

Investigation of Low-Temperature Selective Catalytic Reduction of NO_x with Ammonia over Mn-Modified Fe₂O₃/AC Catalysts

Jiuyu Chen,^{a,#} Baozhong Zhu,^{a,#} Yunlan Sun,^{*,a,b} Shoulai Yin,^a Zicheng Zhu^a and Jiaxin Li^a

^aSchool of Energy and Environment and ^bKey Laboratory of Metallurgical Emission Reduction & Resources Recycling, Ministry of Education, Anhui University of Technology, 243002 Maanshan, Anhui, PR China

A series of iron-based catalysts supported on activated carbon (AC) and manganese-modified Fe₂O₃/AC catalysts with various Fe:Mn ratios were prepared for selective catalytic reduction (SCR) of NO by ammonia at low temperature. It was found that the addition of a small amount of MnO_x into the Fe₂O₃/AC catalyst contributed to an improvement of the NO conversion. NO conversions of approximately 100% were obtained for the 10FexMn/AC (x = 1, 3, 5) catalysts at 180-240 °C. The characterization results indicate that the addition of a certain amount of MnO_x into the 10Fe/AC catalyst increased the dispersion of the amorphous state, increased the ratios of Fe³⁺/(Fe³⁺ + Fe²⁺) and Mn⁴⁺/Mn³⁺, and improved the surface area and pore volume.

Keywords: Fe₂O₃/AC, manganese oxide, NH₃-SCR, NO conversion

Introduction

Nitrogen oxides (NO_x) are an important part of atmospheric pollution and are harmful to the environment as well as to human health.¹ NO_x can cause the formation of photochemical smog, acid rain and ozone depletion.² Several methods have been used to control NO_x emissions, such as selective catalytic reduction (SCR),^{3,4} the use of some absorbents (i.e. Fe^{II}-EDTA),^{2,5} plasma catalysis,⁶ and selective catalytic oxidation (SCO).⁷ Particularly, the SCR of NO_x with NH₃ is currently the most effective and commercialized way to realize the abatement of NO_x from power plants.⁸⁻¹⁰ In coal-fired power plants, the SCR reactor is installed upstream of the electrostatic precipitator and flue-gas desulfurization scrubber, so the catalyst suffers from deactivation due to dust accumulation and sulfur chemical poisoning. To solve these problems, the SCR reactor should be set downstream of the electrostatic precipitator and flue gas desulfurization scrubber. However, the temperature of flue gas is only approximately 200 °C. For conventional SCR catalysts, such as the V_2O_5 -WO₃/TiO₂ catalyst,^{10,11} the optimum operation temperature is usually approximately 300 to 400 °C. One better method to solve

*e-mail: yunlansun@163.com #Both authors contributed equally to this work. this problem is to develop a suitable low-temperature SCR catalyst.

In recent years, many studies have focused on low-temperature SCR catalysts. Transition metal oxides are often used as active components of catalysts, such as MnO_{x} , ^{12,13} VO_{x} , ¹⁴ $Co_{3}O_{4}$, ¹⁵ CuO_{x} , ¹⁶ and FeO_{x} . ¹⁷ It was reported that the oxides of Mn and Fe have good redox properties and oxygen storage capacities due to their changeable valences, low cost and lack of toxicity, especially for outstanding low-temperature SCR activity.^{18,19} Peña et al.²⁰ revealed that NO conversion on Mn/TiO₂ catalysts was approximately 100% at 140 °C, and the selectivity of N₂ was excellent at 120 °C. Zhu et al.17 prepared a series of Fe-Mn/TiO2 catalysts with holmium (Ho) and found that Fe_{0.3}Ho_{0.1}Mn_{0.4}/TiO₂ catalysts with a molar ratio of 1:10 (Ho:Ti) had the highest low-temperature SCR activity among these catalysts, and NO_x conversion was 90% when temperatures were in the range of 120 to 200 °C.

