatened Species, <https://www.iucnredlist.org/species/15419/123790805>
- Niu FQ, Yang YM, Zhou ZM, Wang XY, Monanunsap S and Junchompoo C (2019) Echolocation clicks of free-ranging Irrawaddy dolphins (*Orcaella brevirostris*) in Trat Bay, the eastern Gulf of Thailand. *J Acoust Soc Am* 145:3031–3037.
- Peakall R, Smouse PE and Huff DR (1995) Evolutionary implications of allozyme and RAPD variation in diploid populations of dioecious buffalograss *Buchloë dactyloides*. *Mol Ecol* 4:135–148.

- Peakall R and Smouse PE (2006) GenAlEx 6: Genetic analysis in Excel. Population genetic software for teaching and research. *Mol Ecol Notes* 6:288–295.
- Peakall R and Smouse PE (2012) GenAlEx 6.5: Genetic analysis in Excel. Population genetic software for teaching and research - an update. *Bioinformatics* 28:2537–2539.
- Piry S, Luikart G and Cornuet JM (1999) BOTTLENECK: A computer program for detecting recent reductions in the effective population size using allele frequency data. *J Hered* 90:502–503.
- Pompanon F, Bonin A, Bellemain E and Taberlet P (2005) Genotyping errors: Causes, consequences and solutions. *Nat Rev Genet* 6:847–859.
- Pritchard JK, Stephens M and Donnelly P (2000) Inference of population structure using multilocus genotype data. *Genetics* 155:945–959.
- Riesch R, Barrettlennard LG, Ellis GM, Ford JK and Deecke VB (2012) Cultural traditions and the evolution of reproductive isolation: Ecological speciation in killer whales? *Biol J Linn Soc Lond* 106:1–17.
- Rosenberg NA (2004) DISTRUCT: A program for the graphical display of population structure. *Mol Ecol Notes* 4:137–138.
- Rousset F (2008) GENEPOP' 007: A complete re-implementation of the GENEPOP software for Windows and Linux. *Mol Ecol Resour* 8:103–106.
- Selkoe KA and Toonen RJ (2006) Microsatellites for ecologists: A practical guide to using and evaluating microsatellite markers. *Ecol Lett* 9:615–629.
- Shen Z, Qu W, Wang W, Lu Y, Wu Y, Li Z, Hang X, Wang X, Zhao D and Zhang C. (2010) MPprimer: A program for reliable multiplex PCR primer design. *BMC Bioinformatics* 11:143.
- Smith BD and Hobbs L (2002) Status of Irrawaddy dolphins *Orcaella brevirostris* in the upper reaches of the Ayeyarwady River, Myanmar. *Raffles Bull Zool* 10:67–73.
- Smith BD, Beasley I and Krebs D (2003) Marked declines in populations of Irrawaddy dolphins. *Oryx* 37:401–406.
- Smith BD, Braulik GT, Strindberg S, Ahmed B and Mansur RM (2006) Abundance of Irrawaddy dolphins (*Orcaella brevirostris*) and Ganges River dolphins (*Platanista gangetica gangetica*) estimated using concurrent counts made by independence teams in waterways of the Sundarbans mangrove forest in Bangladesh. *Mar Mamm Sci* 22:527–547.
- Smouse PE and Peakall R (1999) Spatial autocorrelation analysis of individual multiallele and multilocus genetic structure. *Heredity* 82:561–573.
- Stacey PJ and Arnold PW (1999) *Orcaella brevirostris*. *Mammalian Species* 616:1–8.
- Sutaria D and Marsh H (2011) Abundance estimates of Irrawaddy dolphins in Chilka Lagoon, India, using photo-identification based mark-recapture methods. *Mar Mamm Sci* 27:338–348.
- Taberlet P, Griffin S, Goossens B, Questiau S, Manceau V, Escaravage N, Bouvet, J (1996) Reliable genotyping of samples with very low DNA quantities using PCR. *Nucleic Acids Res* 24:3189–3194.
- Tamura K, Peterson DS, Peterson N, Stecher G, Nei M and Kumar S (2011) MEGA5: Molecular evolutionary genetics analysis using maximum likelihood, evolutionary distance, and maximum parsimony methods. *Mol Biol Evol* 28:2731–2739.
- Tongnunui S, Wattanakornsiri A, Pachana K, Beamish FW and Tongsukdee S (2011) Preliminary investigation of Irrawaddy dolphin (*Orcaella brevirostris*) in the Bangpakong Estuary, Inner Gulf of Thailand. *Environ Nat Resour J* 9:48–57.
- Van Oosterhout C, Hutchinson WF, Wills DPM and Shipley P (2004) Micro-checker: Software for identifying and correcting genotyping errors in microsatellite data. *Mol Ecol Notes* 4:535–538.
- Weir BS and Cockerham CC (1984) Estimating F-statistics for the analysis of population structure. *Evolution* 38:1358–1370.
- Wilcoxon F (1945) Individual comparisons by ranking methods. *Biometrics Bull* 1:80–83.
- Wilson AC, Cann RL, Carr SM, George MR, Gyllensten UB, Helm-Bychowski KM, Higuchi R, Palumbi S, Prager E, Sage R and Stoneking M (1985) Mitochondrial DNA and two perspectives on evolutionary genetics. *Biol J Linn Soc Lond* 26:375–400.

Supplementary Material

The following online material is available for this article:

Table S1 – Multiplex design information for all the tested 18 microsatellite loci for *O. brevirostris* in Thailand.

Table S2 – GenBank accession number of all the 32 mtDNA sequences including 15 haplotypes for *O. brevirostris* in Thailand.

Table S3 – Genetic diversity parameters in four sampling locations of *O. brevirostris* in Thailand.

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