

## Article

## Benthic macroinvertebrate communities in sites with native forest presence and absence in north Patagonia

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**ABSTRACT.** The benthic fauna in northern Patagonian streams is characterized by presence of water quality indicator species. The aim of the present study was characterized the benthic communities in streams associated to sites with presence of *Nothofagus* forests and with human intervention without *Nothofagus* forests using null models in ecology, unified neutral theory of biodiversity (UNTB) and unweighted pair group method with arithmetic mean (UPGMA). The results of species co-occurrence null model revealed that species associations are random, whereas the results of niche overlap null models revealed the presence of niche overlap. The UNTB results revealed low species number. The results of UPGMA revealed that species composition is specific for each site existing marked differences. The exposed results agree with results observed for central and northern Patagonian Chilean rivers, where the species composition is variable in according to surrounding basin properties and seasonal conditions.

**KEYWORDS.** Benthic macroinvertebrates, *Nothofagus*, communities, null models, aquatic insects.

**RESUMO.** Comunidades de macroinvertebrados bentônicos em locais com presença e ausência de floresta nativa no norte da Patagônia. A fauna bentônica nos córregos do norte da Patagônia é caracterizada pela presença de espécies indicadoras de qualidade da água. O objetivo do presente estudo foi caracterizar as comunidades bentônicas em riachos associados a locais com presença de florestas de *Nothofagus* e com intervenção humana sem florestas de *Nothofagus* usando modelos nulos em ecologia, teoria unificada de biodiversidade unificada (UNTB) e método de grupo de par não ponderado com média aritmética (UPGMA). Os resultados do modelo nulo de co-ocorrência de espécies revelaram que as associações de espécies são aleatórias, enquanto os resultados de modelos nulos de sobreposição de nicho revelaram a presença de sobreposição de nicho. Os resultados do UNTB revelaram baixo número de espécies. Os resultados da UPGMA demonstraram que a composição de espécies é específica para cada local, existindo diferenças marcadas. Os resultados expostos concordam com os resultados observados para os rios chileno central e norte da Patagônia, onde a composição das espécies é variável de acordo com as propriedades da bacia circundante e às condições sazonais.

**PALAVRAS-CHAVE.** Macroinvertebrados bentônicos, *Nothofagus*, comunidades, modelos nulos, insetos aquáticos.

The aquatic invertebrates in Chilean inland waters are characterized by its marked endemism (OYANEDEL *et al.*, 2008; MOYA *et al.*, 2009; DE LOS RÍOS-ESCALANTE *et al.*, 2013) at different spatial scales, becoming marked microendemism situations (PALMA *et al.*, 2013). The benthic species composition is variable in function of geographical conditions (PALMA *et al.*, 2013), and water quality due changes in their surrounding basin, specifically native forest replace by agricultural, urban and industrial towns (FIGUEROA *et al.*, 2003, 2007; ENCINA *et al.*, 2017; VEGA *et al.*, 2017) or fish predation (SOTO *et al.*, 2006, 2007; VARGAS *et al.*, 2010; ENCINA *et al.*, 2017; VEGA *et al.*, 2017)

Northern Patagonian rivers can have mixture regime, it is originated by winter rains and summer snow smelt, or originated by lake effluents (NIEMEYER & CERECEDA, 1984; ENCINA *et al.*, 2017). In Araucania region (38°S), it is possible found both kinds of rivers (ENCINA *et al.*, 2017), and many of these rivers have surrounding basins with *Nothofagus dombeyi* (Mirb.) Oerst. perennial native forest or with agricultural, urban and industrial zones (VEGA *et al.*, 2017), also many of these rivers and streams have introduced salmonids populations that generate marked alterations in benthic communities (ENCINA *et al.*, 2017; VEGA *et al.*, 2017). In this context, the rivers and lakes of Araucania region in

mountain zones have marked presence of native forest, that would generate a kind of pristine condition of these habitats with marked differences in comparison to sites without native forest in their surrounding basin (DE LOS RÍOS & ROA, 2010; DE LOS RÍOS-ESCALANTE, 2010).