Catalyst carriers, such as TiO_2 ,²¹ Al_2O_3 ,^{22,23} active carbon²⁴ and modified activated carbon,²⁵⁻²⁷ are also very important for SCR performance at low temperature due to their high surface areas and thermal stabilities, which can help the active substance become highly dispersed. Among them, carbon materials have attracted much attention due to their well-developed pore structure, high surface area and low price. Active carbon (AC) is an ideal candidate for

use as a support for catalysts in the low-temperature NH₃-SCR reaction. Li et al.25 reported that the catalyst with a Cr/sargassum-based activated carbon (SAC) mass ratio of 0.02:1 exhibited the best NO_x-removing performance and the NO_x conversion is greater than 90% at 125-150 °C. Yang et al.9 found that the addition of a small amount of vanadium oxide into a Fe₂O₃/AC catalyst contributed to the best activities for the low-temperature NH₃-SCR of NO. However, few studies have involved FeO./AC over Mn-modified catalysts for the low-temperature NH₃-SCR of NO. In this paper, a series of Fe₂O₃/AC and Mn-modified Fe₂O₃/AC catalysts were prepared by an impregnation method to investigate the effect of different contents of Mn and Fe on the low-temperature NH₃-SCR of NO. The properties of the catalysts were characterized by isothermal N₂ adsorption/desorption (Brunauer-Emmett-Teller (BET)), X-ray fluorescence (XRF), X-ray diffraction (XRD), X-ray photoelectron spectroscopy (XPS) and temperature program reduction (H_2 -TPR).

Experimental

Catalyst preparation

Fe₂O₃/AC and Mn-modified Fe₂O₃/AC catalysts were prepared by an impregnation method using AC as the carrier and ferric nitrate (Fe(NO₃)₃·9H₂O) as the metal precursor. First, AC was washed with deionized water and dried for 6 h at 105 °C. Then, the AC was ground and sieved to obtain particles of 40-100 mesh. The impregnation process to obtain different masses of Fe-based catalysts was carried out as follows. A certain amount of ferric nitrate was dissolved into deionized water to form a solution. Then, AC was added to the solution. After stirring for 3 h, the mixtures were dried for 6 h at 105 °C and then calcined at 400 °C for 5 h.

Meanwhile, Mn-modified Fe_2O_3/AC catalysts were prepared by the same method as described above. Briefly, the pretreated AC was impregnated in solution, which involved the required amount of iron nitrate and manganese nitrate (Mn(NO_3)_2·6H_2O). The mixtures were stirred for 3 h at 25 °C, then dried for 6 h at 105 °C and calcined at 400 °C for 5 h in air. The catalysts are identified as 10FexMn/AC, where x represents the molar ratio of Mn and Fe, and 10 indicates the mole percent of Fe and AC.

Catalyst characterization

The surface areas and pore characterizations of these catalysts were determined by a Gold App V-Sorb 2800 analyzer (Gold App, China). Prior to BET measurement,

all catalysts were dried for 6 h at 105 $^{\circ}$ C and then degassed under vacuum at 180 $^{\circ}$ C for 12 h. The specific surface areas were calculated by the BET equation.

XRF measurements were carried out on an ARL Advant' X Intellipower 3600 (Thermo Scientific Niton TM, USA). The operating voltage and current were 60 kV and 60 mA, respectively. The spectrometer test environment was vacuum.

XRD measurements were carried out with a Rigaku Ultima IV powder diffractometer (Rigaku Ultima, Japan) using Cu K α radiation as a source to examine the crystallinity and dispersity of the components on the support. The scanning range (2 θ) was 10-80°, and the scanning rate was 0.02° s⁻¹. The operating voltage was 40 kV, and the applied current was 40 mA.

 H_2 -TPR measurements were carried out on a TP5080 automatic adsorption instrument (Xianquan, Tianjin). A calcined catalyst was placed into the reactor, dried at 300 °C for 1 h and cooled to 100 °C in a flow of Ar. Then, a flow of 500 ppm H_2/Ar replaced the Ar flow at a rate of 30 mL min⁻¹ for 30 min to adsorb H_2 . Ar was purged for 1 h, and after the baseline stabilized, the temperature was controlled between 100 and 600 °C at a heating rate of 10 °C min⁻¹. The thermal conductivity detector (TCD) signal was monitored online 10 times *per* second.

XPS analysis was conducted on an Escalab 250Xi X-ray electron spectrometer (Thermo Scientific, USA) with an Al K α X-ray source. The binding energy (BE) of Fe2p and Mn2p were calibrated using the C 1s peak (BE = 284.6 eV) as a standard.

Active test of the catalyst

The measurement of catalytic activity was carried out in a fixed-bed reactor, as shown in Figure 1. The simulated flue gas components included 5 vol% O_2 , 500 ppm NO, 500 ppm NH₃ and N₂ as balance at 120-240 °C. They were controlled by a mass flow rate controller. The simulated flue gas was mixed in a mixer, then heated in a preheater by the electric heating system. At last, it entered the reactor to reaction. The total flow rate was kept at 100 mL min⁻¹, and the space velocity was 6000 h⁻¹. The catalytic particle size was 40-100 mesh. In each test, 300 mg catalyst was packed in a quartz tube, which was positioned in the middle of the fixed-bed reactor. The concentration of NO in the inlet and outlet was analyzed by a ECOM-J2KN multi-function flue gas analyzer (Germany). The conversion of NO was calculated according to the following equation 1:

$$NO_{conversion} = \frac{NO_{in} - NO_{out}}{NO_{in}} \times 100\%$$
(1)

where $NO_{conversion}$ was the NO conversion, and NO_{in} and NO_{out} were the concentrations of NO at the inlet and outlet of the reactor, respectively.



Figure 1. Schematic diagram of the experimental setup to carry out the active test of the catalysts. (1) NO cylinder; (2) O_2 cylinder; (3) N_2 cylinder; (4) NH₃ cylinder; (5) cut-off valve; (6) mass flowmeter; (7) mixing reactor; (8) temperature controller; (9) gas preheater; (10) heating furnace; (11) catalyzer; (12) flue gas analyzer.

Results and Discussion

Catalytic performance

SCR activity of Fe₂O₃/AC catalysts

The NH₃-SCR of NO over AC with different Fe loadings was carried out in a fixed-bed reaction system between 120 and 240 °C. The results are shown in Figure 2. Temperature has a significant influence on the catalytic activity: the activity of AC decreases as temperature increases. However, the catalytic activity of AC with catalysts containing different Fe loadings increases with an increase of the reaction temperature, especially between 180 and 240 °C. The molar ratio of Fe and AC also affected the catalytic activity. The catalytic activity of all catalysts increases first and then decreases as the molar ratio of Fe and AC increases. The best catalytic activity is observed when the molar ratio of Fe and AC is 0.10, corresponding to a maximum NO conversion of 83.9% at 240 °C. However, when the molar ratio of Fe and AC is 0.15, the Fe₂O₃/AC catalyst shows the lowest NO conversion. These observations suggest that the contents of the active component are of great importance to the catalytic activity. The Fe₂O₃/AC catalyst with Fe/AC = 0.1 shows a maximum conversion of 83.9% at 240 °C, which indicates that an optimal dispersion of Fe species over the AC surface is obtained with this content of Fe in the Fe₂O₃/AC catalyst. The poor activity of the Fe_2O_3/AC catalyst with Fe/AC = 0.15 can be attributed to the agglomeration of Fe species on the AC surface, leading

to a lack of active catalytic sites. Therefore, the best molar ratio of Fe and AC is 0.1, which is further studied in the following experiments.



Figure 2. Influence of different molar ratios of Fe and AC on NO conversion in the SCR reaction at 120-240 °C for the Fe₂O₃/AC catalyst.

SCR activity of 10FexMn/AC catalysts

On the basis of the study mentioned above, the catalytic activity of the Fe_2O_3/AC catalyst is best when the molar ratio of Fe and AC is 0.10 (i.e., 10Fe/AC); thus, further studies are carried out using this ratio. To enhance the catalytic activity, Mn is introduced to the 10Fe/AC catalyst.

The NO conversion of the 10Fe/AC catalysts with different Fe and Mn molar ratios at 120-240 °C are shown in Figure 3. The NO conversion clearly increases with increasing temperature. Compared with the 10Fe/AC catalyst, the addition of Mn dramatically increases the catalytic performance. This result is similar to that of Liu et al.²⁸ At 120 °C, the NO conversion reaches 70, 68, 62, and 35% for the 10Fe1Mn/AC, 10Fe3Mn/AC, 10Fe5Mn/AC and 10Fe7Mn/AC catalysts, respectively (the catalysts are recorded as 10Fe1Mn/AC, 10Fe3Mn/AC, 10Fe5Mn/AC and 10Fe7Mn/AC when the ratio of Fe to Mn reaches 10:1, 10:3, 10:5 and 10:7, respectively). When the temperature increases from 180 to 240 °C, approximately 100% NO conversion is achieved for the 10Fe1Mn/AC, 10Fe3Mn/AC and 10Fe5Mn/AC catalysts. The NO conversion of the 10Fe7Mn/AC catalyst increases from 82 to 98% at 180-240 °C. These results may be caused by the aggregation of MnO_x particles on the AC surface, which can be interpreted by the following BET analysis.