The presence of native forest in Andean zones of Araucania region (38°S) would be an important determinant of water quality in associated rivers and lakes (PEDROZO *et al.*, 1993; DE LOS RÍOS-ESCALANTE, 2010), that in consequence would affect the aquatic communities such as was observed for pelagial environment (DE LOS RÍOS-ESCALANTE *et al.*, 2012). There are not published information about benthic fauna in Araucania region (38°S) among river beds, only there are an inventory without site specifications of macroinvertebrate fauna in Cautin river (VEGA *et al.*, 2020), Crustacean in Tolten river (DE LOS RÍOS-ESCALANTE *et al.*, in press) and descriptions of benthic macroinvertebrates as preys on native fishes in Tolten river (BARILE *et al.*, in press). Then there are not studies on variations on benthic fauna in Araucania region among riverbeds and in this context, the aim of the present study is to analyse the benthic communities in sites of a mountain stream with two kinds of surrounding basins, *N. dombeyi* native perennial forest and mixture of native forest with agricultural zones in Araucania region (39°S, North Patagonia). As hypothesis, it suggest that benthic community would change in function of kind of surrounding

basin in a gradient of native perennial forest, transition between perennial forest and agricultural zones.

## MATERIAL AND METHODS

**Study site.** The sub-basins analyzed belong to the high zones of Quepe river basin that is located in central-southern Chile (38°40'S; 71°45'W, Fig. 1). The area has a pluvial regime (3800 mm) and with snow (LUEBERT & PLISCOFF, 2006). It has a warm-rainy template climate with Mediterranean influence and cold-rainy template with Mediterranean influence in high zones with low temperature dominance every year. The soils are deep with good drainage and muddy texture with null to weak erosion in his surface (CIREN-CORFO, 2002). In its high zone, it found native forest, in its median and low zone there are soils with agricultural and poultry zones.

**Field work.** Three sites with *Nothofagus* forest and two with mixed (*Nothofagus* and agriculture use) were sampled (Fig. 1); benthic macroinvertebrates were collected using a 50 \* 50 cm Surber net of 80 mm mesh size, samples were collected on triplicate, in according to the descriptions of FIGUEROA *et al.* (2003, 2007), the collected benthic specimens were fixed with absolute ethanol, identified with specialized literature at low possible taxonomic level, that was family or genus (GONZÁLEZ, 2003; DOMÍNGUEZ & FERNÁNDEZ, 2009).