NH₃/NO effect

As a reductant, NH_3 is an essential part of the SCR catalytic reaction and is very important to the whole process of NO_x conversion. The change of NH_3 /NO is realized by adjusting the gas flow of NH_3 using a mass flow meter. In



Figure 3. Catalytic activity profiles of the 10FexMn/AC catalysts with different Fe:Mn molar ratios.

this process, the O_2 and NO concentrations are constant, and the total flow remains unchanged by adjusting the flow of N_2 . The effect of the NH₃/NO molar ratio on the activity of the 10Fe1Mn/AC catalyst at 240 °C is shown in Figure 4. The NO conversion increased slowly and the conversion is less than 70% when the NH₃/NO molar ratio increases from 0.2 to 0.6. However, it increases rapidly when the ratio is in the range of 0.6 to 1.0, and the highest conversion of NO (close to 100%) is achieved at a molar ratio of NH₃/NO = 1.0. Lower contents of NH₃ cause an obvious decrease of the NO_x conversion from the lack of a reductant. In contrast, higher contents of NH₃ may allow NH₃ to escape and lead to pollution in the environment. Therefore, the best NH₃/NO ratio is 1.0 in this reaction.



Figure 4. NO conversion of the 10Fe1Mn/AC catalyst at different NH₃/NO ratios.

Oxygen concentration effect

The O_2 concentration is an important parameter that significantly affects NO_x conversion, so the effect of the O_2 concentration on the conversion of NO_x was studied.

The change in O_2 concentration is achieved by adjusting the flow of O_2 using a mass flowmeter. The concentrations of NH₃ and NO are constant and the total flow keeps constant by changing the flow of N₂. The result is shown in Figure 5. It is evident that the NO conversion increases as the concentration of O_2 increases from 0 to 1.0% and then remains constant beyond 1.0%. It can be concluded that a concentration of 1.0% oxygen is suitable for the NH₃-SCR reaction for the 10Fe1Mn/AC catalyst.



Figure 5. NO conversion of the 10Fe1Mn catalyst at different oxygen concentrations at 240 $^{\circ}$ C.

Catalyst characterization

XRF and XRD

To identify the compositions of activated carbon, the AC which was calcined at 400 °C was analyzed by XRF, and the result is shown in Table 1. It can be seen from Table 1 that the main oxides detected in the calcined activated carbon are silica, alumina, calcium and potassium oxides. In addition, there are relatively small amounts of iron, manganese, copper oxides, as well as titanium oxides.

Figure 6 shows the X-ray diffraction patterns of different catalysts. The AC catalyst contains a broad band in the range of 20-30°, corresponding to the amorphous carbon structure. In addition, the XRD diffraction peaks corresponding to SiO₂ (at $2\theta = 26.6^{\circ}$, 36.26° , 50.16° , 60.01° , respectively, ICDD PDF-#46-1045) are also present.²⁹ The diffraction peaks of Fe and Mn species are not observed. These results are in agreement with the results of XRF (as shown in Table 1), which indicate that SiO₂ is the major component. The characteristic diffraction peaks of α -Fe₂O₃ (ICDD PDF-#33-0664) and γ -Fe₂O₃ (ICDD PDF-#39-1346) are present for the 10Fe/AC catalyst and the 10Fe/AC catalysts with various Mn contents. It can be seen from Figure 6 that the diffraction peaks of α -Fe₂O₃ and γ -Fe₂O₃ of 10Fe/AC catalyst have

Table 1. Ash composition analysis of activated carbon (AC) by X-ray fluorescence (XRF)

Sample	Composition analysis / %								
	SiO_2	Al_2O_3	CaO	ZnO	K_2O	MgO	Fe_2O_3	MnO	Others ^a
AC	59.93	15.94	5.35	2.92	3.40	2.05	0.659	0.065	9.751

^aOthers: Na₂O, La₂O₃, SrO and so on.

high intensity and narrow band width, suggesting a poor dispersion of iron oxide particles on the Fe-AC sample. After the addition of Mn, the diffraction peaks of α -Fe₂O₃ weaken. The intensity of the diffraction peaks of γ -Fe₂O₃ also decrease, and the shapes of the peaks become wider, which indicates that they have better dispersion. Upon an increase of the Fe:Mn molar ratio, the diffraction peaks of α -Fe₂O₃ and γ -Fe₂O₃ further weaken, and only a very weak peak corresponding to Mn₃O₄ appears ($2\theta = 18.3^{\circ}$) when the molar ratio of Fe to Mn increases from 10:3 to 10:7. No diffraction peaks of MnO_x appear for 10Fe1Mn/AC sample. These results may result from the monolayer dispersion capacity on the surface of carrier. When its loading is lower than the capacity, the oxide will be in a monolayer state, but its loading exceeds the capacity, the surplus oxide will remain as crystalline phase in the system together with its monolayer phase.³⁰ Increasing the molar ratio of Fe and Mn, only a very weak peak corresponding to Mn_3O_4 appears, indicating a better dispersion of manganese oxide particles on the Fe-Mn/AC samples. Fe and Mn species may exist on the catalysts, which would enhance the dispersion and inhibit the crystallization of both.31,32



Figure 6. X-ray diffraction patterns of 10Fe/AC and 10Fe/AC with different Mn loadings: (a) AC; (b) 10Fe/AC; (c) 10Fe1Mn/AC; (d) 10Fe3Mn/AC; (e) 10Fe5Mn/AC; (f) 10Fe7Mn/AC.

N₂ physisorption

The BET surface areas and pore volumes of the 10Fe/AC catalyst and 10Fe/AC catalysts with different Mn contents are summarized in Table 2. It can be observed from

Table 2 that the surface area (349.9 m² g⁻¹) and pore volume (0.40 mL g⁻¹) of 10Fe1Mn/AC are larger than those of the 10Fe/AC catalyst (306.1 m² g⁻¹ and 0.39 mL g⁻¹) owing to the high dispersion of manganese and iron oxides over the AC support. However, with an increase of the Fe:Mn molar ratio (Fe:Mn = 10:3, 10:5, 10:7), the BET surface area and pore volume of the catalysts decrease, which may result from the blocking effect of the support pores by the loading of manganese oxide. As shown in Figures 2 and 3, all the Mn-modified Fe₂O₃/AC catalysts show higher activity than 10Fe/AC catalyst, in spite of their lower surface area and pore volumes (apart from 10Fe1Mn/AC catalyst). The results indicate that the surface area and the pore volume are not the determining factor to improve catalytic activity. This is consistent with previous findings that the catalytic activity of catalyst is more dependent on surface chemistry than surface area.33

Table 2. Physical property of the catalysts

Catalyst	BET surface area / (m ² g ⁻¹)	Pore volume / (cm ³ g ⁻¹)		
10Fe/AC	306.1	0.39		
10Fe1Mn/AC	349.9	0.40		
10Fe3Mn/AC	300.3	0.38		
10Fe5Mn/AC	229.9	0.39		
10Fe7Mn/AC	176.5	0.36		

BET: Brunauer-Emmett-Teller.

XPS analysis

To study the oxidation state of the iron and manganese oxides of the different catalysts and to understand the atomic concentrations and chemical compositions of the surface layer, selected catalysts, such as 10Fe/AC, 10Fe1Mn/AC and 10Fe7Mn/AC, are investigated by XPS. The results are given in Table 3 and Figures 7 and 8.

Figure 7 shows the XPS spectra of Fe2p assigned to different catalysts. It can be seen from Figure 7a that two main peaks, Fe2p_{3/2} (peak at 711 ± 0.5 eV) and Fe2p_{1/2} (peak at 725 ± 0.5 eV), are obtained. There are satellite peaks on the right side of the two peaks, which correspond to Fe2p_{3/2} and Fe2p_{1/2}. Compared with the standard XPS spectrum of Fe³⁺, the peak at 711 ± 0.5 eV corresponds to Fe³⁺; therefore, Fe³⁺ is the main valence state for the

Table 3. Surface atomic concentrations of the catalysts

Catalant	Surface atomic concentration / at.%					
Catalyst	Fe	Mn	0			
10Fe/AC	14.16	_	31.70			
10Fe1Mn/AC	21.53	4.83	47.93			
10Fe7Mn/AC	14.27	7.15	40.98			



Figure 7. XPS spectra of different catalysts: (a) Fe2p; (b) Fe2p of 10Fe7Mn/AC catalyst.

10Fe/AC and 10Fe1Mn/AC catalysts. However, the Fe2p_{3/2} peak of the 10Fe7Mn/AC catalyst is wide and asymmetric and does not have satellite peaks. To further analyze the composition of the Fe species and the valence states on the 10Fe7Mn/AC catalyst surface, the Fe2p_{3/2} spectra is separated into two peaks by performing a peak-fitting technique, and the results are shown in Figure 7b. It can be seen from Figure 7b that the aforementioned peaks can be attributed to Fe²⁺ (709.8 eV) and Fe³⁺ (711.9 eV). The ratio of Fe³⁺/(Fe²⁺ + Fe³⁺) is 65.5%. This result shows that the content of Fe³⁺ in the 10Fe7Mn/AC catalyst. Previous reports have indicated that Fe³⁺ is conducive to NO conversion at low temperature.^{34,35} In addition, it can be



Figure 8. XPS spectra of $Mn2p_{3/2}$ and $Mn2p_{1/2}$ for the 10Fe1Mn/AC and 10Fe7Mn/AC catalysts.