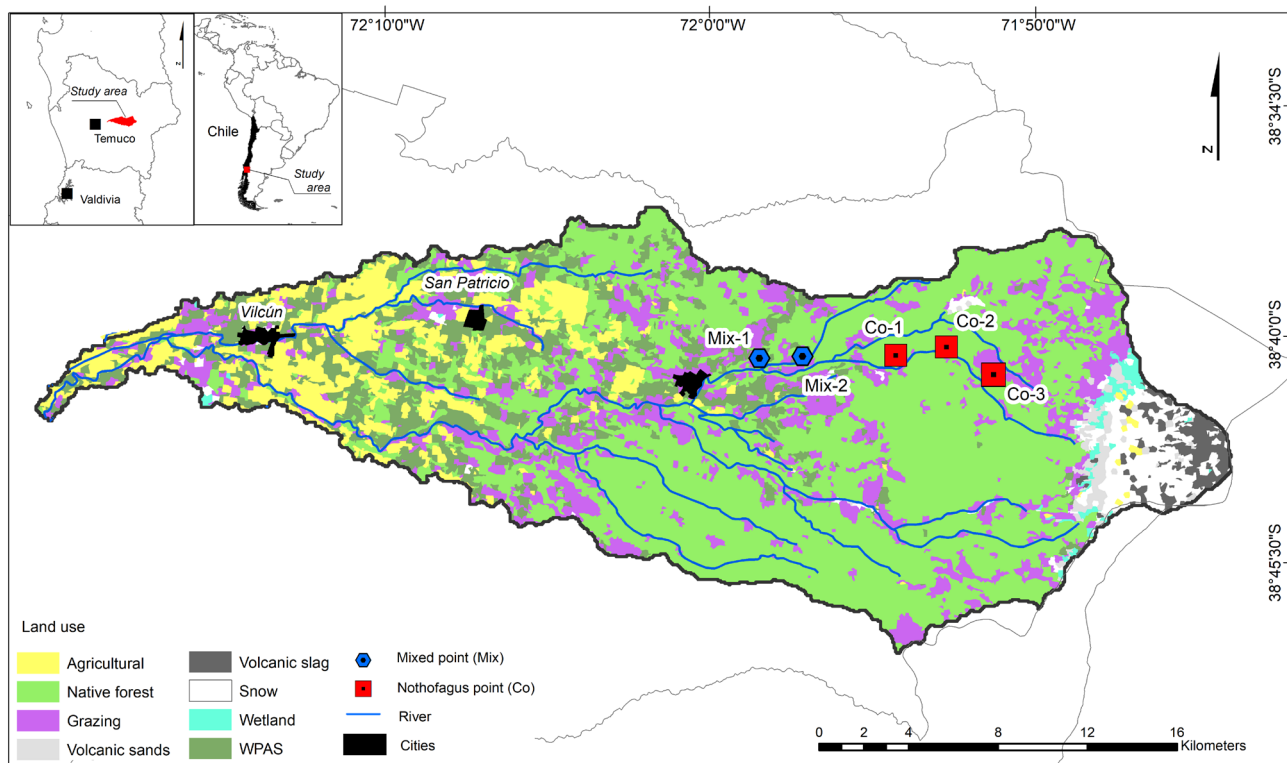


Fig 1. Map with studied site in Araucania region, Chile (Mix-1, Mixed P1; Mix-2, Mixed P2; Co-1, Coigüe P1; Co-2, Coigüe P2; Co-3, Coigüe P3).

**Data analysis.** A species absence/presence matrix was constructed, with the species in rows and the sites in columns. We calculated the Checkerboard score (“C-score”), which is a quantitative index of occurrence that measures the extent to which species co-occur less frequently than expected by chance (GOTELLI, 2000). A community is structured by competition when the C-score is significantly larger than expected by chance (GOTELLI, 2000; TONDOH, 2006; TIHO & JOHENS, 2007). In addition, we compared co-occurrence patterns with null expectations via simulation. GOTELLI & ELLISON (2013) suggested the Fixed-Fixed model as statistical null model; in this model the sums of the rows and columns of the matrix are preserved. Thus, each random community contains the same number of species as the original community (fixed column), and each species occurs with the same frequency as in the original community (fixed row). The null model analyses were performed using “R” software (R CORE TEAM, 2009) and the package EcosimR version 7.0 (GOTELLI & ELLISON, 2013; CARVAJAL-QUINTERO *et al.*, 2015).

As second step were analysed under a niche-sharing null model using Pianka’s overlap index with retained niche breadth and reshuffled zero states using “R” software (R CORE TEAM, 2009) and the package EcosimR version 7.0 (GOTELLI & ELLISON, 2013; CARVAJAL-QUINTERO *et al.*, 2015). The EcosimR program also determines whether measured overlap values differ from what would be expected in a random sampling of the species data. EcosimR performs Monte Carlo randomisations to create pseudo-communities and then statistically compares the patterns of these randomised communities with those in the real data matrix (GOTELLI & ELLISON, 2013). In our analysis all values of the general matrix were randomised 1000 times and the niche breadth was retained for each species; in this way, the algorithm retained the amount of specialization for each species (GOTELLI & ELLISON, 2013; CARVAJAL-QUINTERO *et al.*, 2015). In a second step we applied UPGMA cluster analysis using the neighbour-joining method to determine potential similarities between sites on the base of species associations using “R” software (R CORE TEAM, 2009) and the R package phangorn (SCHLIEP, 2018).

In a third step, it was applied analysis on the view point of unified neutral theory of biodiversity (UNTB; HUBBELL, 2001) using the R package UNTB (HANKIN, 2018), it was applied the  $\theta$  estimator and Preston analysis. Finally, the data were ordered to make species richness estimations considering the presence–absence data using Chao 2, Jackknife 1 and Jackknife 2 indexes with “R” software (R CORE TEAM, 2009) and the R package fossil (VAVREK, 2011)

with the aim of understanding the community properties (CHAO, 1984; GOTELLI & CHAO, 2013).