seen from Table 3 that the Fe concentration on the surface of 10Fe/AC is 14.16%. After the addition of Mn, the Fe concentration increases to 21.53 and 14.27% on the surface of 10Fe1Mn/AC and 10Fe7Mn/AC, respectively. This indicates that the addition of Mn causes Fe to accumulate on the surface of catalysts, which would help to promote the SCR reaction.³⁶ However, the surface concentration of Fe over the 10Fe7Mn/AC catalyst is lower than that of the 10Fe1Mn/AC catalyst, indicating the introduction of excessive manganese oxide may cause the iron oxide into the porous system.³⁷ The above analysis may correspond to the activity of the 10Fe1Mn/AC catalyst being higher than that of the 10Fe7Mn/AC and 10Fe/AC catalysts.

Figure 8 shows the Mn2p photoelectron peaks of the 10Fe1Mn/AC and 10Fe7Mn/AC catalysts through deconvolutions of the spectra. The two main peaks at 642 and 654 eV are assigned to Mn2p_{3/2} and Mn2p_{1/2}, respectively. Using peak fitting to fit Mn2p_{3/2}, the spectra of Mn⁴⁺, Mn³⁺ and Mn²⁺ are obtained, and the binding energy of various valence states are listed in Table 4. Figure 8 shows that three kinds of valence states of Mn are present on the surface of 10Fe1Mn/AC and 10Fe7Mn/AC. The Mn2p_{3/2} binding energy of Mn⁴⁺ (MnO₂) is 644.6 ± 0.5 eV, and the corresponding binding energy of Mn³⁺ (Mn₂O₃) and Mn²⁺ (MnO), found in the literature, are 642 ± 0.5 and 640.5 ± 0.5 eV, respectively. The obtained binding energy of Mn2p_{3/2} in 10Fe1Mn/AC and 10Fe7Mn/AC are similar to those of MnO₂, Mn₂O₃ and MnO.³⁸

The ratios of the different valence states of Mn on the surface of the 10Fe1Mn/AC and 10Fe7Mn/AC catalysts are listed in Table 4, which show that the Mn⁴⁺/Mn³⁺ atomic ratio reaches 1.04 for the 10Fe1Mn/AC catalyst and decreases to 0.93 for the 10Fe7Mn/AC catalyst. Kapteijn *et al.*³⁹ reported that the NO conversion on pure manganese oxide was ranked by MnO₂ > Mn₅O₈ > Mn₂O₃ > Mn₃O₄. The higher

H₂-TPR

Chen et al.

0 1	Mn ²⁺		Mn ³⁺		Mı	Mn ⁴⁺	
Sample	BE/eV	Area	BE / eV	Area	BE / eV	Area	$- A_{Mn} / A_{Mn}$
10Fe1Mn/AC	640.9	1163	642.1	2506	644.6	2605	1.04
10Fe7Mn/AC	640.7	4129	642.0	6415	645.1	5968	0.93

Table 4. Binding energy (BE) and valence-state ratios of Mn on the surfaces of the obtained samples

 Mn^{4+}/Mn^{3+} ratio indicates that there are more MnO_2 in the catalysts, so 10Fe1Mn/AC catalyst exhibits a better lowtemperature catalytic activity.

The H₂-TPR curves of the Fe₂O₃, 10Fe/AC, 10Fe1Mn/AC

and 10Fe7Mn/AC catalysts are shown in Figure 9. As

observed from Figure 9, the reduction peak of Fe_2O_3 is

mainly divided into two parts. The reduction peaks of the

10Fe/AC and 10Fe1Mn/AC catalysts show little difference.

In addition, for the 10Fe7Mn/AC catalyst, the reduction

peak at low temperature disappears and changes at high

temperature. To analyze this precisely, fitting processes for

the 10Fe/AC, 10Fe1Mn/AC and 10Fe7Mn/AC catalysts are carried out, and the results are shown in Figure 10.

Figure 10 shows that Fe₂O₃ displays two reduction peaks,

to Fe²⁺ at 364 °C and Fe₃O₄ to FeO at 589 °C.⁴⁰ For the 10Fe/AC catalyst, the three peaks at 370, 592 and 721 °C are assigned to the reduction of Fe³⁺, Fe₃O₄ and oxygen-

450°C 604°C 10Fe7Mn/AC 563°C 393°C TCD Signal / (a.u.) 10Fe1Mn/AC 594°C 373°C 10Fe/AC 589°C 364°C Fe.O 300 400 500 Temperature / °C 100 200 600 700 800

Figure 9. H₂-TPR profiles of the catalysts.



Figure 10. H₂-TPR profiles of the catalysts: (a) Fe₂O₃; (b) 10Fe/AC; (c) 10Fe1Mn/AC; (d) 10Fe7Mn/AC.

containing functional groups in AC, respectively. The 10Fe1Mn/AC catalyst exhibits three reduction peaks at 391, 544 and 684 °C. Compared with the 10Fe/AC catalyst, the reduction temperature of Fe³⁺ increases, while the other two reduction peaks shift to lower temperatures because of the synergetic effect between MnO_x and FeO_x.⁴¹ For the 10Fe7Mn/AC catalyst, the reduction peaks occur at 442 and 631 °C. Compared with all of the catalysts tested, it is worth noting that the 10Fe1Mn/AC catalyst possesses the largest reduction area, which suggests that it has the strongest reduction ability.

Conclusions

The catalytic activity of the 10Fe/AC catalyst was increased by the addition of manganese oxide. The 10Fe1Mn/AC, 10Fe3Mn/AC, and 10Fe5Mn/AC catalysts showed the best performance at 180-240 °C among the various manganese-modified Fe₂O₃/AC catalysts. The modification with a certain amount of Mn species resulted in a better dispersion of the amorphous state, a larger surface area, higher existing amounts of Fe³⁺/(Fe³⁺ + Fe²⁺) and Mn⁴⁺/Mn³⁺, and a stronger reduction ability as detected by XRD, BET, XPS and H₂-TPR, which were all beneficial for improving the SCR performance of the catalyst. There are high NO conversion of 10Fe1Mn/AC catalyst at a molar ratio of NH₃/NO = 1.0 and a concentration of 1.0% oxygen.

In this study, the effect of SO_2 and H_2O on the catalytic activity was not conducted. As we know, the catalyst resistances to SO_2 and H_2O are key problems in industrial practice, which will be studied in the future.

Acknowledgments

We greatly appreciate the financial support provided by the National Natural Science Foundation of China (No. 51676001, 51376007, U1660206), the Anhui Provincial Natural Science Foundation (No. 1608085ME104), key projects of Anhui Province University Outstanding Youth Talent (No. gxyqZD2016074, gxyqZD2017038) and the National Key R&D Program (2016YFB0601402-02).

References

- Sahu, S. K.; Beig, G.; Parkhi, N.; Aerosol Air Qual. Res. 2015, 15, 1137.
- He, F. Q.; Deng, X. H.; Chen, M.; *Chemosphere* 2017, 168, 623.
- Li, W. M.; Liu, H. D.; Chen, Y. F.; Front. Environ. Sci. Eng. 2017, 11, 6.