## RESULTS

The results revealed that for autumn samples, Gastropoda appears a Mixed P2, whereas Oligochaeta appears only in Coigüe P1; Ephemeroptera, specifically Nesameletidae, in sites Coigüe P1-Mixed P1 is most abundant; Diptera, has low abundance only with Athericidae at site Coigüe P3, increase at site Coigüe P2, decreasing in successive low sites; Megaloptera (Corydalidae and Sialidae) appear at Coigüe P1 (Tab. I). In order Plecoptera, Gripopterygidae is the most abundant appearing in four sites, the second dominant is Austroperlidae with two sites, the most abundance was in Coigüe P2 and Mixed P1; Trichoptera more abundant was *Smicridea* that was abundant in Coigüe P1, Coigüe P2, Mixed P2, also, Ecnomidae was present in Coigüe P3 and Mixed P1, and Sericostomatidae was present in Mixed P2; Crustacea, has only *Hyalella chiloensis* at Coigüe P1 (Tab. I).

In winter, Oligochaeta was present with high abundance at MX2, and low abundances in decreasing order at Coigüe P1 and Coigüe P3; Ephemeroptera was most abundant with high diversity at all Coigüe sites, with high diversity at Coigüe P1, and high abundance at Coigüe P3, and in Mixed sites was only Notonemouridae with low abundance (Tab. I). The order Diptera has high diversity in Coigüe P1, and high abundance in Coigüe P3, whereas in Mixed sites it has low abundances; the order Megaloptera, was present only in Mixed P2; Plecoptera has high diversity at site Coigüe P1, and high abundance at Coigüe P2, decreasing its abundance at Mixed sites; finally order Trichoptera, genus *Smicridea* (Hydropsychidae) was present in all sites being abundant in Coigüe P2 site (Tab. I).

The results revealed a high species number and individual abundances for autumn samples in comparison to winter samples, also the sites with *N. dombeyi* has high species and individual abundances in comparison to sites with mixture of native forest with agricultural zones (Tabs I, II). The results of null model analysis revealed that species associations are random and there is niche overlapping for both sampling periods (Tabs I, II).

The UPGMA revealed for winter a marked similarity between native forest 2 with mixed 2, continuing native forest 3, mixture 3 and native forest 1; whereas for autumn it was marked similarity between native forest 2 and 3, and successively with mixture 3, native forest 1 and mixture 2 (Fig. 2). The results of UNTB revealed low species number, the optimal  $\theta$  for autumn and winter was 5.72 and 3.72,

respectively, the singletons were 4 and 2 for autumn and winter respectively, this mean that autumn has high species number that would be equivalent that each species are practically equivalent to one individual. Finally for autumn

the Chao-2, Jack-1 and Jack-2 values were 42.16; 32.77 and 39.81 respectively, whereas for winter Chao-2, Jack-1 and Jack-2 values were 13.25; 13.97 and 12.99 respectively, that would show the low species number for winter period.

Tab. I. Macroinvertebrate abundances (ind/m<sup>2</sup>) for studied sites in Araucania region, Chile during Autumn and results of co-occurrence and niche sharing null model analysis.

		Coigüe P1	Coigüe P2	Coigüe P3	Mixed P1	Mixed P2
Location		38°40'37"S; 71°54'29"W	38°40'12"S; 71°53'35"W	38°41'00"S; 71°50'52"W	38°40'37"S; 71°58'05"W	38°40'44"S; 71°57'25"W
Gastropoda						
	Gastropoda indet.	0.0	0.0	0.0	2.0	0.0
Oligochaeta						
	Oligochaeta indet.	2.7	0.7	0.7	0.0	0.7
Ephemeroptera						
	Baetidae	0.0	0.0	3.3	0.0	0.0
	Coloboruscidae	0.3	0.0	0.0	0.0	0.0
	Leptophlebiidae	0.0	0.3	0.0	0.0	0.0
	Nesameletidae	27.7	18.0	15.0	27.3	9.0
Diptera						
	Athericidae	1.3	6.7	4.7	1.0	0.0
	Ceratopogonidae	0.3	0.0	0.0	2.0	0.0
	Culicidae	0.0	0.0	0.0	0.3	0.0
	Chironomidae	6.0	3.0	0.0	3.3	2.0
	Psychocidae	3.7	0.3	0.0	0.3	0.0
	Simulidae	3.7	6.0	0.0	0.0	0.7
Megaloptera						
	Corydalidae	0.0	0.0	0.0	0.7	0.0
	Sialidae	0.0	0.0	0.0	1.0	0.0
Plecoptera						
	Austroperlidae	0.0	1.3	0.0	0.3	0.0
	Diamphipnoidae	0.0	0.0	0.0	1.7	0.0
	Eustheniidae	0.0	0.0	0.0	0.0	0.3
	Gripopterygidae	0.0	1.3	1.7	1.0	0.0
Trichoptera						
	<i>Smicridea</i> sp.	7.3	21.0	18.0	17.3	2.7
	Ecnomidae	0.0	0.0	4.0	1.3	0.0
	Sericostomatidae	0.0	0.0	0.0	0.0	0.3
Crustacea						
	<i>Hyalella chilensis</i>	1.7	0.0	0.0	0.0	0.0
Results of null model analysis						
	Observed index	Mean index		Standard Effect size	P	
Species co-occurrence	0.813	0.800		0.546	0.375 n.s	
Niche sharing	0.299	0.324		0.501	0.240 n.s	