- Yang, Y.; Xu, W. Q.; Wu, Y. H.; Wang, J.; Zhu, T. Y.; Catal. Commun. 2017, 94, 82.
- 5. He, F. Q.; Deng, X. H.; Chen, M.; Fuel 2017, 199, 523.
- Patila, B. S.; Cherkasovb, N.; Langc, J.; Ibhadonb, A. O.; Hessela, V.; Wang, Q.; *Appl. Catal.*, *B* 2016, *194*, 123.
- 7. Hung, C. M.; Aerosol Air Qual. Res. 2015, 8, 447.
- Shen, B. X.; Ma, H. Q.; He, C.; Zhang, X. P.; *Fuel Process. Technol.* 2014, 119, 121.
- Yang, W. W.; Liu, F. D.; Xie, L. J.; Lian, Z. H.; Hong, H.; *Ind. Eng. Chem. Res.* 2016, 55, 2677.
- Huo, Y. T.; Chang, Z. D.; Li, W. J.; Liu, S. X.; Dong, B.; Waste Biomass Valorization 2015, 6, 159.
- Shen, M. Q.; Li, C. X.; Wang, J. Q.; Xu, L. L.; Wang, W. L.; Wang, J.; *RSC Adv.* 2015, *5*, 35155.
- Peng, Y. J.; Li, H.; Si, W. Z.; Li, X.; Shi, W. B.; Luo, J.; Fu, M. J.; John, C.; Hao, J. M.; *Chem. Eng. J.* **2015**, *269*, 44.
- Grzybek, T.; Klinik, J.; Motak, M.; Papp, H.; *Catal. Today* 2008, 137, 235.
- Lei, Z. G.; Long, A. B.; Wen, C. P.; Zhang, J.; Chen, B. H.; *Ind. Eng. Chem. Res.* 2007, *50*, 5360.
- Qiu, L.; Pang, D. D.; Zhang, C. L.; Meng, J. J.; Zhu, R. S.; Ouyang, F.; *Appl. Surf. Sci.* 2015, *357*, 189.
- Kang, M.; Park, E. D.; Kim, J. M.; Yie, J. E.; *Appl. Catal.*, B 2007, 68, 21.
- Zhu, Y. W.; Zhang, Y. P.; Xiao, R.; Huang, T. J.; Shen, K.; *Catal. Commun.* **2017**, 88, 64.
- Xiong, Z. B.; Wu, C.; Hu, Q.; Wang, Y. Z.; Jin, J.; Lu, C. M.; Guo, D. X.; *Chem. Eng. J.* **2016**, 286, 459.
- 19. Thirupathi, B.; Smirniotis, P.; J. Catal. 2012, 288, 74.
- 20. Peña, D. A.; Uphade, B. S.; Smirniotis, P. G.; *J. Catal.* **2004**, *221*, 421.
- Luo, S. P.; Zhou, W. T.; Xie, A. J.; Wu, F. Q.; Yao, C.; Li, X. Z.; Zuo, S. X.; Liu, T. H.; *Chem. Eng. J.* 2015, 286, 291.
- 22. Zahaf, R.; Jung, J. W.; Coker, Z.; Kim, S.; Choi, T. Y.; Lee, D.; Aerosol Air Qual. Res. 2015, 15, 2409.
- 23. Chansai, S.; Burch, R.; Hardacre, C.; Norton, D.; Bao, X.; Lewis, L.; *Appl. Catal.*, *B* **2014**, *160*, 356.
- 24. Sun, D. K.; Liu, Q. Y.; Liu, Z. Y.; Gui, G. Q.; Huang, Z. G.; Appl. Catal., B 2009, 92, 462.
- Li, S. J.; Wang, X. X.; Tan, S.; Shi, Y.; Li, W.; *Fuel* 2017, 191, 511.
- Liu, Y.; Ning, P.; Li, K.; Tang, L. H.; Hao, J. M.; Song, X.;
 Zhang, G. J.; Wang, C.; *Russ. J. Phys. Chem. A* 2017, *91*, 490.
- Samojeden, B.; Motak, M.; Grzybek, T.; C. R. Chim. 2015, 18, 1049.
- Liu, L.; Gao, X.; Song, H.; Zheng, C. H.; Zhu, X. B.; Luo, Z. Y.; Ni, M. J.; Cen, K. F.; *Aerosol Air Qual. Res.* 2014, *14*, 1038.
- Gao, X.; Liu, S. J.; Zhang, Y.; Luo, Z. Y.; Cen, K. F.; J. Hazard. Mater. 2011, 188, 58.
- 30. Xie, Y. C.; Tang, Y. Q.; Adv. Catal. 1990, 37, 1.
- 31. Qi, G. S.; Yang, R. T.; Appl. Catal., B 2003, 44, 217.

- 32. Huang, J. H.; Tong, Z. Q.; Huang, Y.; Zhang, J. F.; *Appl. Catal., B* **2008**, *78*, 309.
- 33. Li, W.; Tan, S.; Shi, Y.; Li, S. J.; Fuel 2015, 160, 35.
- Devadas, M.; Krocher, O.; Elsener, M.; Wokaun, A.; Mitrikas, G.; Soger, N.; Pfeifer, M.; Memel, Y.; Mussmann, L.; *Catal. Today* 2007, *119*, 137.
- Delahay, G.; Valade, D.; Guzmanvargas, A.; Coq, B.; *Appl. Catal.*, B 2005, 55, 149.
- Wang, X. B.; Wu, S. G.; Zou, W. X.; Yu, S. H.; Gui, K. T.; Dong, L.; Chin. J. Catal. 2016, 37, 1314.
- 37. Samojeden, B.; Grzybek, T.; Energy 2016, 116, 1484.

- Chen, Z. H.; Yang, Q.; Li, H.; Li, X. H.; Wang, L. F.; Tsang, S. C.; J. Catal. 2010, 276, 56.
- Kapteijn, F.; Singoredjo, L.; Andreini, A.; Moulijn, J. A.; *Appl. Catal.*, B 1994, 3, 173.
- 40. Mauvezin, M.; Delahay, G.; Coq, B.; Kieger, S.; Jumas, J. C.; Olivier-Fourcade J.; *J. Phys. Chem. B* **2001**, *105*, 928.
- 41. Qi, G. S.; Yang, R. T.; J. Catal. 2003, 217, 434.

Submitted: March 21, 2017 Published online: June 26, 2017