Tab. II. Macroinvertebrate abundances (ind/m<sup>2</sup>) for studied sites in Araucania region, Chile during Winter and results of co-occurrence and niche sharing null model analysis.

		Coigüe P1	Coigüe P2	Coigüe P3	Mixed P2	Mixed P3
Oligochaeta						
	Oligochaeta indet.	2.3	0.0	0.3	0.0	13.3
Ephemeroptera						
	Ameletopsidae	0.3	0.0	6.0	0.0	0.0
	Coloboriscidae	2.7	0.0	0.0	0.0	0.0
	Notonemouridae	2.0	3.3	4.0	0.3	0.3
Diptera						
	Athericidae	0.3	1.3	0.3	1.0	0.0
	Chironomidae	1.0	1.0	0.0	0.3	0.3
	Psychocidae	0.3	0.0	2.0	0.7	0.0
	Simulidae	0.3	0.0	0.7	0.0	0.0
Megaloptera						
	Sialidae	0.0	0.0	0.0	0.0	0.3
Plecoptera						
	Gripopterygidae	0.7	0.3	0.0	0.0	0.0
	<i>Murphyella</i> sp.	4.3	21.3	20.3	10.0	7.7
	<i>Notoperla</i> sp.	0.7	0.0	0.0	0.0	0.0
Trichoptera						
	<i>Smicridea</i> sp.	2.0	20.0	1.3	4.0	1.7
Results of null model analysis						
	Observed index	Mean index	Standard Effect Size	<i>P</i>		
Species co-occurrence	1.842	1.807	0.795	0.230 n.s		
Niche sharing	0.377	0.377	3.536	0.407 n.s		

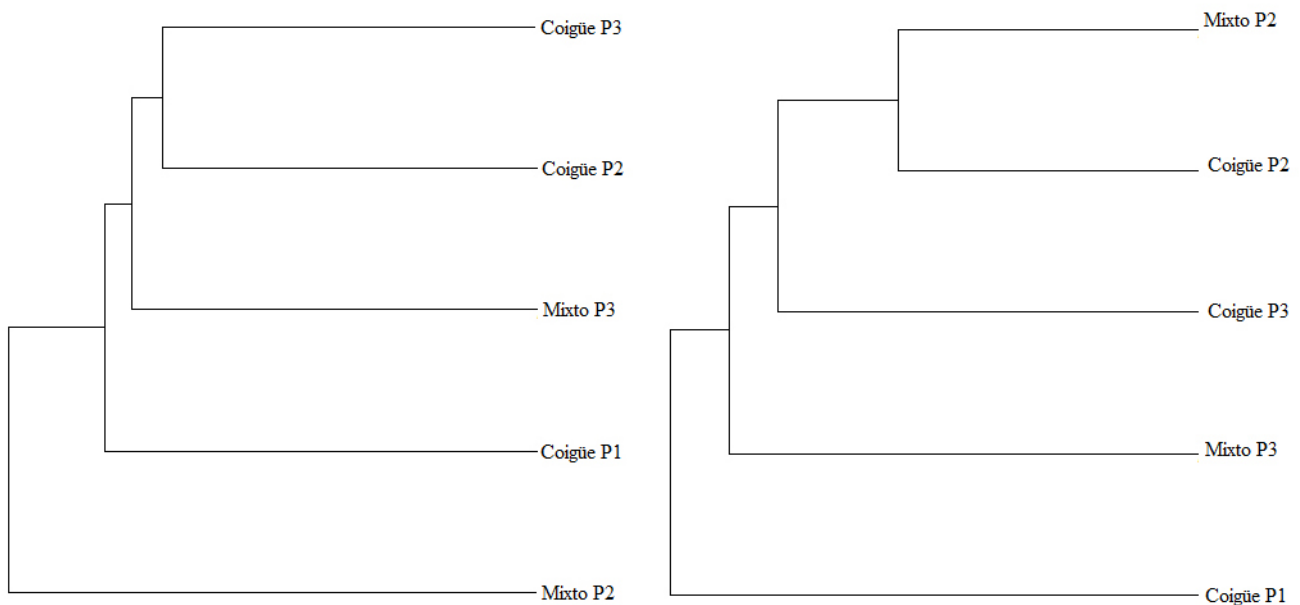


Fig 2. Results of UPGMA for autumn (left) and winter (right) data of macroinvertebrates studied at Araucaria Region, Chile.



## DISCUSSION

The exposed results would agree with observations for southern-central Chilean rivers that has a marked differences in their composition in according to the human intervention of the surrounding basin, specifically the marked abundance of shredders groups such as Ephemeroptera, Plecoptera and Trichoptera in zones with low human intervention and native forest, where there are many particulated organic matter (FIERRO *et al.*, 2012, 2015; FIGUEROA *et al.*, 2003, 2007; ENCINA *et al.* 2017; VEGA *et al.*, 2017). A different situation was reported for sites with mixture of native forest and human intervention, where the mentioned groups were scarce, and Diptera increase its abundances (FIGUEROA *et al.*, 2003, 2007, 2010, 2013). Then, on the view point of continuum river, there is a marked gradient between macroinvertebrate fauna in the origin of the riverbed, that was changes gradually in the medium and low zones of the riverbed, that was described with details for central Chilean rivers (FIGUEROA *et al.*, 2003, 2007) and Patagonian Argentinean rivers (MISERENDINO *et al.*, 2011, 2018).

In this scenario, the presence of a wide gradient of water quality parameters and the presence of potential predator fishes that predate on benthic invertebrates would regulate the benthic community structure (FIERRO *et al.*, 2016; ENCINA *et al.*, 2017; VEGA *et al.*, 2017, 2020; BARILE *et al.*, in press). Similar results have been reported also for Argentinean Patagonia where it is possible found rivers with native forest, and differents kind of human intervention (MISERENDINO, 1995, 1999, 2001; MISERENDINO *et al.*, 2011, 2018; PESSACQ & MISERENDINO, 2008), and for rivers in pristine zone in southern Chilean Patagonia (MOYA *et al.*, 2009; OYADEDEL *et al.*, 2008).

The results obtained in the present study, were collected in mountain zone, three stations in zone with native forest without human intervention, and two stations in a zone with human intervention, revealing the marked differences in benthic fauna, these results were similar to descriptions for mountain rivers in Catamarca, in northern Argentinean Andes (COLLA *et al.*, 2013) and Rio Grande do Sul in Brazil (BERTASO *et al.*, 2015). The exposed results are different to other descriptions of Chilean rivers, because, these studies were conducted in rivers bed with a marked human intervention gradient without include mountain zones where is located the riverbed origin (FIGUEROA *et al.*, 2003, 2007), perhaps, the similarities would be with studies in coastal mountain rivers with altered basins due agriculture and forestry activities between 38-39°S (FIERRO *et al.*, 2012, 2015).

If it is considered in first instance the marked endemism in benthic species fauna (DOMÍNGUEZ & FERNÁNDEZ, 2009), and the complex trophic interactions due abiotic environmental parameters and fish predation

(FIERRO *et al.*, 2016; ENCINA *et al.*, 2017; VEGA *et al.*, 2017; BARILE *et al.*, in press), it would explain the null model results observed in the present study: first the absence of structured pattern of species co-occurrence null model, and second the absence of niche segregation for both sampled periods. These results would indicate that it would be necessary more studies about trophic interactions in benthic fauna (INGS *et al.*, 2009; WOODWARD *et al.*, 2010; GRAY, 2005). Also, the role of shredders as main abundant component in benthic communities was reported for Andean rivers and would agree with the results of the present study (SCHMID-ARAYA *et al.*, 2012).

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